IMPACTS OF SILVER NANOPARTICLES ON BACTERIAL SPECIES *B*. SUBTILIS AND *E*. COLI AND THE MAJOR CROP PLANT *Z*. MAYS

by

Michael A. Doody

A thesis submitted to the Faculty of the University of Delaware in partial fulfillment of the requirements for the degree of Master of Science in Plant and Soil Sciences

Summer 2014

© 2014 Michael A. Doody All Rights Reserved UMI Number: 1567800

All rights reserved

INFORMATION TO ALL USERS The quality of this reproduction is dependent upon the quality of the copy submitted.

In the unlikely event that the author did not send a complete manuscript and there are missing pages, these will be noted. Also, if material had to be removed, a note will indicate the deletion.



UMI 1567800

Published by ProQuest LLC (2014). Copyright in the Dissertation held by the Author.

Microform Edition © ProQuest LLC. All rights reserved. This work is protected against unauthorized copying under Title 17, United States Code



ProQuest LLC. 789 East Eisenhower Parkway P.O. Box 1346 Ann Arbor, MI 48106 - 1346

IMPACTS OF SILVER NANOPARTICLES ON BACTERIAL SPECIES *B*. SUBTILIS AND *E*. COLI AND THE MAJOR CROP PLANT *Z*. MAYS

by

Michael A. Doody

Approved:	
	Yan Jin, Ph.D.
	Professor in charge of thesis on behalf of the Advisory Committee
Approved:	
	Blake Meyers, Ph.D.
	Chair of the Department of Plant and Soil Sciences
Approved:	
	Mark Rieger, Ph.D.
	Dean of the College of Agriculture and Natural Resources
Approved:	
	James G. Richards, Ph.D.
	Vice Provost for Graduate and Professional Education

ACKNOWLEDGMENTS

I would like to express my deepest gratitude to my advisor, Dr. Yan Jin for her unlimited encouragement and mentoring. Without her guidance this thesis would not have been possible. I would also like to thank Dr. Harsh Bais, who provided me with an excellent foundation in rhizosphere ecology and plant-microbe interactions. Additionally, Dr. Chin-Pao Huang and Dr. Angelia Seyfferth have been wonderful committee members, offering their continued guidance and support. Special thanks go to William Bartz, Caroline Golt, Shannon Modla, Deborah Powell, Jean Ross, Jerry Pourier, and Teclemariam Weldekidan for their assistance.

I am also thankful to the many post-doctoral researchers and fellow graduate students with whom I worked, including Venkatachalam Lakshmanan, Olga Lazouskaya, Gang Li, Amanda Roberson, Rafael Saldana, Gopinath Selvaraj, Carla Spence, Taozhu Sun, Gang Wang, Jing Yan, and Wenjuan Zheng. I am extremely grateful to have shared the laboratory with such bright and talented scientists.

I would also like to thank my parents and brother for their unconditional support and encouragement.

Finally, I would like to thank my fiancée, Jacqueline O'Neill. She has been a constant source of compassion, encouragement, and reassurance throughout this process, and I am forever grateful for her love.

TABLE OF CONTENTS

LIST LIST ABST	OF TA OF FI 'RAC'	ABLES GURES T	5	vi vii ix
Chapt	er			
1	INT	RODU	CTION	1
	1.1 1.2	Introd Literat	uction to Silver Nanoparticles ture Review	1 2
		1.2.1 1.2.2 1.2.3 1.2.4	Silver Nanoparticles Antibacterial Activity Phytotoxicity Disruption of Mutually Beneficial Relationships	2 3 9 13
	1.3	Resear	rch Goals and Objectives	15
2	SILV SOI MAX	VER NA L BACT YS	ANOPARTICLES LIMIT INTERACTIONS BETWEEN THE FERIUM <i>B. SUBTILIS</i> AND THE MAJOR CROP PLANT <i>Z.</i>	16
	2.1 2.2	Introd Mater	uction ials and Methods	16 17
		2.2.1 2.2.2 2.2.3	Silver and Silver Nanoparticles Biological Materials Determination of c-AgNP Morphology, Stability, and Ion	17 17
		2.2.4 2.2.5 2.2.6	Release Bacterial Susceptibility to c-AgNPs <i>Z. mays</i> Susceptibility to c-AgNPs Susceptibility of Bacteria-Inoculated <i>Z. mays</i> to c-AgNPs	18 20 22 24
	2.3	2.2.7 Result	statistical Analysis	25
		2.3.1	Particle Characterization	25

 2.3.3 c-AgNPs Sorb onto Bacterial Cell Surfaces 2.3.4 Comparing Effects of Exposure between Bacterial Species 	34 35 37
2.3.4 Comparing Effects of Exposure between Bacterial Species	35 37
	37
2.3.5 Phytotoxicity of AgNPs on Z. mays roots	
2.3.6 Stunting of Tertiary and Fine Roots	42
2.3.7 Decreased Beneficial Interactions in Bacteria-Inoculated Z.	
mays Exposed to c-AgNPs	42
2.4 Conclusion	47
2.1 Contractor	• /
3 CONCLUSION AND SYNTHESIS	48
3.1 Introduction	48
3.2 Summary of Findings and Connection to Current Scientific	-0
Understanding	48
	10
3.2.1 c-AgNP Stability and Dissolution	49
3.2.2 Bacterial Toxicity	49
3.2.3 Sub-lethal Toxicity to Z. mays	51
3.2.4 Impacts on Plant-Microbe Interactions	51
3.3 Limitations and Recommendations for Future Research	52
3.4 Conclusion	53
KEFERENCES	54
Appendix	

А	PLANT NUTRIENT SOLUTION COMPOSITION	63
В	PARTICLE SIZE ANALYSIS	. 64
С	STATISTICAL TABLES	.73

LIST OF TABLES

HDD and Zeta Potential of c-AgNPs in Various Media	. 27
Kinetics Parameters	. 33
Plant Nutrient Solution	. 63
Particle Size Analysis	. 64
ANOVA Results for Inhibition of <i>B. subtilis</i> and <i>E. coli</i> by c-AgNPs	. 73
Tukey's Post-Hoc Results for Inhibition of <i>B. subtilis</i> and <i>E. coli</i> by c-AgNPs	. 74
ANOVA Results for Growth Kinetics of <i>B. subtilis</i> and <i>E. coli</i> exposed to c-AgNPs	. 79
Tukey's Post-Hoc Results for Growth Kinetics of <i>B. subtilis</i> and <i>E. coli</i> Exposed to c-AgNPs	. 80
ANOVA Results for Length, Biomass, and Root Ag Content of Z. mays Seedlings	. 89
Tukey's Post-Hoc Results for Length, Biomass and Root Ag Content of <i>Z. mays</i> Seedlings	. 90
	 HDD and Zeta Potential of c-AgNPs in Various Media Kinetics Parameters Plant Nutrient Solution Particle Size Analysis ANOVA Results for Inhibition of <i>B. subtilis</i> and <i>E. coli</i> by c-AgNPs Tukey's Post-Hoc Results for Inhibition of <i>B. subtilis</i> and <i>E. coli</i> by c-AgNPs ANOVA Results for Growth Kinetics of <i>B. subtilis</i> and <i>E. coli</i> exposed to c-AgNPs Tukey's Post-Hoc Results for Growth Kinetics of <i>B. subtilis</i> and <i>E. coli</i> exposed to c-AgNPs ANOVA Results for Length, Biomass, and Root Ag Content of <i>Z. mays</i> Seedlings Tukey's Post-Hoc Results for Length, Biomass and Root Ag Content of <i>Z. mays</i> Seedlings

LIST OF FIGURES

Figure 1:	Current understandings and questions of environmental transformations of AgNPs (From Levard et al., 2012)
Figure 2:	TEM micrographs showing impacts of AgNPs on cellular membranes of <i>B. subtilis</i> . A) Control cells. B-D) Cells exposed to AgNPs. White arrows refer to AgNPs and black arrows refer to the cellular membrane disruptions. (From El Badawy et al., 2011)
Figure 3:	Proposed mechanisms of Ag ⁺ toxicity to bacterial cells, including blocking of respiratory chain, collapsing membrane potential and stopping ATP production, promoting the runaway production of ROS and damaging membrane lipids and DNA, and influencing metabolic activity by binding to intracellular proteins and chromosomes. (From Rizzello and Pompa, 2014)
Figure 4:	Current understanding of the potential pathways for particle-plant associations. (From Dietz and Herth, 2011)
Figure 5:	Phenotypic difference in <i>A. thaliana</i> seedlings exposed to AgNPs, including reductions in shoot length and browning of root tips. (From Geisler-Lee et al., 2013)
Figure 6:	Dose-response effect on wheat seedlings due to exposure to AgNPs (From Dimpka et al., 2013)
Figure 7:	Impacts of Cd exposure to rhizobia include reduction in thalli numbers and alterations to thalli morphology (From Chen et al., 2003)
Figure 8:	A) TEM micrograph of c-AgNPs in stock solution. Scale bar is 100 nm. B) Size Distribution of c-AgNPs in stock solution. Mean TEM diameter = 44.9 ± 7.2 nm, n = 607
Figure 9:	A) Dynamic growth curve for <i>B. subtilis</i> exposed to c-AgNPs; B) Dynamic growth curve for <i>E. coli</i> exposed to c-AgNPs
Figure 10:	A) Dynamic growth curve for <i>B. subtilis</i> exposed to AgNO ₃ ; B) Dynamic growth curve for <i>E. coli</i> exposed to AgNO ₃ 31

Figure 11: FE-SEM micrographs of <i>B. subtilis</i> (A) and (B) and <i>E. coli</i> (C) and (D). Arrows denote c-AgNPs or aggregates	35
Figure 12: A) Digital photograph of <i>Z. mays</i> roots after 7-d exposure; B) Digital photograph of <i>B. subtilis</i> -inoculated <i>Z. mays</i> seedlings after 7-d exposure.	.38
Figure 13: Root length (A) and Wet Root Biomass (B) of <i>Z. mays</i> seedlings after 7- d exposures. Error bars are ± 1 SD	. 39
Figure 14: Ag content of Z. mays roots after 7-d exposure to c-AgNPs. Error bars are ± 1 SD.	41
Figure 15: FE-SEM micrographs of Z. mays roots. A-C) Control; D-F) 5.0 mg/L c- AgNP.	.44
Figure 16: Concentrations of bacteria in PNS+RE after 7-d exposure. Error bars are ± 1 SD	.46

ABSTRACT

This thesis examines the impacts of citrate-coated silver nanoparticles (c-AgNPs) on two species of bacteria (Bacillus subtilis and Escherichia coli), the major crop plant Zea mays, and the beneficial plant-microbe relationship between Z. mays and B. subtilis. AgNPs are an increasing component of antimicrobial consumer, industrial, and military products. This has led to widespread scientific concern for the ecological safety outside their intended use. An overview of their history, use, and toxicity was used to inform the design of experiments and resulting data. Growth inhibition and sub-lethal toxic effects were used to assess the effects of c-AgNP exposure to bacteria. Similar analytical methods were used to quantify the response of Z. mays to c-AgNP exposure. Results showed that exposure to c-AgNP significantly reduced the growth of bacterial populations and alters their growth kinetics. Z. mays experienced significant sub-lethal effects due to exposure, including reduced root length and biomass, and hyper-accumulated Ag in root tissues. Beneficial interactions between B. subtilis and Z. mays were reduced as both species suffered sub-lethal effects of exposure to c-AgNPs.

Chapter 1

INTRODUCTION

1.1 Introduction to Silver Nanoparticles

Silver nanoparticles (AgNPs), defined as engineered individual silver particles or small aggregates of silver particles measuring no more than 100 nm in any dimension (USEPA, 2007), are used in a variety of applications, including commercial antimicrobial products such as bandages and socks and other textiles, as well as various military and industrial products (Morones et al., 2005). In fact, AgNPs are one of the most commonly used nanomaterials in consumer products (Fabrega et al., 2011; El-Temsah and Joner, 2010; Klaine et al., 2008). The increasing use of AgNPs in such applications has greatly increased environmental risk of exposure (Gottschalk et al., 2009; Mueller and Nowack, 2008; Nowack and Bucheli, 2007; USEPA, 2012) and associated scientific concern.

Numerous studies on the toxicity of AgNPs to aquatic organisms (Wijnhoven et al., 2009; Lee et al., 2012b; Morones et al., 2005) and soil-dwelling nematodes (Kim et al., 2012; Yang et al., 2012; Meyer et al., 2010; Roh et al., 2009;) have been published, and the influence of AgNPs on bacteria (Suresh et al., 2010; El Badawy et al., 2011) and a select number of plant species is becoming increasingly known (Geisler-Lee et al., 2013; Qian et al., 2013; Yin et al., 2011; El-Temsah and Joner, 2010). However, studies on the influence of AgNPs on specific plant-microbe interactions are lacking. Because plant-microbe interactions are ubiquitous in both natural and agricultural soils (Berg, 2009; Berg et al., 2005), it is critical to not only

understand how each organism is impacted by exposure to AgNPs, but if and how those critical interactions between the organisms are altered.

1.2 Literature Review

In the following research, I integrate traditional toxicology with microbiology, plant developmental biology, and colloid science to investigate the influence of AgNPs on the growth and survival of two species of bacteria and the major crop plant *Zea mays*, while also determining if exposure to AgNPs limits plant-microbe interactions. I present an overview of the significant literature pertaining to AgNPs in the review below. Studies of specific relation to this research and the larger picture of AgNP use and safety are further highlighted.

1.2.1 Silver Nanoparticles

Silver has been used for nearly a century for its antimicrobial and biocidal properties (Nowack et al., 2011). Over the past 20 years, the military and industry have begun to make use of silver's antimicrobial properties by applying silver in nanoform to their products, such as field dressings, socks, and even washing machines and dishwashers (Ma et al., 2010; Klaine et al., 2008). Recent studies have shown that silver is the most widely used metallic nanoparticle in consumer products (Maynard and Michelson, 2014; Benn and Westerhoff, 2008; Klaine et al., 2008). Some debate over the safety and regulation of AgNPs has arisen recently, resulting in increased research and policy inquiries by the government and interest groups (Nowack et al., 2012; Stone et al., 2010; Gottschalk et al., 2009; Blaser et al., 2008). This research focuses on the scientific concern for AgNPs' ecological safety outside their intended use.

2

1.2.2 Antibacterial Activity

Presently, there are two major explanations for how AgNPs are toxic to bacteria – as well as other species such as algae, nematodes, and fish (Kittler et al., 2010; Meyer et al., 2010; Choi et al., 2008; Navarro et al., 2008). Several authors attribute toxicity to some nano-scale property or combination of such properties differing from bulk ionic silver (Sharma et al., 2014; Levard et al., 2012; Navarro et al., 2008) while others identify the release of ionic silver from AgNPs as the primary mechanism of toxicity (Kittler et al., 2010; Choi and Hu, 2009; Choi et al., 2008). Nano-scale properties that may differentiate toxicity from bulk species include increased specific surface area and reactivity or photocatalytics that alter how organisms, especially microorganisms such as bacteria, interact with the particles (Meyer et al., 2010). These properties are ultimately altered through environmental interactions based on particle behavior and environmental conditions (Fig. 1).



Figure 1: Current understandings and questions of environmental transformations of AgNPs (From Levard et al., 2012).

Particles in natural environments will quickly undergo transformation, including dissolution, transformation through redox reactions, and complexation with thiols and natural organic matter, and the rate at which transformations occur is strongly linked to the stability of the particle coating (Levard et al., 2012).

1.2.2.1 AgNP-Specific Toxicity

AgNP-related bacterial toxicity mechanisms are becoming increasingly understood, and possible mechanisms include attachment of particles or particle aggregates to cellular membranes. Such interaction may result in changes to membrane permeability and the cytosol redox cycle, accumulation of intracellular radicals, or disruption of ATP synthesis (Nel et al., 2009; Lok et al., 2006; Morones et al., 2005; Sondi and Salopek-Sondi, 2004). Several authors have demonstrated robust evidence for toxicity mechanisms directly related to AgNPs (El Badawy et al., 2011; Suresh et al., 2010; Choi and Hu, 2008). For example, Choi and Hu showed greater levels of inhibition by AgNPs than equal concentrations of ionic silver (2008). El Badawy et al. (2011) showed greater levels of toxicity due to treatment with surfacecharged AgNPs, whereby cell-NP interactions resulted in the disruption of the organism's cellular membrane (Fig. 2) and ultimately cell death. Furthermore, Suresh et al. (2010) showed significant growth inhibition of B. subtilis and E. coli in AgNP solutions with less than 5% Ag⁺ by mass. Such Ag⁺ concentrations are more than 10x less than previously identified threshold concentrations for Ag⁺ toxicity (Suresh et al., 2010; Li et al., 1997).



Figure 2: TEM micrographs showing impacts of AgNPs on cellular membranes of *B. subtilis*. A) Control cells. B-D) Cells exposed to AgNPs. White arrows refer to AgNPs and black arrows refer to the cellular membrane disruptions. (From El Badawy et al., 2011).

1.2.2.2 Ionic Ag-Related Toxicity

However, several studies have also shown toxicity related to the extended release of ionic silver. This popular viewpoint is partially justified by the historical use of Ag⁺ antimicrobial agent (Maynard and Michelson, 2014) and even the recent incorporation of AgNPs into consumer products marketed as antimicrobial. The mechanism behind Ag⁺ toxicity is well-studied; for bacteria, the positive charge of the ion promotes sorption onto the negatively charged cell wall, resulting in the deactivation of cellular enzymes, and the disruption of membrane permeability (Sambhy et al., 2006; Ratte, 1999). Additionally, uptake of Ag⁺ has been shown to generate intracellular reaction oxygen species (ROS) leading to cell lysis and death (Ratte, 1999). Other mechanisms include deleterious interactions with nucleic acids and sulfur-containing metabolic enzymes (Ahamed et al., 2008; Morones et al., 2005). Rizzello and Pompa provide an excellent schematic of AgNP-related Ag+ toxicity to bacteria (Fig. 3) (2014). Choi et al. (2008) found that autotrophic and heterotrophic bacterial species were susceptible to exposure to ionic silver species, and that observed AgNP-toxicity could not be specifically linked to the particles themselves due to a lack of cell membrane disruption. These authors also observed a steady shift in the color of AgNP suspensions from yellow to dark brown during their experiments, indicating oxidative dissolution of AgNPs to Ag⁺ species. Evidence of diffusion of ionic silver across cell membranes by sorbed AgNPs has also been observed (Choi and Hu, 2009). Kittler et al. (2010) found that Ag⁺ present in aged suspensions of AgNPs resulted in greater toxicity to human mesenchymal stem cells than fresh AgNP suspensions with comparatively less Ag^+ .



Figure 3: Proposed mechanisms of Ag⁺ toxicity to bacterial cells, including blocking of respiratory chain, collapsing membrane potential and stopping ATP production, promoting the runaway production of ROS and damaging membrane lipids and DNA, and influencing metabolic activity by binding to intracellular proteins and chromosomes. (From Rizzello and Pompa, 2014).

1.2.2.3 Discrepancies in Toxicity Mechanisms

There currently exists a fundamental lack of concrete knowledge on the toxicity of AgNPs, though significant progress has been made for some classes of organisms. There is significant evidence for both AgNP and Ag⁺ toxicity, yet most studies are inconclusive. The literature is currently incomplete and sometimes contradictory, sometimes providing evidence for both mechanisms of toxicity. For instance, several studies have shown significant growth inhibition by 1.0 mg/L AgNP with no observed disruption of cell membranes using LIVE/DEAD analysis (Choi et

al., 2009; Choi and Hu, 2009). However, other studies showed almost complete cell membrane disruption by similar concentrations (El Badawy et al., 2011). Advanced imaging techniques such as transmission electron microscopy (TEM) and scanning electron microscopy (SEM) have been used extensively by most authors, and in each case significant sorption of AgNPs onto cell surfaces has been observed (El Badawy et al., 2011; Suresh et al., 2010; Choi et al., 2008, 2009). Several authors have also shown evidence of localized uptake of small AgNPs (Choi and Hu, 2009; Choi et al., 2008).

Such discrepancies have not gone unnoticed in the literature, as several review articles have recognized and commented on the variability of results. For example,

Levard et al. (2012) state:

Although the toxicity of Ag-NPs is partly explained by the release of Ag ions, it remains unclear if Ag-NPs are a direct cause of enhanced toxicity. For example, Navarro et al. (2008) presented evidence that toxicity is mainly the result of Ag ions and that Ag-NPs contribute to toxicity as a source of dissolved Ag ions. In contrast, Fabrega et al. (2011) showed a specific nanoparticle effect that could not be explained by dissolved Ag⁺. Similarly, Yin et al. (2011) demonstrated that gum arabic-stabilized Ag-NPs more strongly affected the growth of Lolium multiflorum, a common grass, more than the equivalent dose of Ag ions added as AgNO₃. They concluded that growth inhibition and cell damage can be directly attributed either to the nanoparticles themselves or to the ability of Ag-NPs to deliver dissolved Ag⁺ to critical biotic receptors. Recently Sotiriou et al. (2010) proposed that the antibacterial activity of Ag-NPs depends on their size. They provide some evidence that when Ag-NPs are small and release many Ag ions, the antibacterial activity is dominated by these ions. However, when relatively large (mean diameter >10 nm) Ag-NPs are employed, the concentration of released Ag⁺ is lower, and the particles themselves also influence Ag-NP antibacterial activity.

The conflicting nature of the literature may point toward some combination of the two different mechanisms. One such combination of mechanisms may be that the dissolution of Ag^+ from AgNPs is a nano-specific property related to the incredibly large surface area to volume ratios exhibited by most particles. However, because the

size and coating of AgNPs is varied by use and manufacturer, the amount of dissolution across studies is difficult to compare. Ma et al. have shown that these properties are strongly related to particle dissolution (2012), thus providing the first evidence and argument that any toxicity resulting from dissolved Ag is then fundamentally linked to the specific properties of the initial AgNP source. Other authors have consistently found that particle coatings play an integral role in behavior ranging from stability and dissolution to toxicity (Yang et al., 2012; El Badawy et al., 2011; Meyer et al., 2010)

1.2.3 Phytotoxicity

There is increasing evidence for sub-lethal toxic effects of exposure to metallic NPs for a variety of plant species, including the model system *Arabidopsis thaliana*, and more ecologically and agriculturally relevant species such as *Lolium multiflorum*, *Triticum aestivum*, and *Z. mays*. There are five main modes of biological interaction with NPs that could potentially lead to toxicity. These include: chemical effects, physical toxicity (association of NPs with cell structures or mechanical clogging), catalytic effects, surface effects, and changes to environment (Dietz and Herth, 2011). To date, most studies have focused on surface effects (especially phenotypic responses), while some have evaluated chemical effects, physical toxicity, and to some extent changes in local environments.

Particle-plant associations can occur via the root or the shoot (Dietz and Herth, 2011), and may lead to long distance translocation given the proper conditions and NP size (Fig. 4). Most studies of metal-NP toxicity to plant species have focused on the root system as the point of origin for particle-plant association. This is justified in

9

most cases through the application of biosolids and the major influence of root system health on plant development and overall health.



Figure 4: Current understanding of the potential pathways for particle-plant associations. (From Dietz and Herth, 2011).

Several studies on the impacts of AgNP exposure on the model plant system *A*. *thaliana* have shown significant sub-lethal toxicity. Localized uptake and storage of AgNPs in root caps and intracellular space resulted in browning of root tips and limited development of root hairs in developing *A*. *thaliana* seedlings (Fig. 5) (Geisler-Lee et al., 2013). *A. thaliana* was further shown to be susceptible to AgNP exposure by Qian et al. (2013). These authors demonstrated significant inhibition of root growth by AgNPs, as well as disruptions to the plants' thylakoid membrane structure and a decrease in chlorophyll content.



Figure 5: Phenotypic difference in *A. thaliana* seedlings exposed to AgNPs, including reductions in shoot length and browning of root tips. (From Geisler-Lee et al., 2013).

Studies on more ecologically and agriculturally significant plants have also been conducted. Yin et al. (2011) showed significant growth inhibition, cell damage and alterations to root morphology in *L. multiflorum* exposed to AgNPs. Roots grown in 5 mg/L AgNP nutrient solution accumulated ~ 100 mg/kg Ag with a bioconcentration factor close to 30. These results indicate that Ag can become hyperaccumulated in the root systems of plants grown in NP-spiked media. The authors showed similar uptake rates and bioconcentration factors for higher concentrations of Ag, indicated a dose-response effect. However, the high root [Ag] was not necessarily reflected in the shoots of exposed seedlings, as shoot [Ag] and bioconcentrarion factors were 2 orders of magnitude lower, a well understood phenomenon and typical finding (Yin et al., 2011). Pokhrel and Dubey (2013) showed substantial sub-lethal effects in *Z. mays* seedlings exposed to AgNPs, including alterations in root morphology due to structural changes in primary root cells. AgNPs have also been shown to disrupt the growth of *T. aestivum* (Dimpka et al., 2013). The authors showed decreased biomass due to induced branching of roots, as well as reductions in root and shoot length (Fig 6). Significant accumulation of Ag in the roots was also observed, indicating some degree of uptake and translocation.



Figure 6: Dose-response effect on wheat seedlings due to exposure to AgNPs (From Dimpka et al., 2013).

As with the mechanism of toxicity to bacteria and other microorganisms, there is some debate over the mechanism of toxicity in higher plants. However, there seems to be more robust evidence for some nano-specific mechanism. Yin et al. (2011) argue that growth inhibition and altered root morphology in *L. multiflorum* were the result of AgNP-specific interactions. They specifically cite an increased surface area to volume ratio that promotes increased surface interactions for smaller particles. They also showed that the addition of Ag^+ -binding ligands did not significantly reduce the toxic effects of AgNPs. Additionally, Geisler-Lee et al. (2013) showed that exposure to AgNPs resulted in increased Ag content in *A. thaliana* roots than exposure to equivalent concentrations of AgNO₃. Still, there is minimal evidence of translocation of particles larger than 5 nm, indicating that any plant-particle associations will be mainly limited to the roots (Yin et al., 2011; Shane et al., 2000). Additionally, Dimpka et al. (2013) showed that AgNO₃ applied in doses equivalent to the soluble fraction of their AgNPs did result in significant reduction in plant growth.

To date there is limited vigorous evidence linking toxicity to dissolved Ag species. Lee et al. (2012a) showed significant toxic effects of AgNO₃ released from AgNPs to *Phaseolus radiatus* and *Sorghum bicolor* when grown on agar media, but no discernable toxic effects when grown in soil media. Reductions in root and shoot length in *Hordeum vulgare* and *L. multiflorum* exposed to AgNPs was attributed to the presence of Ag⁺ in solution (El-Temsah and Joner, 2010). However, as with other organisms, sorption of AgNPs to root surfaces may result in the diffusion of Ag⁺ across the membrane where it can be accumulated and result in runaway production of ROS and eventually cell apoptosis (Kim et al., 2009; Choi et al., 2008). Additionally, Ag⁺ has been shown to restrict ethylene activation in plants (Stampoulis et al., 2009) and inhibit mitochondrial function (Knee, 1992).

1.2.4 Disruption of Mutually Beneficial Relationships

Few studies have addressed the impacts of AgNPs on symbiotic relationships, which can be altered in different ways. Most relevant work to date has focused on individual constituents known to participate in such mutually beneficial relationships. Extensive work has been done on the effects of AgNP exposure to nitrogen-fixing bacteria as well as nitrifying and denitrifying bacteria (Bharadway, 2012).

In separate studies, both Priester et al. (2013) and Chen et al. (2003) found that the symbiotic relationship between soybean plants and N_2 -fixing bacteria was disrupted by exposure to a variety of metal or metal NPs. Chen et al. (2003) showed significantly reduced numbers of thalli in rhizobium exposed to bulk Cd, along with drastic morphological changes (Fig. 7).



Figure 7: Impacts of Cd exposure to rhizobia include reduction in thalli numbers and alterations to thalli morphology (From Chen et al., 2003).

However, no studies of the impacts of AgNPs on such symbioses have been completed to date. Fundamental knowledge of the principles of these positive relationships can provide clues as to how participating organism will response to AgNP exposure (Priester et al., 2013). Specifically, direct disruption via toxicity to one or more participants should reduce efficiency of relationship, while indirect disruptions may occur through alteration of soil conditions such as pH, or through larger disruptions of the food web (Priester et al., 2013). However, the lack of understanding of how AgNPs exert toxicity and impact soil processes has led to the inability to describe how whole soil ecosystem processes such as plant-microbe interactions are affected.

1.3 Research Goals and Objectives

The interpretation of toxicity across organisms and classes of organisms should be based on particle characteristics, not just concentration (Stone et al., 2010). Additionally, mechanisms are likely to be different for different types of organisms: bacteria may be affected by cell membrane disruption or direct uptake, while plants may be affected by physical blockage of pores by adsorption of NPs (Dietz and Herth, 2011; Navarro et al., 2008). Until these knowledge gaps are addressed, studies of how biomass production, organic matter breakdown, nutrient cycling, are affected by AgNPs will remain superficial and incomplete. In this research, I addressed knowledge gaps pertaining to bacterial and plant toxicity and the disruption of beneficial relationships between beneficial bacteria and their plant hosts upon exposure to AgNPs. Specifically, this research aimed to quantify and characterize AgNP toxicity to bacterial species *Bacillus subtilis* and *Escherichia coli*; determine the effects of AgNP exposure on the major crop plant *Z. mays*, and to quantify reductions of the beneficial plant-microbe relationship between *Z. mays* and *B. subtilis*.

Chapter 2

SILVER NANOPARTICLES LIMIT INTERACTIONS BETWEEN THE SOIL BACTERIUM B. SUBTILIS AND THE MAJOR CROP PLANT Z. MAYS

2.1 Introduction

Silver nanoparticles (AgNPs), defined as engineered individual silver particles or small aggregates of silver particles measuring no more than 100 nm in any dimension (USEPA, 2007), are used in a variety of applications, including commercial antimicrobial products such as bandages and socks and other textiles, as well as various military and industrial products (Morones et al., 2005). In fact, AgNPs are one of the most commonly used nanomaterials in consumer products (Fabrega et al., 2011; El-Temsah and Joner, 2010; Klaine et al., 2008). The increasing use of AgNPs in such applications has greatly increased environmental risk of exposure (Gottschalk et al., 2009; Mueller and Nowack, 2008; Nowack and Bucheli, 2007; USEPA, 2012) and associated scientific concern.

Numerous studies on the toxicity of AgNPs to aquatic organisms (Wijnhoven et al., 2009; Lee et al., 2012; Morones et al., 2005) and soil-dwelling nematodes (Kim et al., 2012; Yang et al., 2012; Meyer et al., 2010; Roh et al., 2009;) have been published, and the influence of AgNPs on bacteria (Suresh et al., 2010; El Badawy et al., 2011) and a select number of plant species is becoming increasingly known (Geisler-Lee et al., 2013; Qian et al., 2013; Yin et al., 2011; El-Temsah and Joner, 2010). However, studies on the influence of AgNPs on specific plant-microbe interactions are lacking. Because plant-microbe interactions are ubiquitous in both natural and agricultural soils (Berg, 2009; Berg et al., 2005), it is critical to not only understand how each organism is impacted by exposure to AgNPs, but if and how those critical interactions between the organisms are altered.

Here, we investigate the impacts of AgNPs on the growth behavior and kinetics of two bacterial species, *B. subtilis* and *E. coli*; their impacts on the growth of *Z. mays* seedlings; and their effect on the beneficial plant-microbe interaction between *B. subtilis* and *Z. mays*.

2.2 Materials and Methods

2.2.1 Silver and Silver Nanoparticles

Citrate-coated AgNPs (c-AgNPs) were purchased in 2 mM citrate suspension at pH 7.6 from Ted Pella Inc. (CA, USA) with the following manufacturer's specifications: 1.14 mg/mL Ag, 2.7 x 10^{12} particles/mL, 40.6 ± 3 nm average Transmission Electron Microscope (TEM) diameter, 53.7 nm hydrodynamic diameter (HDD), and -40.7 mV zeta potential. AgNO₃ and all other reagents used in this research were analytical grade purchased from Thermo Fisher Scientific (MA, USA).

2.2.2 Biological Materials

Cultures of *B. subtilis* strain FB17 and *E. coli* strain OP50 were provided courtesy of the Bais Lab at the University of Delaware. Cultures were prepared from reserved glycerol stocks and plated on solid Luria-Bertani (LB) plates prior to use. *B. subtilis* strain FB17 was chosen due to its proven participation in plant-beneficial interactions with the model system *A. thaliana* (Kumar et al., 2012; Bais et al., 2004). *E. coli* strain OP50 was chosen as a well-studied common laboratory strain of the bacterium. Additionally, as a Gram-negative bacterium, *E. coli* has a different cell membrane structure than the Gram-positive *B. subtilis* (Ruparelia et al., 2008; Yoon et al., 2007). Seeds of the *Z. mays* cultivar Missouri 17 (Mo-17) were collected from the University of Delaware Greenhouse seed stock (Source ID: 09.1.19716.00359).

2.2.3 Determination of c-AgNP Morphology, Stability, and Ion Release

Characteristics important to the fate, transport, and environmental interactions of c-AgNPs as identified by Stone and colleagues (2010) were quantified prior to testing their effects on bacteria and plants. These characteristics include morphology (size, size distribution, and shape), stability, and ion release.

2.2.3.1 Morphology

Manufacturer-provided particle size and morphology was verified using a Libra 120 TEM (Zeiss AG, DK). Copper grids were treated with poly-1-lysine for 15 min, washed 3x with NanoPure water and dried before being incubated over the AgNP solution for 1 h. After incubation, grids were washed 3x and subsequently dried for 1 h. Images were taken at an accelerating voltage of 120kV. Collected images were processed and analyzed for average TEM diameter and shape using ImageJ v.1.46 software (National Institutes of Health, MD, USA).

2.2.3.2 Stability

Particle stability was determined by measuring HDD and zeta potential of 5.0 mg/L c-AgNP solutions in LB-liquid, basal Plant Nutrient Solution (PNS), and a solution of maize excretes in PNS (PNS+RE). A complete composition of PNS is found in Appendix A (Seyfferth and Parker, 2007; Pedler et al., 2000). Root excretes were collected after growing maize roots in PNS for 7 d. The combined solution was harvested and filtered through sterile 0.22 µm PES membranes (Argos Technologies,

18

IL, USA) and preserved at 4°C. HDD was determined by dynamic light scattering (DLS) and zeta potential by electrophoretic mobility using a Mobius Mobility Instrument (Wyatt Technology, CA, USA).

2.2.3.3 Ion Release

The dissolution of AgNPs in deionized water (DIW), LB-liquid, and PNS+RE was quantified by Inductively Coupled Plasma Mass Spectroscopy (ICP-MS) using modified methods from Lee et al. (2012a) and Ma et al. (2012). Batch experiments were conducted by preparing a 1.0 mg/L AgNP solution in each medium and incubating at room temperature for 48 h in DIW and LB-liquid, and for 7 d in PNS+RE. Two 1 mL aliquots of the sample were taken at each sampling; the first was filtered through a 0.025 µm membrane filter (Millipore, MA, USA) for aqueous phase Ag concentration (Ag_{aq}) determination. The filtrate was then acidified with 9 mL of 1% v/v HNO₃ and allowed to digest for at least 24 h before ICP-MS analysis. The second aliquot was taken for total Ag concentration determination, and was immediately acidified and digested for 24 h. Solutions were analyzed on an Agilent 7500cx ICP-MS (Agilent Technologies, CA, USA). Concentrations of Ag_{aq} obtained from ICP-MS were then compared to theoretical dissolved concentrations calculated using a modified form of the Ostwald-Freundlich equation (Eq. 1):

$$S_r = S_{bulk} \times e^{(2\gamma V_m/RT \times r)} \tag{1}$$

where S_r is the solubility (mg/L) of AgNPs with radius r (m), S_{bulk} is the accepted solubility (mg/L) of a silver particle with a flat surface, γ is the surface tension of the particle (J m⁻²), V_m is the molar volume of the particle (m³/mol), R is the gas constant (8.314 J/mol*K), and T is the temperature (K) (Ma et al., 2012).

2.2.4 Bacterial Susceptibility to c-AgNPs

B. subtilis and *E. coli* were exposed to varying concentrations of c-AgNPs and the best-estimated concentrations of AgNO₃ based on literature reports and dissolution results. c-AgNPs were applied at doses of 0.1, 1.0, and 5.0 mg/L (El Badaway et al, 2011; Stebounova et al, 2011, Suresh et al, 2010) and AgNO₃ was applied at concentrations of 0.1 and 1.0 mg/L based on literature values (Ma et al, 2012; Stebounova et al, 2011; Zhang et al., 2011).

2.2.4.1 Determinations of Minimum Inhibitory Concentrations and Half Maximal Effective Doses

To determine the minimum inhibitory concentration (MIC) of c-AgNPs batch experiments were carried out in $\text{Costar}^{\&} 2 \ \mu\text{L} 96$ -well plates (Corning Inc., NY, USA). Fifty μL of freshly prepared bacterial culture was diluted in 150 μL of Ag-spiked LBliquid and incubated at 30°C and 200 rpm for 8 h. Ag-free LB-liquid was used as a control media to assess the standard growth of the bacteria. The absorbance of the cultures at 630 nm (OD₆₃₀) was taken every 1 h through the log phase of growth and every 2 h to the stationary phase using a Dynex Opsys MR Microplate Reader (Dynex Technologies, VA, USA). Bacteria-free Ag-spiked LB-liquid at each concentration was used to eliminate any artifacts of the suspended particles on the OD₆₃₀. Each treatment was replicated 8 times, and experiments were run 3 separate times to ensure reproducibility of data. Dynamic growth curves were generated from the collected data and the MIC was taken as the lowest concentration of c-AgNPs that showed growth less than the control group.

The half maximal effective concentration (EC_{50}) was calculated by averaging growth of each treatment over the full 8-h exposure and determining the percent

20

inhibition. Data was analyzed using Toxicity Relationship Analysis Program (TRAP)v.1.22 (USEPA, 2013) using the dose-response model as follows:

$$Y = \frac{Y_0}{1 + e^{4S(X - X50)}} \tag{2}$$

where *Y* is the response, Y_0 is the response of the control, *S* is the slope of the curve, *X* is the dose concentration, and X_{50} is the dose which has an effect on 50% of the population (Metzler et al., 2012; Metzler et al., 2011; Huang et al, 2010).

2.2.4.2 Growth Kinetics

Absorbance data was transformed to bacterial number by convention: 1.0 OD $= 5.0 \times 10^8$ cells/mL (Bio-Rad Laboratories, Inc., 2002) and then plotted in *Mathematica v. 9* (Wolfram Research Company, USA) and using a modified form of the logistic function (eq. 3) in order to determine the effect of c-AgNP exposure on bacterial growth kinetics. This form of the logistic function is used to model bacterial growth, and is written as:

$$N_t = \frac{K}{1 + \frac{K - N_0}{N_0} e^{-rt}}$$
(3)

where N_t is the number of bacteria at time t, K is the carrying capacity, N_0 is the initial bacterial population, all expressed as cells/mL, and r is the growth rate of the population expressed as cells mL/h. Obtained values included K, N_0 , and r. These parameters were than used to assess the impact of c-AgNPs on the bacterial population over the entire 8-h exposure.

Impacts of c-AgNPs on growth kinetics can also be assessed by comparing maximum specific growth rates of control and exposed cultures using the following equation (Schacht et al., 2013): $\mu = \frac{lnx_1 - lnx_2}{lnx_2}$ (4)

$$\mu = \frac{mx_1 + mx_2}{t_1 - t_2} \tag{4}$$

where μ is the specific growth rate between two time points (cells/mL/h), and x is the cell concentration (cells/mL) at some time t (h) (Schacht et al., 2013).

2.2.4.3 Bacteria-Particle Associations

Bacteria-particle associations were investigated using a Hitachi S4700 fieldemission Scanning Electron Microscope (SEM) (Hitachi, JP). Samples were prepared by addition of 1 mL of AgNP solution to freshly prepared cultures of bacteria with O.D.600nm ~ 2.00 and vortexing for 1 min to ensure adequate mixing of bacteria and NPs. Mixtures were then fixed in 2% gluteraldehyde for 1 h and then washed 1x PBS followed by a 1-h incubation in 1% Osmium Tetroxide. Samples were rinsed with distilled water and dehydrated in an ethanol dilution series before being dried and mounted on aluminum stubs and coated with Au/Pd on a Denton Vacuum Bench Top Turbo III sputter-coater (NJ, USA). All images were taken at 3.0kV to avoid charging of the c-AgNPs and damaging the sample.

2.2.5 Z. mays Susceptibility to c-AgNPs

Z. mays seeds were sowed in sterile Pro-Mix soil (Premier Tech Horticulture, PA, USA) and grown for 10-14 d in a greenhouse at 25°C and 65% relative humidity under natural day-night light cycles. Prior to use, seedlings were removed from soil and rinsed with tepid tap water to remove soil and soil aggregates. Root systems of the seedlings were then soaked in sterile distilled water for 30 min to further remove bound soil.

2.2.5.1 Phytotoxicity Assays

Z. mays seedlings were grown in PNS spiked with varying concentrations of c-AgNPs based on results from bacterial toxicity experiments. c-AgNPs were applied at 1.0 and 5.0 mg/L, and AgNO₃ was applied at a concentration of 0.1 mg/L. Ag-free PNS was used as a control. After rinsing and soaking, seedlings were transferred to sterile double Magenta boxes with 30 mL of sterile PNS. Boxes were then transferred to a shaker table at a rotating speed of 100 rpm and placed under a plant lighting system on a 12-h day/night light cycle, where they grew for 7 days. Each treatment was replicated 3 times and experiments were repeated 3 separate times to ensure reproducibility.

After 7 d seedlings were removed from the Magenta boxes and roots triple rinsed in sterile DIW. Morphology of roots and plant health were examined by digital photographs of the seedlings. Roots were then separated from the shoot, and the shoot discarded. Small sections of the primary root and tertiary/fine roots were taken and preserved for analysis by FE-SEM. The remaining roots were weighed for wet biomass. Roots were then dried for 48 h at 70°C (Yin et al, 2011). After removal, roots were cooled in a desiccator and dry root biomass was weighed.

Dried roots were preserved at 4°C prior to digestion in preparation of analysis for Ag content by ICP-MS. Roots were digested by adding 2 mL of concentrated trace metal grade HNO₃ and heating at 120°C for 30 min. Digestion tubes were removed from the heating block and the solution cooled to room temperature. Two mL of trace metal grade H_2O_2 (30% w/w) was then added and the tubes were reheated at 120°C for an additional 30 min (Geisler-Lee et al, 2013). The remaining solution was cooled and diluted gravimetrically with DIW. A blank and an AgNO₃ reference standard were digested following the same procedure. Digested and diluted samples were stored at 4°C prior to analysis by ICP-MS.

23

Remaining PNS+RE was collected and filtered as previously described. Collected PNS+RE solutions were acidified with 1% HNO₃ and allowed to digest for 24 h. Samples were then analyzed for total Ag by ICP-MS.

2.2.5.2 Detection of AgNPs on *Z. mays* Roots by Field Emission Scanning Electron Microscopy

Harvested *Z. mays* roots were thoroughly rinsed in sterile DIW and sliced into 1 mm sections of primary roots and root hairs for analysis under FE-SEM. Sections were fixed in a solution of 2% paraformaldehyde and 2% gluteraldehyde in 0.1M sodium cacodylate buffer at pH 7.4. Preserved sampled were stored at 4°C prior to preparation for imaging. The samples were then washed in 0.1M phosphate buffer, dehydrated in ethanol stepwise from 25% to 100% then dried in a Tousimis Autosamdri 815B critical point dryer (MD, USA). Samples were then mounted on SEM mounts and coated with carbon using a Denton Vacuum Bench Top Turbo III sputter-coater (NJ, USA). All images were taken at 3.0kV to avoid charging of any c-AgNPs and damaging the root tissue.

2.2.6 Susceptibility of Bacteria-Inoculated Z. mays to c-AgNPs

Z. mays exposures to c-AgNPs were repeated with modifications to assess the impact of exposure on plant-microbe interactions. After removal from soil and washing/soaking to remove soil particles, the root systems of *Z. mays* seedlings were inoculated with 1 mL of *B. subtilis* culture at a concentration of 1.0×10^5 cells/mL. The plants were then placed in sterile plant nutrient solution and placed under a growth lamp on a shaker table as previously described. The same endpoints of root length, mass, morphology, and Ag content were used to assess impacts *Z. mays* seedlings.

After 7-d of growth, 1 mL of the remaining nutrient solution was plated on LB-solid media and incubated for 24 h at 30°C to quantify *B. subtilis* growth.

2.2.7 Statistical Analysis

Results were analyzed using IMB SPSS Statistics Software v.22 (NY, USA). One-way Analysis of Variance (ANOVA) was used to compare means between experimental treatments, and Tukey's post-hoc tests were used to identify which treatments were different. Independent t-Tests were used to compare differences across experiments.

2.3 Results and Discussion

2.3.1 Particle Characterization

2.3.1.1 Size and Morphology

More than 600 individual c-AgNPs were identified in all collected TEM micrographs (Fig. 8A). The mean TEM diameter of c-AgNPs in stock solution was determined to be 44.9 ± 7.2 nm (Fig. 8B). Of all analyzed particles, 52.1% were in the 40-49 nm size range, and no analyzed particles were less than 25 nm or greater than 72 nm. Particles were roughly spherical with some polygonal features, such as sharp corners or defined edges (Fig. 8A).



Figure 8: A) TEM micrograph of c-AgNPs in stock solution. Scale bar is 100 nm. B) Size Distribution of c-AgNPs in stock solution. Mean TEM diameter = 44.9 ± 7.2 nm, n = 607.

2.3.1.2 Decreased Particle Stability in Tested Media

c-AgNPs were less stable in tested media compared to stock solution, as indicated by larger HDDs and less negative zeta potentials (Table 1). Particles are smallest but least stable in LB-liquid, though still more than 3x the size of c-AgNPs in stock solution. c-AgNPs are the largest but most stable in sterile PNS. c-AgNPs are ~15% smaller in PNS+RE, indicating that low molecular weight carbon compounds – such as the organic acids excreted or exuded by *Z. mays* roots – may act as a stabilizer (Akaighe et al., 2011), though this cannot be confirmed given the observed increase in zeta potential.
Solution	$HDD / HH (Avg. \pm 5D)$	$\zeta / m V (Avg. \pm 5D)$
Stock*	53.7	-40.7
LB-Liquid	180.2 ± 11.2	-12.8 ± 2.4
PNS	222.6 ± 18.2	-17.1 ± 1.8
PNS+RE	188.2 ± 11.4	-14.8 ± 3.2

Table 1: HDD and Zeta Potential of c-AgNPs in Various Media

HDD / nm (Avg + SD) ℓ / mV (Avg + SD)

*No SD was provided by the manufacturer for these values

Although c-AgNPs have larger HDDs and less stable zeta potentials, they appear stable over the duration of various experiments (solution color shifts and settling out of particles were not observed).

2.3.1.3 Ion Release in Tested Media is Negligible

Solution

Batch experiments conducted in DIW, LB-liquid, and PNS+RE revealed negligible dissolution of Ag^+ from c-AgNPs in the time period studied. Maximum dissolution occurred in DIW, while c-AgNPs suspended in LB-liquid and PNS+RE dissolved considerably less. Expected concentrations obtained using Eq. 1 were generally found to overestimate the dissolution of Ag^+ in the tested media (Ma et al., 2012). For particles with a radius of 45 nm, and assuming $S_{bulk} = 0.009$ mg/L, $\gamma = 1$ J/m², and $V_m = 6.02 \times 10^{-4}$ m³/mol, maximum S_r was calculated as 0.18 mg/L. This value is similar to the final concentration of Ag^+ observed in DIW (0.10 ± 0.02 mg/L), but roughly 50x higher than the final concentration in LB-liquid (0.004 ± 0.002 mg/L). Ag^+ concentrations in PNS+RE after 7d were below instrument detection limit. These findings agree with several authors touting the increased stability of c-AgNPs over AgNPs stabilized with other coatings (Ma et al., 2012; Kittler et al., 2010; Zhang et al., 2011).

2.3.2 Suppression of Bacterial Growth

Exposure to c-AgNPs resulted in the suppression of growth of both *B. subtilis* and *E. coli*. Eight-hour growth experiments showed differences between non-exposed control populations and populations exposed to even low concentrations of c-AgNPs. Dynamic growth curves generated from absorbance data in Fig. 9 show qualitative MICs of 0.1 mg/L c-AgNPs for both *B. subtilis* and *E. coli*. B. subtilis was significantly inhibited by exposure to 5.0 mg/L c-AgNPs (P < 0.05), while inhibitions of E. coli by c-AgNPs were not statistically significant (P > 0.05). These observations are in general agreement with previously reported values (Suresh et al., 2010; Krishnaraj et al., 2010; Verma et al., 2010; Gade et al., 2008). *B. subtilis* was significantly more affected by exposure to 5.0 mg/L c-AgNPs than *E. coli* (P < 0.05). Additionally, dose-response effects of exposure to c-AgNPs were more pronounced in the *B. subtilis* culture than for *E. coli*, as further indicated by 8-h average inhibition and EC₅₀ calculations.

Both species were relatively unaffected by exposure to 0.1 mg/L AgNO_3 (Fig. 10), a concentration equivalent to literature-reported solubility of c-AgNPs (Ma et al., 2012; Zhang et al., 2011). The dynamic growth curves show a slight lag in the growth behavior of these cultures, specifically in the time to reach the log and stationary phases. This phenomenon has previously been observed in cultures exposed to AgNPs (Suresh et al., 2010). However, cultures had statistically significant (P < 0.05) greater cell concentrations after 8-h than control. And while the cultures outperformed

control, *E. coli* concentrations were significantly greater than *B. subtilis* (P < 0.05). Both species were completely growth-inhibited by exposure to 1.0 mg/L AgNO₃. Complete toxicity at this level of Ag⁺ has been previously observed (Li et al., 1997) and is a testament to the incredible antimicrobial power of this form of silver. In fact, toxicity was so high that further kinetic analysis was made impossible, as growth did not fit the logistic model of growth.

c-AgNPs were significantly more toxic to both species of bacteria than equivalent concentrations of soluble Ag (P < 0.05). These findings are in agreement with previously reported data (El Badawy et al., 2011; Suresh et al., 2010). The difference in toxicity between equal concentrations of c-AgNPs and Ag⁺ is clear, as the latter was 5x more toxic to *B. subtilis* and 12x more toxic to *E. coli*.



Figure 9: A) Dynamic growth curve for *B. subtilis* exposed to c-AgNPs; B) Dynamic growth curve for *E. coli* exposed to c-AgNPs.



Figure 10: A) Dynamic growth curve for *B. subtilis* exposed to AgNO₃; B) Dynamic growth curve for *E. coli* exposed to AgNO₃.

Calculated EC_{50} values further demonstrate the increased susceptibility of *B*. subtilis to c-AgNPs compared to E. coli. The Toxicity Relationship Analysis Program (TRAP) v. 1.22 (USEPA, 2013) was used to determine the EC₅₀ for c-AgNPs for both species of bacteria; B. subtilis: 6.5 ± 3.6 mg/L, E. coli: 10.9 ± 5.8 mg/L. In each instance, the calculated EC_{50} value is higher than any of the experimental concentrations used in this study, which reduced the predictive power of the software (as indicated in the large standard error). Thus it may prove more beneficial to examine the effects of c-AgNPs on the growth kinetic behavior of each species (Table 2). The carrying capacity (K) for both species of bacteria was significantly reduced by exposure to 5.0 mg/L c-AgNPs, while the 0.1 mg/L AgNO₃ had a significantly higher K than the untreated control. No significant differences were observed for the 0.1 and 1.0 mg/L c-AgNP treatments. However, there is still an observable dose-response effect for *B. subtilis* that mirrors the response observed in the MIC experiments. The weak dose-response observed in the MIC experiments for *E. coli* is also seen in *K*. Growth rates (r) generally followed the same trend as K. In logistic growth, r tends to decrease as populations increase; thus higher growth rates are associated with smaller populations typifying boom and bust growth behavior (Vandermeer, 2010). This explains the general decrease in r observed in cultures treated with increasing concentrations of c-AgNPs. However, this trend was not statistically significant due to relatively high degrees of variability inherited during the curve-fitting process. Using μ_{max} instead of r alleviates some of the fitting variability associated with the logistic function by focusing on growth between only two time points (Schacht et al., 2013). The trend in r is also apparent in μ_{max} , with a general decrease across increasing c-

AgNP concentrations; μ_{max} was significantly greater for the cultures exposed to 5.0 mg/L c-AgNP for both species of bacteria.

		Parameter (Avg. \pm SD)				
Treatment		<i>K</i> *	г*	$\mu_{max}^{\#}$		
B. subtilis	Control	3.72 ± 0.21	1.181 ± 0.230	0.999 ± 0.285		
	0.1 mg/L c-AgNP	3.75 ± 0.26	1.148 ± 0.270	0.931 ± 0.207		
	1.0 mg/L c-AgNP	3.76 ± 0.33	1.193 ± 0.295	1.029 ± 0.231		
	5.0 mg/L c-AgNP	$3.67\pm0.25^\dagger$	1.321 ± 0.288	$1.584\pm0.674^\dagger$		
	0.1 mg/L AgNO ₃	$3.96\pm0.27^{\dagger}$	$0.981\pm0.176^{\dagger}$	0.901 ± 0.374		
E. coli	Control	3.18 ± 0.19	1.007 ± 0.120	0.850 ± 0.103		
	0.1 mg/L c-AgNP	$3.02\pm0.23^{\dagger}$	0.998 ± 0.153	0.840 ± 0.166		
	1.0 mg/L c-AgNP	3.03 ± 0.20	0.966 ± 0.103	0.926 ± 0.174		
	5.0 mg/L c-AgNP	$2.97\pm0.23^{\dagger}$	1.057 ± 0.134	$1.501\pm0.427^\dagger$		
	0.1 mg/L AgNO ₃	$3.42\pm0.18^\dagger$	$0.915\pm0.042^\dagger$	0.815 ± 0.139		
*Logistic function parameters: $K: 1.0 \times 10^8$ cells/mL and r . cells/mL/h						
[#] Specific growth rate parameters: μ_{max} : cells/mL [/] h						

Table 2: Kinetics Parameters

[†]Statistically different from control (P < 0.05)

The alteration of kinetics due to exposure to c-AgNPs is most significant for cultures treated with 5.0 mg/L c-AgNPs. Although dose-response was not as strong for *E. coli* as for *B. subtilis*, the general suppression of *K* and elevation of *r* and μ_{max}

indicate a suppressive effect due to exposure to c-AgNPs. Treatment with low levels of AgNO₃ resulted in better performing cultures, while higher concentrations led to complete inhibition of bacterial growth. The added nitrate may be responsible for the increased performance of the bacterial culture exposed to 0.1 mg/L AgNO₃, considering its extremely low background concentration in the untreated LB-liquid. Overall though, these findings support previous studies that found high levels of AgNO₃ to be extremely toxic to bacteria (Suresh et al., 2010; Li et al., 1997).

2.3.3 c-AgNPs Sorb onto Bacterial Cell Surfaces

c-AgNPs and small aggregates sorbed to both species of bacteria during exposures (Fig. 11). The differing surface charges of the bacteria did not appear to alter sorption behavior, and sorption of c-AgNPs did not induce any observable morphology changes on cellular surfaces. This lack of impact on cellular surfaces is contrary to previously reported data. El Badawy et al. (2011), Suresh et al. (2010) and Choi and Hu (2009) showed significant impacts on cell membranes of different bacteria types due to exposure to AgNPs. Impacts include loss of cell height and restricted morphology, as well as cell wall pitting and formation of lumps. Possible explanations for differing results include a lack of resolution due to charging of cellular surfaces, as well as shorter incubation times prior to sample analysis.

Interestingly, the citrate coating of the particles may have provided an additional carbon source the bacteria, as several colonies formed on top of masses of c-AgNP aggregates (Fig. 11B and D). This finding provides important insight into the role of particle coatings to overall behavior, especially toxicity. Further investigation is necessary to quantify the utilization of the citrate component of c-AgNPs by

bacteria, and to determine the time-dependent effects of any utilization and resulting exposure of fresh Ag surfaces that may exert toxicity.



Figure 11: FE-SEM micrographs of *B. subtilis* (A) and (B) and *E. coli* (C) and (D). Arrows denote c-AgNPs or aggregates

2.3.4 Comparing Effects of Exposure between Bacterial Species

Experimental results revealed that the highest concentration of c-AgNPs had a more pronounced effect on *B. subtilis* than *E. coli*. Additionally, the EC₅₀ value for *E. coli* was nearly double that for *B. subtilis*. This phenomenon is somewhat in contrast to previous findings that indicated the negative surface charge of c-AgNPs limits their

toxic effects to Gram positive bacteria such as *B. subtilis* and increasing their effects on Gram negative bacteria such as *E. coli* (El Badawy et al., 2011). The increasingly touted mechanism for surface-charge dependent associations of AgNPs and bacterial surfaces does not completely explain the observed toxicity. Under this mechanism, the lipopolysaccharide-rich cell membrane of Gram negative bacterial species tends to have a neutral to slightly positive surface charge, thereby increasing cell-particle interactions and resulting toxicity (El Badawy et al., 2011). The carboxyl, phosphate, and amino groups present in the cell membranes of Gram positive bacterial species results in a strongly negative surface charge (van der Wal et al., 1997), supposedly increasing repulsion and limiting cell-particle interactions.

This phenomenon was not observed in the tested bacterial cultures; instead, these findings support previous data showing that the lipopolysaccharides of the outer membrane of Gram-negative bacteria provides innate resistance to NPs (Suresh et al., Ruparelia et al., 2008; Yoon et al., 2007; Brayner et al., 2006). Comparatively, the lack of a protective outer membrane and periplasmic space in Gram-postive bacteria such as *B. subtilis* may increase the likelihood of cell-particle interactions compared to Gram-negative species.

Another facet of the observed toxicity is related to the carbon-rich citrate coating. It is well known in the microbial ecology field that citrate is a ubiquitous compound that acts as a carbon source for both species tested here (Brocker et al., 2009; Yamamoto and Sekiguchi, 2000; Meyer et al., 2001, 1997; Bott et al., 1995), and thus it is not surprising to observe such behavior. This phenomenon may override surface charge dependent interactions and the impacts of cellular membrane differences, and toxicity may simply occur as bacteria exploit the citrate coating for its

carbon, exposing fresh and highly toxic Ag surfaces (Fig. 11B and D) or releasing Ag⁺ into solution.

2.3.5 Phytotoxicity of AgNPs on Z. mays roots

Exposure to c-AgNPs results in sub-lethal phytotoxicity to *Z. mays* seedlings, including reduced root length and biomass, (Figs. 12 and 13). Control plants had a root length of 17.0 ± 6.1 cm and a root biomass of 0.579 ± 0.307 g. Plants exposed to 1.0 mg/L AgNPs had respective lengths and biomass of 13.2 ± 6.8 cm and 0.454 ± 0.307 g. Increasing the c-AgNP concentration to 5.0 mg/L resulted in a root length of 12.9 ± 4.1 cm and a root biomass of 0.480 ± 0.144 g. Treatment with 0.1 mg/L Ag⁺ as AgNO₃ resulted in the shortest roots (11.1 ± 0.9 cm) and lowest root biomass (0.418 ± 0.222 g).



Figure 12: A) Digital photograph of *Z. mays* roots after 7-d exposure; B) Digital photograph of *B. subtilis*-inoculated *Z. mays* seedlings after 7-d exposure.



Figure 13: Root length (A) and Wet Root Biomass (B) of *Z. mays* seedlings after 7-d exposures. Error bars are ± 1 SD.

While a limited dose-response is evident, differences between control and Ag treatments were not statistically significant (P > 0.05), likely due to relatively high

degrees of variability within treatments. However, these same general trends and phenomena have been observed in a variety of plant species exposed to AgNPs, including *L. multiflorum* (Yin et al, 2011) and the model system *A. thaliana* (Geisler-Lee et al, 2013). Yin et al. found no difference in root length and biomass between control plants and those exposed to 1.0 mg/L AgNP. However, they did show some significant effects when increasing exposure concentrations to 5.0 mg/L and above. El-Temsah and Joner found that exposure to AgNPs of various sizes resulted in significant decreases in shoot length of *L. perenne*, although root length was not measured (2010).

Ag content of *Z. mays* roots increased with c-AgNP treatment concentration (Fig. 14). Control plants had Ag contents lower than the instrument detection limit. Roots of seedlings exposed to 1.0 mg/L AgNP had an average Ag content of $2.44 \pm 2.90 \times 10^{-5}$ mg. Increasing the exposure concentration to 5.0 mg/L led to an average root Ag content of $1.15 \pm 2.03 \times 10^{-4}$ mg, while treatment with 0.1 mg/L AgNO₃ led to an average root Ag content of $1.92 \pm 5.78 \times 10^{-6}$ mg. Seedlings grown in 5.0 mg/L c-AgNP nutrient solution accumulated more Ag than all other treatments, but differences were not statistically significant (*P* > 0.05).



Figure 14: Ag content of Z. mays roots after 7-d exposure to c-AgNPs. Error bars are ± 1 SD.

Yin et al. (2011) found a similar trend in concentrations of Ag in the roots of exposed *L. multiflorum* plants. They also observed a weak-dose response in the bioconcentration factor for plants exposed to a range of [Ag] for gum-arabic coated 6 nm AgNPs. The authors found greater [Ag] in their root systems at similar concentrations than in the present study, likely due to the large size difference (6 nm compared to 45 nm). Macroscopic root features of *Z. mays* seedlings (Fig. 12) are consistent with previous findings, are likely representative of smaller-scale alterations in morphology, including shortened or missing root hairs, highly vacuolated and collapsed cortical cells, and broken epidermal and root cap cells (Geisler-Lee et al., 2013; Yin et al., 2011).

2.3.6 Stunting of Tertiary and Fine Roots

c-AgNP treated plants had fewer tertiary roots than control plants (Fig. 12). Biomass differences are not necessarily reflective of this due to the overwhelmingly larger mass of primary and secondary root structures. And while primary and secondary root structure and architecture were not impacted by exposure to c-AgNPs, fine roots were significantly impacted by exposure to AgNPs (Fig. 12). Stunting of these roots is a common response in plants exposed to AgNPs (Geisler-Lee et al., 2013; Yin et al., 2011). Impacts on the development of tertiary roots and root hair structures, though not as significant as damages to primary or secondary roots, reduces the plant's ability to regulate moisture and nutrient uptake (Eissenstat et al., 2000; Davies and Zhang, 1991), and their ability to communicate through excretions and exudations (Zobel, 2005).

These same authors have shown "browning" of root tips caused by exposure to AgNPs. It is unknown whether this phenomenon is the result of uptake of AgNPs or by complexation of Ag^+ with secondary plant compounds. However, significant localized uptake of AgNPs into the intracellular spaces within the root cap and associated border cells was observed, indicating some nano-specific impact (Geisler-Lee et al., 2013).

2.3.7 Decreased Beneficial Interactions in Bacteria-Inoculated Z. mays Exposed to c-AgNPs

Inoculation of *Z. mays* seedlings with *B. subtilis* resulted in increased root length and root biomass for all treatments (Figs. 12 and 13). Root length and biomass increased 35% and 15% respectively between control groups. Treatment of inoculated *Z. mays* seedlings with c-AgNPs reduced the root length and biomass compared to control. Root length and biomass were still greater than non-inoculated seedlings,

indicating that treatment with bacteria can mitigate the effects of c-AgNPs through general growth promotion. However, increases in root length and biomass were less significant for Ag-treated seedlings than control (24% and 15% respectively for seedlings treated with 5.0 mg/L c-AgNP). Inoculation with *B. subtilis* dramatically increased the root length and biomass of *Z. mays* seedlings treated with 0.1 mg/L AgNO₃ (68% and 34%, respectively). While dose-response is clear, differences in root length and biomass were not statistically significant between treatments or between experiments for the control group and seedlings exposed to 1.0 mg/L c-AgNPs (P > 0.05). Biomass differences were significantly different for seedlings exposed to 5.0 mg/Lc-AgNPs, while differences in root length were significantly different for those exposed to 0.1 mg/L AgNO₃.

Interestingly, roots of seedlings inoculated with *B. subtilis* had higher Ag contents than seedlings alone across all concentrations (Fig. 14). Control group seedlings again had Ag contents less than the instrument detection limit. Treatment with 1.0 and 5.0 mg/L c-AgNPs resulted in root Ag contents of $1.72 \pm 0.35 \times 10^{-4}$ mg and $4.63 \pm 2.99 \times 10^{-4}$ mg, respectively. The increase in root Ag content for seedlings exposed to c-AgNPs compare to control was not statistically significant. Seedlings exposed to 0.1 mg/L AgNO₃ had root Ag contents of $0.16 \pm 3.36 \times 10^{-5}$ mg. These contents represent marked increases of over non-inoculated seedlings. Increases in root Ag content were statistically significant for seedlings exposed to 5.0 mg/L c-AgNPs (P < 0.05). However, all other increases were not statistically significant.



Figure 15: FE-SEM micrographs of *Z. mays* roots. A-C) Control; D-F) 5.0 mg/L c-AgNP.

Analysis of roots by FE-SEM (Fig. 15) showed large aggregates of c-AgNPs on root surfaces (Fig. 15F). Smaller particles were sorbed in smaller random clusters over surfaces. In each case, the presence of c-AgNPs or aggregates was also associated with bacterial populations (Fig. 15F). Images are markedly similar to those taken of bacteria after incubation with c-AgNPs (Fig. 11). Like c-AgNPs, bacteria were mostly randomly distributed across root surfaces (Fig. 15C), but were more concentrated in areas of greater local topography. There were no observable differences in c-AgNP sorption or bacterial distribution between primary roots and tertiary roots or root hairs (images not shown).

Bacterial concentrations in solution were also reduced by treatment with c-AgNPs (Fig 16). Control group bacteria numbered $9.39 \times 10^7 \pm 3.57 \times 10^7$ cells/mL. Treatment with c-AgNPs reduced numbers to $5.83 \times 10^7 \pm 0.80 \times 10^7$ and $3.92 \times 10^7 \pm 0.60 \times 10^7$ cells/mL for 1.0 mg/L and 5.0 mg/L, respectively. This reduction is significantly different from control (P < 0.05). Bacteria exposed to 0.1 mg/L c-AgNO₃ numbered 7.38 $\times 10^7 \pm 1.48 \times 10^7$ cells/mL. FE-SEM images show qualitative evidence of reduced bacterial numbers, though quantitative differences could not be determined (Fig 15). Inhibition levels of bacteria in these inoculation experiments were greater than in previous experiments (Fig. 9A), which showed inhibitions of $11.0 \pm 6.0\%$ and 33.3 ± 23.0 for *B. subtilis* exposed to 1.0 and 5.0 mg/L c-AgNPs, respectively. Bacteria inoculated on *Z. mays* seedlings showed inhibitions of $42.0 \pm 11.4\%$ at 1.0 mg/L c-AgNP, and $56 \pm 11\%$ at 5.0 mg/L c-AgNP (Fig. 16). Additionally, bacteria exposed to 0.1 mg/L AgNO₃ did not outperform control in inoculation experiments as in previous growth inhibition experiments (Figs. 9A and 16).



Figure 16: Concentrations of bacteria in PNS+RE after 7-d exposure. Error bars are ± 1 SD.

A possible simple mechanism for reduced bacterial populations is competition for binding/colony sites on the root surfaces. As the c-AgNPs sorb to the roots of *Z*. *mays* seedlings, the effective available area for bacterial colonization is reduced. It is well known that bacteria have preferred root sites for colonization and that their ability to form and maintain stable relationships with their plant hosts is strongly linked to their colonization ability (Bais et al., 2004; Lutenberg and Dekkers, 1999; Schippers et al., 1987). Thus available surface area reductions (by sorption of c-AgNPs, natural soil particles, or other microbes) can have a significant impact on the colonization effort (Schippers et al., 1987). Our data show limited evidence for this mechanism; however the magnitude of the root sites occupied by c-AgNPs is comparably smaller than those free from c-AgNPs. Still, these findings demonstrate the ecological significance of exposure to even low levels of c-AgNPs and are indicative of the delicacies required to maintain a healthy and mutually beneficial relationship.

2.4 Conclusion

Though c-AgNPs have been routinely found to be less toxic than AgNPs with other coatings (Sharma et al, 2014), they impart sub-lethal toxicity on both Gram positive and Gram negative bacteria and the major crop plant *Z. mays*. These sub-lethal effects of exposure to c-AgNPs ultimately resulted in dramatically reduced beneficial interactions between *Z. mays* and the bacterium *B. subtilis*. Reductions in such beneficial plant-microbe interactions are of great concern due to the indiscriminate nature of silver's antimicrobial activity. While the mechanism of toxicity to both bacteria and plants remains unknown, increasing evidence points towards a combination of effects imparted by AgNPs and ionic Ag alike. Further research is needed to elucidate the exact mechanism of reductions in beneficial interactions between soil microbes and their plant hosts.

Chapter 3

CONCLUSION AND SYNTHESIS

3.1 Introduction

Because the use of AgNPs in consumer and other types of products has risen dramatically, their release into the environment has become an increasing concern to scientists and policy makers (Gottschalk et al., 2009; Mueller and Nowack, 2008; Nowack and Bucheli, 2007; USEPA, 2012). And while their impacts on aquatic organisms and certain species of bacteria and plants are becoming increasingly wellknown, the mechanism and degree of toxicity remain unclear. Additionally, the effects of AgNPs on inter-species relationships have yet to be studied. In order to accurately assess the impacts of c-AgNPs on such relationships and begin to gain an understanding of how they may impact whole soil ecosystem processes, their impacts on individual organisms must first be well-understood and quantified.

In this research, the impacts of exposure to c-AgNPs on bacterial species *B. subtilis*, *E. coli*, and the major crop plant *Z. mays* were quantified and characterized according to concentration as well as particle characteristics (Stone et al., 2010). Furthermore, the effects of c-AgNPs on the beneficial plant-microbe interaction between *Z. mays* and *B. subtilis* were similarly quantified and characterized.

3.2 Summary of Findings and Connection to Current Scientific Understanding

Experimental results demonstrate that c-AgNPs exert significant sub-lethal toxicity to bacteria as well as plants. The findings are in general agreement with the

literature, and the demonstration of the disruption of beneficial plant-microbe interactions between *Z. mays* and *B. subtilis* is an important contribution to the growing body of knowledge concerning AgNP toxicity.

3.2.1 c-AgNP Stability and Dissolution

The nature of particle stability and dissolution are vastly important when considering the mechanism of toxicity to both bacterial species and plants. It is thus necessary to consider how measured values compare to the literature.

The decreased stability of c-AgNPs in tested solution compared to stock suspension is evidence that NPs undergo environmental transformations, as stated and diagramed (Fig. 1) by Levard et al., (2012). However, the lack of solution color change and settling out of particles indicated that c-AgNPs were relatively stable over the time periods tested, a finding in agreement with literature reports of stability on the order of weeks (Ma et al., 2012; Kittler et al, 2010).

The low dissolution of c-AgNPs to Ag⁺ in the tested media is also in general agreement with the literature (Ma et al., 2012; Kittler et al., 2010, Zhang et al., 2011). Use of the modified Ostwald-Freundlich equation (Eq. 1) accurately predicted dissolution in DIW, but overestimated dissolution in LB-liquid and PNS+RE, further supporting claims of the importance of environmental conditions to particle fate (Levard et al., 2012).

3.2.2 Bacterial Toxicity

c-AgNPs resulted in observable reductions in the growth of *B. subtilis* and *E. coli* during 8-h exposures at concentrations as low as 1.0 mg/L, and statistically significant reductions to *B. subtilis* at 5.0 mg/L c-AgNPs. Increasing concentrations to

5.0 mg/L resulted in even more significant reductions and alterations in growth kinetics of both species. These findings support literature reports of significant toxicity at such concentrations (El Badawy et al., 2011; Suresh et al., 2010).

FE-SEM images show significant sorption of c-AgNPs or aggregates onto bacterial surfaces, as previously observed (El Badawy et al., 2011; Suresh et al., 2010; Choi and Hu, 2009). However, cell membrane damage was not observed in the analyzed images, contrary to prior findings. Images do show that both species may initially colonize the c-AgNPs, making use of the easily available carbon before succumbing to the toxic effects of the underlying Ag (Brocker et al., 2009; Yamamoto and Sekiguchi, 2000; Meyer et al., 2001, 1997; Bott et al., 1995).

The more significant impacts of c-AgNP exposure on the Gram-positive *B*. subtilis may be more related to its cellular membrane structure than composition and surface charge (Suresh et al., 2010). Gram-positive species lack a protective outer membrane and periplasmic space present in the cellular membranes of Gram-negative species such as *E. coli*, and this triple layer membrane offers innate protection against effects of c-AgNPs or Ag⁺.(Suresh et al., 2010; Ruparelia et al., 2008, Yoon et al., 2007; Brayner et al., 2006). The relatively thin and less rigid cell membrane leaves Gram-positive species more susceptible to sorption of NPs, subsequent morphological changes including cell wall pitting, as well as diffusion of Ag⁺ across the membrane to inside the cytoplasm (El Badawy et al., 2011, Choi and Hu, 2009).

AgNO₃ applied at 0.1 mg/L (a concentration equal to the predicted maximum soluble fraction of c-AgNPs) resulted in statistically significant increases in populations for both bacterial species. Increasing AgNO₃ concentrations to 1.0 mg/L resulted in complete culture collapse, a well-observed and reported phenomenon

(Suresh et al., 2010; Li et al., 1997). Combined with the significant sorption of c-AgNPs onto surfaces, these data suggest that the primary mechanism of toxicity is related to nano-scale properties, not Ag^+ . However, the potential breakdown of the stabilizing citrate coating by bacteria may introduce fresh Ag^+ into solution at rates greater than predicted or measured in dissolution experiments.

3.2.3 Sub-lethal Toxicity to Z. mays

Reductions in root length and biomass of *Z. mays* seedlings exposed to c-AgNPs are in agreement with studies of such impacts on other major plant species. Results mirror findings by Yin et al. (2011) that significant responses were not observed until Ag concentrations reached 5.0 mg/L. Accumulation of Ag in root tissues at concentrations greater than the treatment solution are indicative of some degree of bioconcentration, a phenomenon also observed by other authors (Geisler-Lee et al., 2013; Yin et al., 2011).

3.2.4 Impacts on Plant-Microbe Interactions

Inoculation of *Z. mays* seedlings with *B. subtilis* resulted in observable increases in biomass and root length across all experimental treatments, indicating at least general growth promotion (Beauregard et al., 2013; Kumar et al., 2012; Mohamed and Gomaa, 2012; Bais et al., 2004). However, inoculated seedlings exposed to AgNPs showed less dramatic increases in endpoints. Increases in root Ag content compared to non-inoculated plants may be related to the accelerated breakdown of the citrate coating by bacterial activity and subsequent increased availability. However, increased root Ag content did not increase sub-lethal effects compared to non-inoculated seedlings, except at 5.0 mg/L c-AgNPs. These findings are in general agreement with previously reported instances of disruptions of beneficial plant-microbe relationships (Priester et al., 2012; Chen et al., 2003), though few studies of this nature exist.

3.3 Limitations and Recommendations for Future Research

There are two categories of potential limitations to this research. The first concerns limited test subjects; bacterial toxicity was tested on only two species of bacteria, and phytotoxicity on only one species, albeit a major crop plant. While the bacteria tested represent the two major classes (Gram-positive and Gram-negative), a more diverse selection of species within these classes may have provided additional insight into specific toxicity mechanisms, as well as inter- and intra-species response to c-AgNP exposure. Testing the toxicity of c-AgNPs to other major crop plants such as rice (*Oryza* spp.), wheat (*Triticum* spp.) and soy (*Glycine* spp.) would increase the broader impacts of this research. Additionally, testing these species may show differences in responses between monocots and dicots, as well as grain and legume species.

The second stems from the narrow concentration range, comparatively large size of tested c-AgNPs, and the use of only one type of coated AgNP. The narrow concentration range was most constraining when calculating the EC_{50} values for *B*. *subtilis* and *E*. *coli*, where c-AgNP concentrations were not high enough to reduce bacterial populations by 50%. Higher concentrations also may have resulted in greater sub-lethal toxicity in exposed *Z*. *mays* seedlings. Other authors used concentrations have demonstrated significant effects of exposure at concentrations between 10 and 50 mg/L, depending on the size of the NPs used (Yin et al., 2011). However, using higher concentrations reduces the applicability to natural systems, thus our lower

concentrations are still justified. The (relatively) large size may have been responsible for the lower degree of toxicity observed in experiments compared to the literature. Several authors have shown that smaller particles (ranging from 6 to 20 nm in diameter) have the most impact on bacteria and plants alike (Geisler-Lee et al., 2013; El Badawy et al., 2011; Yin et al., 2011; Choi and Hu, 2009). As previously mentioned, citrate has been shown by several authors to create the most stable and least bio-interactive particles (Ma et al., 2012; El Badawy et al., 2011; Yang et al., 2012). Using other coatings, such as polyvinylpyrrolidone (PVP), polyvinyl alcohol (PVA), gum-arabic (GA), or branched polyethyleneimine (BPEI) may have produced different results owing to altered surface properties and dissolution behavior.

Additional sources of error and limitations include cross-contamination of 96well plates in MIC experiments that reduced usable data, and fungal contamination of *Z. mays* seedlings during exposures that likewise reduced sample size.

3.4 Conclusion

This research contributes to the growing evidence for a new paradigm concerning NP toxicity in general in which it is not necessary to discriminate between NP-specific and ionic-specific toxicity effects. Because NPs have different dissolution characteristics than their bulk counterparts (Ma et al, 2010; Kittler et al, 2008), the release of ionic species from NPs is novel property in and of its self. Additionally, the interaction of these coatings with biological materials may further alter particle characteristics. What is critical, however, is that c-AgNPs did impart toxic effects on *B. subtilis, E. coli*, and *Z. mays*, as well as the beneficial plant-microbe interactions between *Z. mays* and *B. subtilis*.

REFERENCES

- Ahamed, M., M. Karns, M. Goodson, J. Rowe, S. M. Hussain, J. J. Schlager, and Y. Hong. 2008. DNA damage response to different surface chemistry of silver nanoparticles in mammalian cells. *Toxicology and Applied Pharmacology* 233 (3): 404-10.
- Bais, Harsh P., Ray Fall, and Jorge M. Vivanco. 2004. Biocontrol of *Bacillus subtilis* against infection of Arabidopsis roots by *Psuedomonas syringae* is facilitated by biofilm formation and surfactin production. *Plant Physiology* 134: 307-19.
- Beauregard, Pascale B., Yunrong Chai, Hera Vlamakis, Richard Losick, and Roberto Kolter. 2013. *Bacillus subtilis* biofilm induction by plant polysaccharides. *Proceedings of the National Academy of Sciences* 110 (17) (April 23): E1621-30.
- Benn, Troy M., and Paul Westerhoff. 2008. Nanoparticle silver released into water from commercially available sock fabrics. *Environmental Science and Technology* 42: 4133.
- Berg, Gabrielle. 2009. Plant-microbe interaction promoting plant growth and health: Perspectives for controlled use of microorganisms in agriculture. *Applied Microbiology and Biotechnology* 84 (1): 11.
- Berg, Gabriele, Annette Krechel, Michaela Ditz, Richard A. Sikora, Andreas Ulrich, and Johannes Hallmann. 2005. Endophytic and ectophytic potato-associated bacterial communities differ in structure and antagonistic function against plant pathogenic fungi. *FEMS Microbiology Ecology* 51 (2): 215-29.
- Bharadway, Punita S. 2012. Silver or silver nanoparticle: A safety or a risk. *Journal of Environmental Research and Development* 7 (1A): 452.
- Blaser, Sabine A., Martin Scheringer, Matthew Macleod, and Konrad Hungerbuhler.
 2008. Estimation of cumulative aquatic exposure and risk due to silver:
 Contribution of nano-functionalized plastics and textiles. *Science of the Total Environment* 390: 396.

- Bott, M., M. Meyer, and P. Dimroth. 1995. Regulation of anaerobic citrate metabolism in *Klebiseilla pneumoniae*. *Molecular Microbiology* 18: 533-46.
- Brayner, R., R. Ferrari-Iliou, N. Brivois, S. Djediat, M. F. Benedetti, and F. Fievet. 2006. Toxicological impact studies on *Escherichia coli* bacteria in ultrafine ZnO nanoparticles colloidal medium. *Nano Letters* 6 (4): 866-70.
- Brocker, Melanie, Steffen Schaffer, Christina Mack, and Michael Bott. 2009. Citrate utilization by *Corynebacterium glutamicum* is controlled by the *cit*AB two-component system through positive regulation of the citrate transport genes *cit*H and *tct*CBA. *Journal of Bacteriology* 191 (12): 3869-80.
- Chen Y.X., Y. F. He, Y. Yang, Y. L. Yu, S. J. Zheng, G.M. Tian, Y.M. Luo, and M.H. Wong. 2003. Effect of cadmium on nodulation and N₂-fixation of soybean in contaminated soils. *Chemosphere* 50 (6): 781-7.
- Choi, O. K., and Z. Q. Hu. 2009. Nitrification inhibition by silver nanoparticles. *Water Science and Technology* 59 (9): 1699.
- Choi, Okkyoung, Thomas E. Clevenger, Baolin Deng, Rao Y. Surampalli, Jr Ross Louis, and Zhiqiang Hu. 2009. Role of sulfide and ligand strength in controlling nanosilver toxicity. *Journal of Water Research* 43: 1879.
- Choi, Okkyoung, Kathy Kanjun Deng, Nam-Jung Kim, Jr Ross Louis, Rao Y. Surampalli, and Zhiqiang Hu. 2008. The inhibitory effects of silver nanoparticles, silver ions, and silver chloride colloids on microbial growth. *Water Research* 42: 3066.
- Choi, Okkyoung, and Zhiqiang Hu. 2008. Size dependent and reactive oxygen species related nanosilver toxicity to nitrifying bacteria. *Environmental Science and Technology* 42 (12): 4583.
- Davies, W J,Zhang, J., 1991. Root signals and the regulation of growth and development of plants in drying soil. *Annual Review of Plant Physiology and Plant Molecular Biology* 42: 55-76.
- Dietz, Karl-Josef, and Simone Herth. 2011. Plant nanotoxicology. *Trends in Plant Science* 16 (11): 582.
- Dimpka, Christian O., Joan E. McLean, Nicole Martineau, David W. Britt, Richard Haverkamp, and Anne J. Anderson. 2013. Silver nanoparticles disrupt wheat (*Triticum aestivum* L.) growth in a sand matrix. *Environmental Science and Technology* 47: 1082-90.

- Eissenstat, D. M., C. E. Wells, R. D. Yanai, and J. L. Whitbeck. 2000. Building roots in a changing environment: Implications for root longevity. *New Phytologist* 147 (1): 33-42.
- El Badawy, Amro M., Bendahandi G. Silva, Brian Morris, Kirk G. Scheckel, Makram T. Suidan, and Thabet M. Tolaymat. 2011. Surface charge-dependent toxicity of silver nanoparticles. *Environmental Science and Technology* 45 (1): 283.
- El-Temsah, Yehia Sayed, and Erik J. Joner. 2012. Impact of Fe and Ag nanoparticles on seed germination and differences in bioavailability during exposure in aqueous suspension and soil. *Environmental Toxicology* 27 (1): 42.
- Fabrega, Julia, Samuel N. Luoma, Charles R. Tyler, Tamara S. Galloway, and Jamie R. Lead. 2011. Silver nanoparticles: Behaviour and effects in the aquatic environment. *Environment International* 37 : 517.
- Gade, A. K., P. Bonde, A. P. Ingle, P. D. Marcato, N. Duran, and M. K. Rai. 2008. Exploitation of Aspergillus niger for synthesis of silver nanoparticles. Journal of Biobased Materials and Bioenergy 2 (3): 243-7.
- Geisler-Lee, Jane, Qiang Wang, Ying Yao, Wen Zhang, Matt Geisler, Li Kungang, Ying Huang, Yongsheng Chen, Andrei Kolmakov, and Xingmao Ma. 2013. Phytotoxicity, accumulation, and transport of silver nanoparticles by *Arabidopsis thaliana*. *Nanotoxicology* 7 (3): 323.
- Gottschalk, Fadri, Tobias Sonderer, Roland W. Scholz, and Bernd Nowack. 2009.
 Modeled environmental concentrations of engineered nanomaterials (TiO₂, ZnO, Ag, CNT, fullerenes) for different regions. *Environmental Science and Technology* 43: 9216.
- Huang, Chin-Pao, Hsun-Wen Chou, Yao-hsing Tseng, and Maohong Fan. 2010.
 Responses of *Ceriodaphnia dubia* to photocatalytic nano-titanium dioxide particles. In *Environanotechnology*., eds. Maohong Fan, Chin-Pao Huang, Alan E. Bland, Zhonglin Wang, R. Achid Slimane and Ing G. Wright, 1-21. Oxford, UK: Elsevier Publishing.
- Kim, Shin Woong, Sun-Hwa Nam, and Youn-Joo An. 2012. Interaction of silver nanoparticles with biological surfaces of *Caenorhabditis elegans*. *Ecotoxicology and Environmental Safety* 77: 64.
- Kim, Soohee, Ji Eun Choi, Jinhee Choi, Jyu-Hyuck Chung, Kwangsik Park, Jongheop Yi, and Doug-Young Ryu. 2009. Oxidateive stress-dependent toxicity of silver nanoparticles in human hepatoma cells. *Toxicology in Vitro* 23: 1076.

- Kittler, S., C. Greulich, J. Diendorf, M. Koller, and M. Epple. 2010. Toxicity of silver nanoparticles increasing during storage because of slow dissolution under release of silver ions. *Chemistry of Materials* 22: 4548.
- Klaine, Stephen J., Pedro J. J. Alvarez, Graeme E. Batley, Teresa F. Fernandes, Richard D. Handy, Delina Y. Lyon, Shaily Mahendra, Michael J. McLaughlin, and Jamie R. Lead. 2008. Nanomaterial in the environment: Behavior, fate, bioavailability, and effects. *Environmental Toxicology and Chemistry* 27 (9): 1825.
- Knee, Michael. 1992. Sensitivity of ATPases to silver ions suggests that silver acts outside the plasma membrane to block ethylene action. *Phytochemistry* 31 (4): 1093.
- Krishnaraj, C., E. G. Jagan, S. Rajesekar, Selvakumar P., P. T. Kalaichelven, and N. Mohan. 2010. Synthesis of silver nanoparticles using *Acalypha indica* leaf extracts and its antibacterial activity against water borne pathogens. *Colloids* and Surfaces B: Biointerfaces 76 (1): 50-6.
- Kumar, Amutha Sampath, Venkatachalam Lakshmana, Jeffrey L. Caplan, Deborah Powell, Kirk J. Cyzmmek, Delphis F. Levia, and Harsh P. Bais. 2012. Rhizobacteria *Bacillus subtilis* restricts foliar pathogen entry through stomata. *The Plant Journal* 72: 694-706.
- Lee, Woo-Mi, Jin Il Kwak, and Youn-Joo An. 2012a. Effect of silver nanoparticles in crop plants *Phaseolus radiatus* and *sorghum bicolor*: Media effect on phytotoxicity. *Chemosphere* 89: 491.
- Lee, Yong-Ju, Jiwon Kim, Jeehyun Oh, Sungkyu Lee, In Seok Hong, and Sang-Ho Kim. 2012b. Ion-release kinetics and ecotoxicity effects of silver nanoparticles. *Environmental Toxicology and Chemistry* 31 (1): 155.
- Levard, Clement, E. Matt Hotze, Gregory V. Lowry, and Gordon E. Brown Jr. 2012. Environmental transformation of silver nanoparticles: Impact on stability and toxicity. *Environmental Science and Technology* 46: 6900.
- Li, Xian-Zhi, Nikaido, Hiroshi, Williams,Kurt E, 1997. Silver-resistant mutants of *Escherichia coli* display active efflux of Ag⁺ and are deficient in porins. *Journal of Bacteriology*. 179 (19): 6127.
- Lok, Chun-Nam, Chi-Ming Ho, Rong Chen, Qing-Yu He, Wing-Yiu Yu, Hongzhe Sun, Paul Kwong-Ham Tam, Jen-Fu Chiu, and Chi-Ming Che. 2006. Proteomic analysis of antibacterial action of silver nanoparticles. *Journal of Proteome Research* 5: 916.

- Lutenberg, B. J. J., and L. C. Dekkers. 1999. What makes *Pseudomonas* bacteria rhizosphere competent? *Environmental Microbiology* 1: 9-13.
- Ma, Rui, Clement Levard, Stella M. Marinakos, Yingwen Chen, Jie Liu, F. Marc Michel, Gordon E. Brown Jr., and Gregory V. Lowry. 2012. Size-controlled dissolution of organic-coated silver nanoparticles. *Environmental Science and Technology* 46: 752.
- Ma, Xingmao, Jane Geiser-Lee, Yang Deng, and Andrei Kolmakov. 2010. Interactions between engineered nanoparticles (ENPs) and plants: Phytotoxicity, uptake and accumulation. *Science of the Total Environment* 408: 3053.
- Maynard, A., and Michelson, E. The nanotechnology consumer products inventory. 2014 [cited 03/12/2014]. Available from http://www.nanotechproject.org/inventories/consumer/.
- Metzler, D. M., A. Erdem, Y. H. Tseng, and C. P. Huang. 2012. Responses of algal cells to engineered nanoparticles measured as algal cell population, chlorophyll a, and lipid peroxidation: Effect of particle size and type. *Journal of Nanotechnology*.
- Metzler, David M., Minghua Li, Ayca Erdem, and C. P. Huang. 2011. Responses of algae to photocatalytic nano-TiO₂ particles with an emphasis on the effect of particle size. *Chemical Engineering Journal* 170 (2-3): 538-46.
- Meyer, Joel N., Christopher A. Lord, Xinyu Y. Yang, Elana A. Turner, Appala R. Badireddy, Stella M. Marinakos, Ashutosh Chilkoti, Mark R. Wiesner, and Melanie Auffan. 2010. Intracellular uptake and associated toxicity of silver nanoparticles in *Caenorhabditis elegans*. *Aquatic Toxicology* 100: 140.
- Meyer, M., P. Dimroth, and M. Bott. 2001. Catabolite repression of the citrate fermentation genes in *Klebsiella pneumoniae*: Evidence for involvement of the cyclic AMP receptor protein. *Journal of Bacteriology* 183 : 5248-56.
 - ———. 1997. In vitro binding of the response regulator citB and of its carboxyterminal domain to A + T-rich DNA target sequences in the control region of the divergent citC and citS operons of Klebsiella pneumoniae. Journal of Molecular Biology 269: 719-31.
- Mohamed, H. I., and E. Z. Gomaa. 2012. Effect of plant growth promoting Bacillus subtilis and Pseudomonas fluorescens on growth and pigment composition of radish plants (*Raphanus sativus*) under NaCl stress. *Photsynthetica* 50 (2): 263-72.

- Morones, Jose R., Jose L. Elechiguerra, Alejandra Camacho, Katherine Holt, Juan B. Kouri, Jose T. Ramirez, and Miguel J. Yacaman. 2005. The bactericidal effect of silver nanoparticles. *Nanotechnology* 16: 2346.
- Mueller, Nicole C., and Bernd Nowack. 2008. Exposure modeling of engineered nanoparticles in the environment. *Environmental Science and Technology* 42: 4447.
- Navarro, Enrique, Anders Baun, Renata Behra, Nanna B. Hartmann, Juliane Filser, Ai-Jun Miao, Antonietta Quigg, Peter H. Santschi, and Laure Sigg. 2008. Environmental behavior and ecotoxicity of engineered nanoparticles to algae, plants, and fungi. *Ecotoxicology* 17: 372.
- Nel, Andre E., Lutz Madler, Darrell Velegol, Tian Xia, Eric M. V. Hoek, Ponisseril Somasundran, Fred Klaessig, Vince Castranova, and Mike Thompson. 2009. Understanding biophysicochemical interactions at the nano-bio interface. *Nature Materials* 8: 543.
- Nowack, Bernd, and T. D. Bucheli. 2007. Occurrence, behavior, and effects of silver nanoparticles in the environment. *Environmental Pollution* 150: 5.
- Nowack, Bernd, Harold F. Krug, and Murray Height. 2011. 120 years of nanosilver history: Implications for policy makers. *Environmental Science and Technology* 45: 1177.
- Nowack, Bernd, James F. Ranville, James Diamond, Julian A. Gallego-Urrea, Chris Metcalfe, Jerome Rose, Nina Horne, Albert A. Koelmans, and Stephen J. Klaine. 2012. Potential scenarios for nanomaterial release and subsequent alteration in the environment. *Environmental Toxicology and Chemistry* 31 (1): 20.
- Pedler, J. F., D. R. Parker, and D. E. Crowley. 2000. Zinc deficiency-induced phytosiderophore release by the *Tricaceae* is not consistently expressed in solution culture. *Planta* 211: 120-6.
- Pokhrel, Lok R., and Brajesh Dubey. 2013. Evaluation of developmental responses of two crop plants exposed to silver and zinc oxide nanoparticles. *Science of the Total Environment* 452-453: 321.
- Priester, John H., Yuan Ge, Randall E. Mielke, Allison M. Horst, Shelly Cole Moritz, Katherine Espinosa, and Jeff Gelb. 2012. Soybean susceptibility to manufactured nanomaterials with evidence for food quality and soil fertility interruption. *Proceedings of the National Academy of Sciences* 109 (37): 2451.

- Qian, Haifeng, Xiaofeng Peng, Xiao Han, Jie Ren, Liwei Sun, and Zhengwei Fu.
 2013. Comparison of the toxicity of silver nanoparticles and silver ions on the growth of terrestrial plant model *Arabidopsis thaliana*. *Journal of Environmental Science* 25 (9): 1947.
- Ratte, Hans Toni. 1999. Bioaccumulation and toxicity of silver compounds: A review. *ETC Environmental Toxicology and Chemistry* 18 (1): 89-108.
- Rizzello, Loris, and Pier Pablo Pompa. 2014. Nanosilver-based antibacterial drugs and devices: Mechanisms, methodological drawbacks, and guidelines. *Chemical Society Reviews* 43: 1501-18.
- Roh, Ji-Yeon, Sang Jun Sim, Jongheop Yi, Kwangsik Park, Kyu Hyuck Chung, Dong-Young Ryu, and Jinhee Choi. 2009. Ecotoxicity of silver nanoparticles on the soil nematode *Caenorhabditis elegans* using functional ecotoxicogenomics. *Environmental Science and Technology* 43: 3933.
- Ruparelia J. P., A. K. Chatterjee, S. P. Duttagupta, and S. Mukherji. 2008. Strain specificity in antimicrobial activity of silver and copper nanoparticles. *Acta Biomaterialia* 4 (3): 707-16.
- Sambhy, Varun, Megan M. MacBride, Blake R. Peterson, and Ayusman Sen. 2006. Silver bromide Nanoparticle/Polymer composites: dual action tunable antimicrobial materials. *Journal of the American Chemical Society* 128 (30): 9798-808.
- Schacht, V.J., Neumann, L.V., Sandhi, S.K., Chen, L., Henning, T., Klar, P.J., Theophel, K., Schnell, S.,Bunge, M., 2013. Effects of silver nanoparticles on microbial growth dynamics. *JAM Journal of Applied Microbiology*114 (1): 25-35.
- Schippers, B., A. W. Baker, and P. Bakker. 1987. Interactions of deleterious and beneficial rhizosphere microorganisms and the effect of cropping practices. *Annual Review of Phytopathology* 25: 339-58.
- Seyfferth, Angelia L., and David R. Parker. 2007. Effects of genotype and transpiration rate on the uptake and accumulation of perchlorate (ClO₄⁻) in lettuce. *Environmental Science and Technology* 41: 3361-7.
- Shane, M. W., M. E. McCully, and M. J. Canny. 2000. The vascular system of maize revisted: Implications for water transport and xylem safety. *Annals of Botany* 86: 245.

- Sharma, Virender K., Karolina M. Siskova, Radek Zboril, and Jorge L. Garda-Torresdey. 2014. Organic-coated silver nanoparticles in biological and environmental conditions: Fate, stability and toxicity. *Advances in Colloid and Interface Science* 204: 15.
- Sondi, Ivan, and Branka Salopek-Sondi. 2004. Silver nanoparticles as antimicrobial agent: A case study on *E. coli* as a model for gram-negative bacteria. *Journal of Colloid and Interface Science* 275 (1): 177-82.
- Stampoulis, Dimitrios, Saion K. Sinha, and Jason C. White. 2009. Assay-dependent phytotoxicity of nanoparticles to plants. *Environmental Science and Technology* 43 (24): 9473.
- Stebounova, Larissa V., Ethan Guio, and Vicki H. Grassian. 2011. Silver nanoparticles in simulated biological media: A study of aggregation, sedimentation, and dissolution. *Journal of Nanoparticle Research* 13: 233.
- Stone, Vicki, Bernd Nowack, Anders Baun, Nico van den Brink, Frank von der Kammer, Maria Dusinska, Richard Handy, et al. 2010. Nanomaterials for environmental studies: Classification, reference material issues, and strategies for physico-chemical characterisation. *Science of the Total Environment* 408: 1745.
- Suresh, Anil K., Dale A. Pelletier, Wei Wang, Ji-Won Moon, Baohua Gu, Ninell P. Mortensen, David P. Allison, David C. Joy, Tommy J. Phelps, and Mitchel J. Doktycz. 2010. Silver nanocrystallites: Biofabrication using *Shewanella oneidensis*, and an evaluation of their comparative toxicity on gram-negative and gram-positive bacteria. *Environmental Science and Technology* 44: 5210.
- USEPA. 2013. *Toxicity relationship analysis program*. V. 1.22. Mid-Continent Ecology Division.
- ———. Frequently asked questions: Biosolids. 2012 [cited 03/01/2014]. Available from http://water.epa.gov/polwaste/wastewater/treatment/biosolids/genqa.cfm.
- ——. 2007. Nanotechnology white paper. EPA 100/B-011/001.
- van der Wal, A., W. Norde, A. Zehnder, and J. Lyklema. 1997. Determination of the surface charge in the cell walls of gram-positive bacteria. *Colloids and Surfaces B: Biointerfaces* 9: 81-100.
- Vandermeer, J. 2010. How populations grow: The exponential and logistic equations. *Nature Education Knowledge* 3 (10): 15.

- Verma, V. C., R. N. Kharwar, and A. C. Gange. 2010. Biosynthesis of antimicrobial silver nanoparticles by the endophytic fungus *Aspergillus clavatus*. *Nanomedicine: Nanotechnology, Biology, and Medicine* 5 (1): 33-40.
- Wijnhoven S.W.P., W. I. Hagens, A. G. Oomen, S. Dekkers, A. J. A. M. Sips, W. J. G. M. Peijnenburg, J. Bisschops, D. Van De Meent, E. H. W. Heugens, M. Van Zijverden, I. Gosens, W. H. De Jong, C. A. Herberts, B. Roszek, and R.E. Geertsma. 2009. Nano-silver A review of available data and knowledge gaps in human and environmental risk assessment. *Nanotoxicology* 3 (2): 109-38.
- Yamamoto, H., M. Murata, and J. Sekiguchi. 2000. The *cit*ST two-component system regulates the expression of the mg-citrate transporter in *Bacillus subtilis*. *Molecular Microbiology* 37: 898-912.
- Yang, Xinyu, Andreas P. Gondikas, Stella M. Marinakos, Melanie Auffan, Jie Liu, Heileem Hsu-Kim, and Joel N. Meyer. 2012. Mechanism of silver nanoparticle toxicity is dependent on dissolved silver and surface coating in *Caenorhabditis elegans*. *Environmental Science and Technology* 46: 1119.
- Yin, Liyan, Yingwen Cheng, Benjamin Espinasse, Benjamin P. Colam, Melanie Auffan, Mark Wiesner, Jerome Rose, Jie Liu, and Emily S. Bernhardt. 2011.
 More than the ions: The effects of silver nanoparticles on Lolium multiflorum. Environmental Science and Technology 45: 2360.
- Yoon, K.Y., J. Hoon-Byeon, J. H. Park, and J. Hwang. 2007. Susceptibility constants of *Escherichia coli* and *Bacillus subtilis* to silver and copper nanoparticles. *Science of the Total Environment* 373 (2-3): 572-5.
- Zhang, Wen, Ying Yao, Nicole Sullivan, and Yongshen Chen. 2011. Modeling the primary size effects of citrate-coated silver nanoparticles on their ion release kinetics. *Environmental Science and Technology* 45: 4422-8.
- Zobel, Richard W. 2005. Tertiary root systems. In *Roots and soil management: Interactions between roots and the soil.*, eds. Richard W. Zobel, Sara F. Wright, 35-56. Madison, WI: American Society of Agronomy, Crop Science Society of America, Soil Science Society of America.
Appendix A

PLANT NUTRIENT SOLUTION COMPOSITION

Salt	Molecular Weight (g/mol)	Stock Solution (mol/L)	Aliquot per 1 L (mL)	Final Concentration (µM)
$Ca(NO_3)_2*4H_2O$	236.15	0.9500	2.0	1900.0
NH ₄ NO ₃	80.06	0.0500	2.0	100.0
KNO ₃	101.11	0.5000	2.0	1000.0
MgSO ₄ *7H ₂ O	246.48	0.5000	1.0	500.0
KH ₂ PO ₄	136.09	0.2400	0.333	80.0
H ₃ BO ₃	61.83	0.0100	1	10.0
Na ₂ MoO ₄ *2H ₂ O	241.95	0.0001	1	0.1
ZnCl ₂	136.28	0.0080		8.0
MnCl ₂ *4H ₂ O	197.91	0.0006		0.6
$CuCl_2*2H_2O$	170.48	0.0020		2.0
NiCl ₂ *6H ₂ O	237.71	0.0001	1	0.1
FeCl ₃ *6H ₂ O	270.30	0.0200		20.0
HEDTA	374.45	0.0577		57.7
HCl (1 M)				
MES	213.24	0.5000	2	1000.0
NaOH	40	02500	2	500.0

Table 3: Plant Nutrient Solution

Appendix B

PARTICLE SIZE ANALYSIS

Table 4: Particle Size Analysis

Particle #	Line Length (pixels)	Diameter (nm)*	Particle #	Line Length (pixels)	Diameter (nm)*
1	53.67	58.23	31	40.20	43.62
2	53.37	57.91	32	62.29	67.58
3	50.99	55.32	33	43.86	47.59
4	59.46	64.51	34	50.36	54.64
5	55.46	60.17	35	43.17	46.84
6	58.14	63.08	36	52.35	56.80
7	55.17	59.86	37	54.59	59.23
8	50.99	55.32	38	56.60	61.41
9	55.61	60.34	39	52.84	57.33
10	58.14	63.08	40	46.17	50.09
11	44.41	48.18	41	50.16	54.42
12	51.42	55.79	42	50.60	54.90
13	41.18	44.68	43	50.99	55.32
14	48.17	52.26	44	49.48	53.69
15	52.04	56.46	45	46.04	49.95
16	60.00	65.10	46	44.72	48.52
17	51.26	55.62	47	52.61	57.08
18	56.57	61.38	48	46.39	50.33
19	52.00	56.42	49	45.69	49.57
20	56.82	61.65	50	44.94	48.76
21	52.80	57.29	51	47.20	51.21
22	58.58	63.56	52	48.33	52.44
23	50.00	54.25	53	43.68	47.39
24	40.79	44.26	54	48.66	52.80
25	66.48	72.13	55	51.46	55.83

Particle #	Line Length (pixels)	Diameter (nm)*	Particle #	Line Length (pixels)	Diameter (nm)*
26	57.27	62.14	56	38.47	41.74
27	54.92	59.59	57	44.05	47.79
28	54.00	58.59	58	43.68	47.39
59	41.23	44.73	93	36.88	40.01
60	46.17	50.09	94	41.04	44.53
61	40.79	44.26	95	47.41	51.44
62	48.04	52.12	96	37.36	40.54
63	52.84	57.33	97	39.85	43.24
64	46.18	50.11	98	46.52	50.47
65	39.45	42.80	99	46.82	50.80
66	38.47	41.74	100	44.18	47.94
67	33.29	36.12	101	41.18	44.68
68	44.00	47.74	102	37.95	41.18
69	46.86	50.84	103	35.61	38.64
70	44.72	48.52	104	34.99	37.96
71	43.27	46.95	105	36.06	39.13
72	30.07	32.63	106	40.20	43.62
73	46.69	50.66	107	41.76	45.31
74	44.05	47.79	108	43.27	46.95
75	45.61	49.49	109	35.61	38.64
76	46.04	49.95	110	58.14	63.08
77	34.06	36.96	111	34.05	36.94
78	42.05	45.62	112	32.56	35.33
79	39.29	42.63	113	33.29	36.12
80	38.47	41.74	114	38.83	42.13
81	34.53	37.47	115	26.83	29.11
82	46.04	49.95	116	46.86	50.84
83	37.74	40.95	117	41.23	44.73
84	45.34	49.19	118	37.95	41.18
85	48.37	52.48	119	29.73	32.26
86	40.00	43.40	120	32.98	35.78
87	44.41	48.18	117	41.23	44.73
88	39.40	42.75	118	37.95	41.18
89	36.72	39.84	119	29.73	32.26
90	34.93	37.90	120	32.98	35.78

Particle #	Line Length (pixels)	Diameter (nm)*	Particle #	Line Length (pixels)	Diameter (nm)*
91	41.23	44.73	121	37.36	40.54
92	40.20	43.62	122	42.76	46.39
123	35.78	38.82	158	43.86	47.59
124	39.70	43.07	159	30.59	33.19
125	34.93	37.90	160	32.80	35.59
126	54.59	59.23	161	43.08	46.74
127	44.00	47.74	162	42.52	46.13
128	36.06	39.13	163	37.95	41.18
129	45.69	49.57	164	34.06	36.96
130	35.38	38.39	165	32.56	35.33
131	42.05	45.62	166	32.31	35.06
132	35.44	38.45	167	30.27	32.84
133	36.72	39.84	168	40.25	43.67
134	34.00	36.89	169	36.88	40.01
135	46.04	49.95	170	42.05	45.62
136	41.23	44.73	171	38.47	41.74
137	44.72	48.52	172	37.76	40.97
138	36.88	40.01	173	32.00	34.72
139	43.17	46.84	174	34.06	36.96
140	41.76	45.31	175	49.40	53.60
141	34.00	36.89	176	38.42	41.69
142	37.20	40.36	177	38.47	41.74
143	47.41	51.44	178	36.77	39.90
144	46.04	49.95	179	37.95	41.18
145	51.42	55.79	180	30.46	33.05
146	40.50	43.94	181	41.62	45.16
147	37.74	40.95	182	36.88	40.01
148	35.44	38.45	183	34.99	37.96
149	36.72	39.84	184	42.52	46.13
150	38.83	42.13	185	38.63	41.91
151	42.43	46.04	186	38.42	41.69
152	39.60	42.97	187	41.23	44.73
153	38.47	41.74	188	38.05	41.28
154	46.17	50.09	189	43.27	46.95
155	34.93	37.90	190	44.41	48.18

Particle	Line Length	Diameter	Particle	Line Length	Diameter
#	(pixels)	(nm)*	#	(pixels)	(nm)*
156	27.20	29.51	191	39.45	42.80
157	36.88	40.01	192	50.00	54.25
193	41.23	44.73	229	37.58	40.77
194	42.19	45.78	230	33.11	35.92
195	42.38	45.98	231	33.29	36.12
196	36.88	40.01	232	35.44	38.45
197	34.53	37.47	233	40.79	44.26
198	38.05	41.28	234	35.44	38.45
199	49.48	53.69	235	36.77	39.90
200	44.27	48.03	236	39.45	42.80
201	37.58	40.77	237	41.04	44.53
202	37.74	40.95	238	42.94	46.59
203	30.53	33.13	239	36.77	39.90
204	33.29	36.12	240	46.00	49.91
205	43.86	47.59	241	45.25	49.10
206	39.60	42.97	242	37.20	40.36
207	33.29	36.12	243	35.38	38.39
208	39.85	43.24	244	36.72	39.84
209	36.06	39.13	245	31.30	33.96
210	36.22	39.30	246	44.41	48.18
211	26.68	28.95	247	34.00	36.89
212	45.69	49.57	248	36.77	39.90
213	30.59	33.19	249	42.76	46.39
214	36.50	39.60	250	38.83	42.13
215	39.40	42.75	251	47.71	51.77
216	34.23	37.14	252	28.28	30.68
217	40.79	44.26	253	34.93	37.90
218	50.16	54.42	254	29.73	32.26
219	41.23	44.73	255	39.45	42.80
220	34.99	37.96	256	43.27	46.95
221	42.52	46.13	257	36.22	39.30
222	33.53	36.38	258	37.74	40.95
223	39.60	42.97	259	44.72	48.52
224	43.86	47.59	260	46.39	50.33
225	29.12	31.60	261	40.35	43.78
226	41.76	45.31	262	46.00	49.91

Particle	Line Length	Diameter	Particle Line Length Di		Diameter
#	(pixels)	(nm)*	#	(pixels)	(nm)*
227	22.80	24.74	263	47.54	51.58
228	31.24	33.90	264	38.00	41.23
265	45.61	49.49	301	35.38	38.39
266	43.27	46.95	302	43.17	46.84
267	37.20	40.36	303	46.17	50.09
268	38.42	41.69	304	38.63	41.91
269	29.12	31.60	305	33.11	35.92
270	37.74	40.95	306	42.19	45.78
271	39.40	42.75	307	42.05	45.62
272	38.05	41.28	308	46.00	49.91
273	39.70	43.07	309	41.04	44.53
274	42.43	46.04	310	49.40	53.60
275	46.00	49.91	311	37.58	40.77
276	50.60	54.90	312	41.18	44.68
277	36.22	39.30	313	40.25	43.67
278	36.06	39.13	314	40.50	43.94
279	32.56	35.33	315	36.50	39.60
280	41.23	44.73	316	44.18	47.94
281	45.34	49.19	317	40.45	43.89
282	48.70	52.84	318	38.21	41.46
283	40.50	43.94	319	40.20	43.62
284	41.62	45.16	320	40.05	43.45
285	39.53	42.89	321	44.72	48.52
286	26.08	28.30	322	36.06	39.13
287	40.79	44.26	323	29.73	32.26
288	36.77	39.90	324	31.24	33.90
289	35.38	38.39	325	41.23	44.73
290	39.29	42.63	326	38.83	42.13
291	34.06	36.96	327	43.27	46.95
292	31.30	33.96	328	38.83	42.13
293	37.36	40.54	329	43.86	47.59
294	48.08	52.17	330	43.17	46.84
295	31.24	33.90	331	35.44	38.45
296	51.92	56.33	332	49.03	53.20
297	37.95	41.18	333	37.95	41.18
298	44.94	48.76	334	36.88	40.01
			-		

Particle	Line Length	Diameter	Particle Line Length		Diameter
#	(pixels)	(nm)*	#	(pixels)	(nm)*
299	25.61	27.79	335	35.61	38.64
300	39.29	42.63	336	38.42	41.69
337	32.06	34.79	373	48.66	52.80
338	50.12	54.38	374	34.23	37.14
339	53.85	58.43	375	38.05	41.28
340	47.07	51.07	376	50.36	54.64
341	36.77	39.90	377	36.88	40.01
342	41.04	44.53	378	36.06	39.13
343	34.23	37.14	379	48.66	52.80
344	39.45	42.80	380	43.08	46.74
345	39.45	42.80	381	49.68	53.90
346	50.60	54.90	382	47.41	51.44
347	40.20	43.62	383	52.15	56.58
348	38.63	41.91	384	42.38	45.98
349	41.23	44.73	385	44.05	47.79
350	52.13	56.56	386	35.61	38.64
351	41.76	45.31	387	36.05	39.11
352	41.18	44.68	388	43.08	46.74
353	50.04	54.29	389	43.27	46.95
354	32.31	35.06	390	42.94	46.59
355	44.05	47.79	391	39.40	42.75
356	38.42	41.69	392	54.59	59.23
357	41.23	44.73	393	48.83	52.98
358	53.25	57.78	394	53.37	57.91
359	33.53	36.38	395	38.83	42.13
360	50.99	55.32	396	38.47	41.74
361	36.77	39.90	397	40.79	44.26
362	46.17	50.09	398	37.95	41.18
363	53.85	58.43	399	53.85	58.43
364	53.25	57.78	400	44.05	47.79
365	40.50	43.94	401	42.38	45.98
366	36.77	39.90	402	50.64	54.94
367	34.00	36.89	403	41.04	44.53
368	47.41	51.44	404	43.36	47.05
369	48.33	52.44	405	50.00	54.25
370	49.40	53.60	406	37.36	40.54

Particle	Line Length	Diameter	Particle	Line Length	Diameter
#	(pixels)	(nm)*	#	(pixels)	(nm)*
371	52.35	56.80	407	45.69	49.57
372	50.99	55.32	408	48.33	52.44
409	54.04	58.63	445	48.08	52.17
410	41.23	44.73	446	46.69	50.66
411	46.82	50.80	447	42.52	46.13
412	37.58	40.77	448	48.00	52.08
413	42.43	46.04	449	38.83	42.13
414	46.69	50.66	450	53.81	58.38
415	49.52	53.73	451	36.77	39.90
416	36.88	40.01	452	42.19	45.78
417	48.00	52.08	453	38.05	41.28
418	36.72	39.84	454	44.72	48.52
419	48.08	52.17	455	48.66	52.80
420	41.04	44.53	456	40.45	43.89
421	40.50	43.94	457	56.60	61.41
422	38.00	41.23	458	40.50	43.94
423	42.05	45.62	459	43.86	47.59
424	41.23	44.73	460	40.25	43.67
425	44.05	47.79	461	44.94	48.76
426	41.62	45.16	462	34.41	37.33
427	32.80	35.59	463	38.83	42.13
428	47.54	51.58	464	44.18	47.94
429	40.00	43.40	465	40.79	44.26
430	53.25	57.78	466	38.00	41.23
431	52.35	56.80	467	32.80	35.59
432	45.65	49.53	468	42.19	45.78
433	51.61	56.00	469	36.06	39.13
434	42.19	45.78	470	37.58	40.77
435	42.76	46.39	471	32.25	34.99
436	48.04	52.12	472	56.04	60.80
437	38.42	41.69	473	34.06	36.96
438	50.99	55.32	474	37.76	40.97
439	38.42	41.69	475	37.20	40.36
440	39.29	42.63	476	44.27	48.03
441	43.91	47.64	477	54.00	58.59
442	36.50	39.60	478	43.68	47.39

Particle	Line Length	Diameter	Particle	Line Length	Diameter
#	(pixels)	(nm)*	#	(pixels)	(nm)*
443	36.72	39.84	479	41.04	44.53
444	37.58	40.77	480	47.71	51.77
481	40.25	43.67	517	42.05	45.62
482	36.06	39.13	518	38.47	41.74
483	34.93	37.90	519	34.53	37.47
484	46.69	50.66	520	40.00	43.40
485	50.12	54.38	521	48.17	52.26
486	39.45	42.80	522	34.53	37.47
487	51.25	55.61	523	48.37	52.48
488	39.45	42.80	524	44.72	48.52
489	36.72	39.84	525	44.05	47.79
490	40.79	44.26	526	41.23	44.73
491	49.19	53.37	527	45.65	49.53
492	43.91	47.64	528	43.27	46.95
493	41.04	44.53	529	36.50	39.60
494	44.41	48.18	530	45.61	49.49
495	40.25	43.67	531	42.00	45.57
496	39.85	43.24	532	47.54	51.58
497	39.45	42.80	533	46.86	50.84
498	40.45	43.89	534	46.39	50.33
499	32.06	34.79	535	35.38	38.39
500	41.62	45.16	536	46.69	50.66
501	43.91	47.64	537	48.08	52.17
502	43.08	46.74	538	44.05	47.79
503	39.40	42.75	539	40.45	43.89
504	38.63	41.91	540	50.16	54.42
505	45.65	49.53	541	36.00	39.06
506	42.43	46.04	542	45.25	49.10
507	42.83	46.47	543	42.19	45.78
508	48.41	52.52	544	46.04	49.95
509	41.23	44.73	545	40.00	43.40
510	37.36	40.54	546	54.00	58.59
511	39.70	43.07	547	45.61	49.49
512	40.45	43.89	548	36.88	40.01
513	42.05	45.62	549	48.17	52.26
514	52.50	56.96	550	43.27	46.95

Particle Line Length	Diameter	Particle Line Length Dian		Diameter
# (pixels)	(nm)*	#	(pixels)	(nm)*
515 43.17	46.84	551	34.99	37.96
516 34.99	37.96	552	45.34	49.19
553 40.79	44.26	583	49.40	53.60
554 46.04	49.95	584	46.04	49.95
555 38.00	41.23	585	27.86	30.23
556 36.06	39.13	586	39.60	42.97
557 40.00	43.40	587	32.98	35.78
558 30.00	32.55	588	32.25	34.99
559 41.23	44.73	589	36.06	39.13
560 48.37	52.48	590	46.69	50.66
561 42.05	45.62	591	40.20	43.62
562 43.17	46.84	592	34.00	36.89
563 28.68	31.12	593	37.36	40.54
564 34.00	36.89	594	44.94	48.76
565 31.62	34.31	595	31.62	34.31
566 36.06	39.13	596	52.04	56.46
567 36.72	39.84	597	44.05	47.79
568 38.21	41.46	598	45.69	49.57
569 51.22	55.57	599	36.00	39.06
570 52.80	57.29	600	43.68	47.39
571 42.93	46.58	601	49.52	53.73
572 35.78	38.82	602	50.99	55.32
573 44.94	48.76	603	44.05	47.79
574 35.61	38.64	604	42.05	45.62
575 30.53	33.13	605	45.69	49.57
576 30.00	32.55	606	46.17	50.09
577 43.71	47.43	607	47.54	51.58
578 36.88	40.01			
579 32.56	35.33			
580 39.45	42.80			
581 45.61	49.49			
582 49.68	53.90			

* Diameter = Pixel Length x 1.085

Appendix C

STATISTICAL TABLES

Table 5: ANOVA Results for Inhibition of *B. subtilis* and *E. coli* by c-AgNPs

ANOVA									
Absorbance									
	Sum of Squares	df	Mean Square	F	Sig.				
Between Groups	16.221	11	1.475	872.771	.000				
Within Groups	.488	289	.002						
Total	16.709	300							

Multiple Comparisons									
Dependent Va	riable: Absorba	ince							
Tukey HSD									
		Mean Difference (I-			95% Confide	ence Interval			
(I) Treatment	(J) Treatment	J)	Std. Error	Sig.	Lower Bound	Upper Bound			
1	2	.001849	.011922	1.000	03744	.04113			
	3	.022471	.011522	.726	01549	.06044			
	4	.043143 [*]	.011346	.009	.00576	.08053			
	5	041000*	.011346	.018	07839	00361			
	6	.687571 [*]	.011346	.000	.65019	.72496			
	7	.092591*	.009022	.000	.06286	.12232			
	8	.119571 [*]	.011033	.000	.08322	.15593			
	9	.117048 [*]	.011346	.000	.07966	.15443			
	10	.124311 [*]	.011033	.000	.08795	.16067			
	11	.046446*	.010894	.002	.01055	.08234			
	12	.707702 [*]	.011033	.000	.67135	.74406			
2	1	001849	.011922	1.000	04113	.03744			
	3	.020622	.013355	.927	02338	.06463			
	4	.041294	.013203	.081	00221	.08480			
	5	042849	.013203	.058	08636	.00066			
	6	.685722*	.013203	.000	.64222	.72923			
	7	.090742*	.011269	.000	.05361	.12788			
	8	.117722 [*]	.012935	.000	.07510	.16035			
	9	.115198 [*]	.013203	.000	.07169	.15870			
	10	.122461 [*]	.012935	.000	.07984	.16509			
	11	.044597 [*]	.012817	.028	.00236	.08683			
	12	.705853 [*]	.012935	.000	.66323	.74848			
3	1	022471	.011522	.726	06044	.01549			
	2	020622	.013355	.927	06463	.02338			
	4	.020671	.012843	.904	02165	.06299			
	5	063471 [*]	.012843	.000	10579	02115			
	6	.665100 [*]	.012843	.000	.62278	.70742			

Table 6: Tukey's Post-Hoc Results for Inhibition of *B. subtilis* and *E. coli* by c-AgNPs

	7	.070120 [*]	.010845	.000	.03438	.10585
	8	.097100 [*]	.012567	.000	.05569	.13851
	9	.094576 [*]	.012843	.000	.05226	.13689
	10	.101839 [*]	.012567	.000	.06043	.14325
	11	.023975	.012445	.742	01703	.06498
	12	.685230 [*]	.012567	.000	.64382	.72664
4	1	043143 [*]	.011346	.009	08053	00576
	2	041294	.013203	.081	08480	.00221
	3	020671	.012843	.904	06299	.02165
	5	084143 [*]	.012685	.000	12594	04234
	6	.644429 [*]	.012685	.000	.60263	.68623
	7	.049448 [*]	.010658	.000	.01433	.08457
	8	.076429 [*]	.012406	.000	.03555	.11731
	9	.073905 [*]	.012685	.000	.03211	.11570
	10	.081168 [*]	.012406	.000	.04029	.12205
	11	.003304	.012282	1.000	03717	.04378
	12	.664559 [*]	.012406	.000	.62368	.70544
5	1	.041000 [*]	.011346	.018	.00361	.07839
	2	.042849	.013203	.058	00066	.08636
	3	.063471 [*]	.012843	.000	.02115	.10579
	4	.084143 [*]	.012685	.000	.04234	.12594
	6	.728571 [*]	.012685	.000	.68677	.77037
	7	.133591 [*]	.010658	.000	.09847	.16871
	8	.160571 [*]	.012406	.000	.11969	.20145
	9	.158048 [*]	.012685	.000	.11625	.19985
	10	.165311 [*]	.012406	.000	.12443	.20619
	11	.087446 [*]	.012282	.000	.04697	.12792
	12	.748702 [*]	.012406	.000	.70782	.78958
6	1	687571 [*]	.011346	.000	72496	65019
	2	685722 [*]	.013203	.000	72923	64222
	3	665100 [*]	.012843	.000	70742	62278
	4	644429 [*]	.012685	.000	68623	60263
	5	728571 [*]	.012685	.000	77037	68677
	7	594980 [*]	.010658	.000	63010	55986
	8	568000 [*]	.012406	.000	60888	52712

	9	570524 [*]	.012685	.000	61232	52872
	10	563261 [*]	.012406	.000	60414	52238
	11	641125 [*]	.012282	.000	68160	60065
	12	.020130	.012406	.900	02075	.06101
7	1	092591 [*]	.009022	.000	12232	06286
	2	090742 [*]	.011269	.000	12788	05361
	3	070120 [*]	.010845	.000	10585	03438
	4	049448 [*]	.010658	.000	08457	01433
	5	133591 [*]	.010658	.000	16871	09847
	6	.594980 [*]	.010658	.000	.55986	.63010
	8	.026980	.010324	.277	00704	.06100
	9	.024457	.010658	.483	01066	.05957
	10	.031720	.010324	.094	00230	.06574
	11	046145 [*]	.010175	.001	07967	01262
	12	.615111 [*]	.010324	.000	.58109	.64913
8	1	119571 [*]	.011033	.000	15593	08322
	2	117722 [*]	.012935	.000	16035	07510
	3	097100 [*]	.012567	.000	13851	05569
	4	076429 [*]	.012406	.000	11731	03555
	5	160571 [*]	.012406	.000	20145	11969
	6	.568000*	.012406	.000	.52712	.60888
	7	026980	.010324	.277	06100	.00704
	9	002524	.012406	1.000	04340	.03836
	10	.004739	.012121	1.000	03520	.04468
	11	073125 [*]	.011994	.000	11265	03360
	12	.588130 [*]	.012121	.000	.54819	.62807
9	1	117048 [*]	.011346	.000	15443	07966
	2	115198 [*]	.013203	.000	15870	07169
	3	094576 [*]	.012843	.000	13689	05226
	4	073905 [*]	.012685	.000	11570	03211
	5	158048 [*]	.012685	.000	19985	11625
	6	.570524 [*]	.012685	.000	.52872	.61232
	7	024457	.010658	.483	05957	.01066
	8	.002524	.012406	1.000	03836	.04340
	10	.007263	.012406	1.000	03362	.04814

	11	070601 [*]	.012282	.000	11107	03013
	12	.590654 [*]	.012406	.000	.54977	.63153
10	1	124311 [*]	.011033	.000	16067	08795
	2	122461 [*]	.012935	.000	16509	07984
	3	101839 [*]	.012567	.000	14325	06043
	4	081168 [*]	.012406	.000	12205	04029
	5	165311 [*]	.012406	.000	20619	12443
	6	.563261*	.012406	.000	.52238	.60414
	7	031720	.010324	.094	06574	.00230
	8	004739	.012121	1.000	04468	.03520
	9	007263	.012406	1.000	04814	.03362
	11	077864 [*]	.011994	.000	11739	03834
	12	.583391 [*]	.012121	.000	.54345	.62333
11	1	046446*	.010894	.002	08234	01055
	2	044597 [*]	.012817	.028	08683	00236
	3	023975	.012445	.742	06498	.01703
	4	003304	.012282	1.000	04378	.03717
	5	087446 [*]	.012282	.000	12792	04697
	6	.641125 [*]	.012282	.000	.60065	.68160
	7	.046145 [*]	.010175	.001	.01262	.07967
	8	.073125 [*]	.011994	.000	.03360	.11265
	9	.070601 [*]	.012282	.000	.03013	.11107
	10	.077864 [*]	.011994	.000	.03834	.11739
	12	.661255 [*]	.011994	.000	.62173	.70078
12	1	707702 [*]	.011033	.000	74406	67135
	2	705853 [*]	.012935	.000	74848	66323
	3	685230 [*]	.012567	.000	72664	64382
	4	664559 [*]	.012406	.000	70544	62368
	5	748702 [*]	.012406	.000	78958	70782
	6	020130	.012406	.900	06101	.02075
	7	615111 [*]	.010324	.000	64913	58109
	8	588130 [*]	.012121	.000	62807	54819
	9	590654 [*]	.012406	.000	63153	54977
	10	583391 [*]	.012121	.000	62333	54345
	11	661255 [*]	.011994	.000	70078	62173

*. The mean difference is significant at the 0.05 level.

1: *B. subtilis* Control; 2: *B. subtilis* 0.1 mg/L c-AgNP; 3: *B. subtilis* 1.0 mg/L c-AgNP; 4: *B. subtilis* 5.0 mg/L c-AgNP; 5: *B. subtilis* 0.1 mg/L Ag-Ion; 6: *B. subtilis* 1.0 mg/L Ag-Ion; 7: *E. coli* Control; 8: *E. coli* 0.1 mg/L c-AgNP; 9: *E. coli* 1.0 mg/L c-AgNP; 10: *E. coli* 5.0 mg/L c-AgNP; 11: *E. coli* 0.1 mg/L Ag-Ion; 12: *E. coli* 1.0 mg/L Ag-Ion

	ANOVA										
		Sum of Squares	df	Mean Square	F	Sig.					
к	Between Groups	1.126	9	.125	60.124	.000					
	Within Groups	.529	254	.002							
	Total	1.655	263								
N0	Between Groups	.039	9	.004	46.977	.000					
	Within Groups	.024	254	.000							
	Total	.063	263								
r	Between Groups	3.540	9	.393	10.716	.000					
	Within Groups	9.324	254	.037							
	Total	12.864	263								

Table 7: ANOVA Results for Growth Kinetics of *B. subtilis* and *E. coli* exposed to c-AgNPs

Multiple Comparisons									
Tukey HSD									
						95% Con	fidence		
			Mean			Inter	val		
Dependent	(I)	(J)	Difference	Std.		Lower	Upper		
Variable	Treatment	Treatment	(I-J)	Error	Sig.	Bound	Bound		
К	1	2	004903	.01229	1.000	04415	.03434		
		3	.011376	.01249	.996	02851	.05126		
		4	.042542*	.01229	.022	.00330	.08179		
		5	047240 [*]	.01229	.006	08648	00800		
		6	.110478 [*]	.00967	.000	.07959	.14137		
		7	.140975 [*]	.01193	.000	.10287	.17908		
		8	.138683*	.01229	.000	.09944	.17793		
		9	.151573 [*]	.01193	.000	.11347	.18968		
		10	.061941 [*]	.01177	.000	.02434	.09954		
	2	1	.004903	.01229	1.000	03434	.04415		
		3	.016279	.01425	.980	02922	.06178		
		4	.047446 [*]	.01407	.029	.00250	.09239		
		5	042336	.01407	.084	08728	.00260		
		6	.115381 [*]	.01186	.000	.07751	.15325		
		7	.145879 [*]	.01376	.000	.10193	.18983		
		8	.143587 [*]	.01407	.000	.09865	.18853		
		9	.156476 [*]	.01376	.000	.11252	.20043		
		10	.066845 [*]	.01363	.000	.02333	.11036		
	3	1	011376	.01249	.996	05126	.02851		
		2	016279	.01425	.980	06178	.02922		
		4	.031167	.01425	.469	01433	.07667		
		5	058615 [*]	.01425	.002	10411	01312		
		6	.099102*	.01207	.000	.06057	.13763		
		7	.129599*	.01394	.000	.08508	.17412		

Table 8: Tukey's Post-Hoc Results for Growth Kinetics of B. subtilis and E. coliExposed to c-AgNPs

	8	.127308 [*]	.01425	.000	.08181	.17281
	9	.140197 [*]	.01394	.000	.09567	.18472
	10	.050566*	.01381	.011	.00648	.09466
4	1	042542 [*]	.01229	.022	08179	00330
	2	047446 [*]	.01407	.029	09239	00250
	3	031167	.01425	.469	07667	.01433
	5	089782 [*]	.01407	.000	13472	04484
	6	.067935 [*]	.01186	.000	.03007	.10580
	7	.098433*	.01376	.000	.05448	.14239
	8	.096141 [*]	.01407	.000	.05120	.14108
	9	.109030 [*]	.01376	.000	.06508	.15298
	10	.019399	.01363	.919	02411	.06291
5	1	.047240 [*]	.01229	.006	.00800.	.08648
	2	.042336	.01407	.084	00260	.08728
	3	.058615 [*]	.01425	.002	.01312	.10411
	4	.089782 [*]	.01407	.000	.04484	.13472
	6	.157717 [*]	.01186	.000	.11985	.19559
	7	.188215 [*]	.01376	.000	.14426	.23217
	8	.185923 [*]	.01407	.000	.14098	.23086
	9	.198812 [*]	.01376	.000	.15486	.24277
	10	.109181 [*]	.01363	.000	.06567	.15269
6	1	110478 [*]	.00967	.000	14137	07959
	2	115381 [*]	.01186	.000	15325	07751
	3	099102 [*]	.01207	.000	13763	06057
	4	067935 [*]	.01186	.000	10580	03007
	5	157717 [*]	.01186	.000	19559	11985
	7	.030497	.01149	.199	00619	.06719
	8	.028205	.01186	.344	00966	.06607
	9	.041095 [*]	.01149	.015	.00440	.07778
	10	048537 [*]	.01132	.001	08470	01237
7	1	140975 [*]	.01193	.000	17908	10287
	2	145879 [*]	.01376	.000	18983	10193
	3	129599 [*]	.01394	.000	17412	08508
	4	098433 [*]	.01376	.000	14239	05448
	5	188215 [*]	.01376	.000	23217	14426

	6	030497	.01149	.199	06719	.00619
	8	002292	.01376	1.000	04624	.04166
	9	.010598	.01345	.999	03234	.05354
	10	079034 [*]	.01331	.000	12153	03654
8	1	138683 [*]	.01229	.000	17793	09944
	2	143587 [*]	.01407	.000	18853	09865
	3	127308 [*]	.01425	.000	17281	08181
	4	096141 [*]	.01407	.000	14108	05120
	5	185923 [*]	.01407	.000	23086	14098
	6	028205	.01186	.344	06607	.00966
	7	.002292	.01376	1.000	04166	.04624
	9	.012889	.01376	.995	03106	.05684
	10	076742 [*]	.01363	.000	12026	03323
9	1	151573 [*]	.01193	.000	18968	11347
	2	156476 [*]	.01376	.000	20043	11252
	3	140197 [*]	.01394	.000	18472	09567
	4	109030 [*]	.01376	.000	15298	06508
	5	198812 [*]	.01376	.000	24277	15486
	6	041095 [*]	.01149	.015	07778	00440
	7	010598	.01345	.999	05354	.03234
	8	012889	.01376	.995	05684	.03106
	10	089631 [*]	.01331	.000	13212	04714
10	1	061941 [*]	.01177	.000	09954	02434
	2	066845*	.01363	.000	11036	02333
	3	050566*	.01381	.011	09466	00648
	4	019399	.01363	.919	06291	.02411
	5	109181 [*]	.01363	.000	15269	06567
	6	.048537*	.01132	.001	.01237	.08470
	7	.079034 [*]	.01331	.000	.03654	.12153
	8	.076742*	.01363	.000	.03323	.12026
	9	.089631 [*]	.01331	.000	.04714	.13212
1	2	.000608	.00259	1.000	00767	.00889
	3	.002588	.00263	.993	00583	.01100
	4	.009033*	.00259	.020	.00075	.01731
	5	002424	.00259	.995	01070	.00586

N0

	6	023467*	.00204	.000	02999	01695
	7	023229*	.00251	.000	03127	01519
	8	024073 [*]	.00259	.000	03235	01579
	9	015693 [*]	.00251	.000	02373	00765
	10	021299*	.00248	.000	02923	01336
2	1	000608	.00259	1.000	00889	.00767
	3	.001980	.00300	1.000	00762	.01158
	4	.008425	.00297	.130	00106	.01791
	5	003032	.00297	.991	01252	.00645
	6	024076 [*]	.00250	.000	03207	01608
	7	023837*	.00290	.000	03311	01456
	8	024681 [*]	.00297	.000	03416	01520
	9	016301 [*]	.00290	.000	02558	00703
	10	021907 [*]	.00287	.000	03109	01272
3	1	002588	.00263	.993	01100	.00583
	2	001980	.00300	1.000	01158	.00762
	4	.006445	.00300	.499	00316	.01605
	5	005011	.00300	.813	01461	.00459
	6	026055*	.00254	.000	03419	01792
	7	025816 [*]	.00294	.000	03521	01642
	8	026660*	.00300	.000	03626	01706
	9	018281 [*]	.00294	.000	02768	00889
	10	023887*	.00291	.000	03319	01458
4	1	009033*	.00259	.020	01731	00075
	2	008425	.00297	.130	01791	.00106
	3	006445	.00300	.499	01605	.00316
	5	011457 [*]	.00297	.006	02094	00197
	6	032501 [*]	.00250	.000	04049	02451
	7	032262*	.00290	.000	04154	02299
	8	033106 [*]	.00297	.000	04259	02362
	9	024726 [*]	.00290	.000	03400	01545
	10	030332 [*]	.00287	.000	03951	02115
5	1	.002424	.00259	.995	00586	.01070
	2	.003032	.00297	.991	00645	.01252
	3	.005011	.00300	.813	00459	.01461

	4	.011457 [*]	.00297	.006	.00197	.02094
	6	021044 [*]	.00250	.000	02903	01305
	7	020805 [*]	.00290	.000	03008	01153
	8	021649 [*]	.00297	.000	03113	01217
	9	013269 [*]	.00290	.000	02254	00399
	10	018875 [*]	.00287	.000	02806	00969
6	1	.023467 [*]	.00204	.000	.01695	.02999
	2	.024076 [*]	.00250	.000	.01608	.03207
	3	.026055*	.00254	.000	.01792	.03419
	4	.032501 [*]	.00250	.000	.02451	.04049
	5	.021044 [*]	.00250	.000	.01305	.02903
	7	.000239	.00242	1.000	00750	.00798
	8	000605	.00250	1.000	00860	.00739
	9	.007774 [*]	.00242	.048	.00003	.01552
	10	.002168	.00239	.996	00546	.00980
7	1	.023229 [*]	.00251	.000	.01519	.03127
	2	.023837 [*]	.00290	.000	.01456	.03311
	3	.025816 [*]	.00294	.000	.01642	.03521
	4	.032262 [*]	.00290	.000	.02299	.04154
	5	.020805 [*]	.00290	.000	.01153	.03008
	6	000239	.00242	1.000	00798	.00750
	8	000844	.00290	1.000	01012	.00843
	9	.007536	.00283	.198	00153	.01660
	10	.001930	.00280	1.000	00704	.01090
8	1	.024073 [*]	.00259	.000	.01579	.03235
	2	.024681 [*]	.00297	.000	.01520	.03416
	3	.026660*	.00300	.000	.01706	.03626
	4	.033106 [*]	.00297	.000	.02362	.04259
	5	.021649 [*]	.00297	.000	.01217	.03113
	6	.000605	.00250	1.000	00739	.00860
	7	.000844	.00290	1.000	00843	.01012
	9	.008380	.00290	.116	00090	.01765
	10	.002774	.00287	.994	00641	.01196
9	1	.015693 [*]	.00251	.000	.00765	.02373
	2	.016301 [*]	.00290	.000	.00703	.02558

		3	.018281 [*]	.00294	.000	.00889	.02768
		4	.024726 [*]	.00290	.000	.01545	.03400
		5	.013269 [*]	.00290	.000	.00399	.02254
		6	007774*	.00242	.048	01552	00003
		7	007536	.00283	.198	01660	.00153
		8	008380	.00290	.116	01765	.00090
		10	005606	.00280	.603	01457	.00336
	10	1	.021299*	.00248	.000	.01336	.02923
		2	.021907 [*]	.00287	.000	.01272	.03109
		3	.023887 [*]	.00291	.000	.01458	.03319
		4	.030332*	.00287	.000	.02115	.03951
		5	.018875 [*]	.00287	.000	.00969	.02806
		6	002168	.00239	.996	00980	.00546
		7	001930	.00280	1.000	01090	.00704
		8	002774	.00287	.994	01196	.00641
		9	.005606	.00280	.603	00336	.01457
r	1	2	.032838	.05163	1.000	13197	.19765
		3	011554	.05247	1.000	17905	.15594
		4	140182	.05163	.173	30499	.02463
		5	.200504 [*]	.05163	.005	.03569	.36532
		6	.174573 [*]	.04064	.001	.04483	.30431
		7	.182917 [*]	.05013	.012	.02287	.34296
		8	.214677 [*]	.05163	.002	.04986	.37949
		9	.124442	.05013	.283	03560	.28449
		10	.266147 [*]	.04947	.000	.10823	.42406
	2	1	032838	.05163	1.000	19765	.13197
		3	044392	.05986	.999	23548	.14669
		4	173020	.05912	.104	36176	.01572
		5	.167667	.05912	.130	02108	.35641
		6	.141735	.04982	.127	01730	.30077
		7	.150079	.05782	.225	03451	.33467
		8	.181839	.05912	.070	00690	.37058
		9	.091604	.05782	.855	09299	.27620
		10	.233309 [*]	.05725	.002	.05056	.41606
	3	1	.011554	.05247	1.000	15594	.17905

	2	.044392	.05986	.999	14669	.23548
	4	128627	.05986	.495	31971	.06246
	5	.212059 [*]	.05986	.017	.02097	.40315
	6	.186128 [*]	.05069	.011	.02431	.34794
	7	.194471 [*]	.05857	.034	.00748	.38146
	8	.226231 [*]	.05986	.007	.03514	.41732
	9	.135996	.05857	.379	05099	.32299
	10	.277701 [*]	.05800	.000	.09253	.46287
4	1	.140182	.05163	.173	02463	.30499
	2	.173020	.05912	.104	01572	.36176
	3	.128627	.05986	.495	06246	.31971
	5	.340686*	.05912	.000	.15194	.52943
	6	.314755 [*]	.04982	.000	.15572	.47379
	7	.323098*	.05782	.000	.13850	.50769
	8	.354858*	.05912	.000	.16612	.54360
	9	.264623 [*]	.05782	.000	.08003	.44922
	10	.406329*	.05725	.000	.22358	.58908
5	1	200504 [*]	.05163	.005	36532	03569
	2	167667	.05912	.130	35641	.02108
	3	212059 [*]	.05986	.017	40315	02097
	4	340686*	.05912	.000	52943	15194
	6	025931	.04982	1.000	18497	.13311
	7	017588	.05782	1.000	20218	.16701
	8	.014172	.05912	1.000	17457	.20291
	9	076063	.05782	.949	26066	.10853
	10	.065643	.05725	.979	11711	.24839
6	1	174573 [*]	.04064	.001	30431	04483
	2	141735	.04982	.127	30077	.01730
	3	186128 [*]	.05069	.011	34794	02431
	4	314755 [*]	.04982	.000	47379	15572
	5	.025931	.04982	1.000	13311	.18497
	7	.008343	.04827	1.000	14575	.16243
	8	.040104	.04982	.998	11893	.19914
	9	050131	.04827	.990	20422	.10396
	10	.091574	.04757	.652	06030	.24345

7	1	182917 [*]	.05013	.012	34296	02287
	2	150079	.05782	.225	33467	.03451
	3	194471 [*]	.05857	.034	38146	00748
	4	323098 [*]	.05782	.000	50769	13850
	5	.017588	.05782	1.000	16701	.20218
	6	008343	.04827	1.000	16243	.14575
	8	.031760	.05782	1.000	15283	.21635
	9	058475	.05649	.990	23882	.12187
	10	.083230	.05590	.896	09523	.26169
8	1	214677 [*]	.05163	.002	37949	04986
	2	181839	.05912	.070	37058	.00690
	3	226231 [*]	.05986	.007	41732	03514
	4	354858 [*]	.05912	.000	54360	16612
	5	014172	.05912	1.000	20291	.17457
	6	040104	.04982	.998	19914	.11893
	7	031760	.05782	1.000	21635	.15283
	9	090235	.05782	.866	27483	.09436
	10	.051470	.05725	.996	13128	.23422
9	1	124442	.05013	.283	28449	.03560
	2	091604	.05782	.855	27620	.09299
	3	135996	.05857	.379	32299	.05099
	4	264623 [*]	.05782	.000	44922	08003
	5	.076063	.05782	.949	10853	.26066
	6	.050131	.04827	.990	10396	.20422
	7	.058475	.05649	.990	12187	.23882
	8	.090235	.05782	.866	09436	.27483
	10	.141705	.05590	.255	03676	.32017
10	1	266147 [*]	.04947	.000	42406	10823
	2	233309 [*]	.05725	.002	41606	05056
	3	277701 [*]	.05800	.000	46287	09253
	4	406329 [*]	.05725	.000	58908	22358
	5	065643	.05725	.979	24839	.11711
	6	091574	.04757	.652	24345	.06030
	7	083230	.05590	.896	26169	.09523
	8	051470	.05725	.996	23422	.13128

9	141705	.05590	.255	32017	.03676
*. The mean difference is significant at the	0.05 level.				

1: *B. subtilis* Control; 2: *B. subtilis* 0.1 mg/L c-AgNP; 3: *B. subtilis* 1.0 mg/L c-AgNP; 4: *B. subtilis* 5.0 mg/L c-AgNP; 5: *B. subtilis* 0.1 mg/L Ag-lon; 6: *E. coli* Control; 7: *E. coli* 0.1 mg/L c-AgNP; 8: *E. coli* 1.0 mg/L c-AgNP; 9: *E. coli* 5.0 mg/L c-AgNP; 10: *E. coli* 0.1 mg/L Ag-lon;

ANOVA									
		Sum of							
		Squares	df	Mean Square	F	Sig.			
Length	Between Groups	878.906	7	125.558	4.979	.000			
	Within Groups	1210.488	48	25.219					
	Total	2089.394	55						
Biomass	Between Groups	.762	7	.109	1.114	.370			
	Within Groups	4.693	48	.098					
	Total	5.455	55						
AgContent	Between Groups	.004	7	.001	10.641	.000			
	Within Groups	.003	48	.000					
	Total	.006	55						

 Table 9: ANOVA Results for Length, Biomass, and Root Ag Content of Z. mays

 Seedlings

Multiple Comparisons										
Tukey HSD										
						95% Cor	fidence			
			Mean			Inter	val			
Dependent	(I)	(J)	Difference			Lower	Upper			
Variable	Treatment	Treatment	(I-J)	Std. Error	Sig.	Bound	Bound			
Length	1.00	2.00	3.8014	2.4402	.772	-3.930	11.533			
		3.00	4.1389	2.4402	.690	-3.592	11.870			
		4.00	5.9264	2.4402	.251	-1.805	13.658			
		5.00	-6.8944	2.6467	.179	-15.280	1.491			
		6.00	-3.4611	2.6467	.891	-11.847	4.924			
		7.00	.9889	2.8010	1.000	-7.886	9.863			
		8.00	-1.6944	2.6467	.998	-10.080	6.691			
	2.00	1.00	-3.8014	2.4402	.772	-11.533	3.930			
		3.00	.3375	2.5109	1.000	-7.618	8.293			
		4.00	2.1250	2.5109	.989	-5.830	10.080			
		5.00	-10.6958 [*]	2.7121	.006	-19.288	-2.103			
		6.00	-7.2625	2.7121	.155	-15.855	1.330			
		7.00	-2.8125	2.8629	.975	-11.883	6.258			
		8.00	-5.4958	2.7121	.476	-14.088	3.097			
	3.00	1.00	-4.1389	2.4402	.690	-11.870	3.592			
		2.00	3375	2.5109	1.000	-8.293	7.618			
		4.00	1.7875	2.5109	.996	-6.168	9.743			
		5.00	-11.0333*	2.7121	.004	-19.626	-2.441			
		6.00	-7.6000	2.7121	.118	-16.193	.993			
		7.00	-3.1500	2.8629	.954	-12.220	5.920			
		8.00	-5.8333	2.7121	.399	-14.426	2.759			
	4.00	1.00	-5.9264	2.4402	.251	-13.658	1.805			
		2.00	-2.1250	2.5109	.989	-10.080	5.830			
		3.00	-1.7875	2.5109	.996	-9.743	6.168			
		5.00	-12.8208 [*]	2.7121	.001	-21.413	-4.228			
		6.00	-9.3875 [*]	2.7121	.023	-17.980	795			
		7.00	-4.9375	2.8629	.672	-14.008	4.133			

Table 10: Tukey's Post-Hoc Results for Length, Biomass and Root Ag Content of Z. mays Seedlings

		8.00	-7.6208	2.7121	.116	-16.213	.972
	5.00	1.00	6.8944	2.6467	.179	-1.491	15.280
		2.00	10.6958 [*]	2.7121	.006	2.103	19.288
		3.00	11.0333 [*]	2.7121	.004	2.441	19.626
		4.00	12.8208 [*]	2.7121	.001	4.228	21.413
		6.00	3.4333	2.8993	.933	-5.753	12.619
		7.00	7.8833	3.0409	.184	-1.751	17.518
		8.00	5.2000	2.8993	.627	-3.986	14.386
	6.00	1.00	3.4611	2.6467	.891	-4.924	11.847
		2.00	7.2625	2.7121	.155	-1.330	15.855
		3.00	7.6000	2.7121	.118	993	16.193
		4.00	9.3875 [*]	2.7121	.023	.795	17.980
		5.00	-3.4333	2.8993	.933	-12.619	5.753
		7.00	4.4500	3.0409	.822	-5.184	14.084
		8.00	1.7667	2.8993	.999	-7.419	10.953
	7.00	1.00	9889	2.8010	1.000	-9.863	7.886
		2.00	2.8125	2.8629	.975	-6.258	11.883
		3.00	3.1500	2.8629	.954	-5.920	12.220
		4.00	4.9375	2.8629	.672	-4.133	14.008
		5.00	-7.8833	3.0409	.184	-17.518	1.751
		6.00	-4.4500	3.0409	.822	-14.084	5.184
		8.00	-2.6833	3.0409	.986	-12.318	6.951
	8.00	1.00	1.6944	2.6467	.998	-6.691	10.080
		2.00	5.4958	2.7121	.476	-3.097	14.088
		3.00	5.8333	2.7121	.399	-2.759	14.426
		4.00	7.6208	2.7121	.116	972	16.213
		5.00	-5.2000	2.8993	.627	-14.386	3.986
		6.00	-1.7667	2.8993	.999	-10.953	7.419
		7.00	2.6833	3.0409	.986	-6.951	12.318
Biomass	1.00	2.00	.125125	.151931	.991	35624	.60649
		3.00	.099375	.151931	.998	38199	.58074
		4.00	.161000	.151931	.962	32036	.64236
		5.00	160500	.164793	.976	68261	.36161
		6.00	157000	.164793	.979	67911	.36511
		7.00	017000	.174400	1.000	56955	.53555

	8.00	093000	.164793	.999	61511	.42911
2.00	1.00	125125	.151931	.991	60649	.35624
	3.00	025750	.156336	1.000	52107	.46957
	4.00	.035875	.156336	1.000	45944	.53119
	5.00	285625	.168862	.693	82063	.24938
	6.00	282125	.168862	.705	81713	.25288
	7.00	142125	.178251	.992	70687	.42262
	8.00	218125	.168862	.897	75313	.31688
3.00	1.00	099375	.151931	.998	58074	.38199
	2.00	.025750	.156336	1.000	46957	.52107
	4.00	.061625	.156336	1.000	43369	.55694
	5.00	259875	.168862	.783	79488	.27513
	6.00	256375	.168862	.794	79138	.27863
	7.00	116375	.178251	.998	68112	.44837
	8.00	192375	.168862	.945	72738	.34263
4.00	1.00	161000	.151931	.962	64236	.32036
	2.00	035875	.156336	1.000	53119	.45944
	3.00	061625	.156336	1.000	55694	.43369
	5.00	321500	.168862	.555	85650	.21350
	6.00	318000	.168862	.569	85300	.21700
	7.00	178000	.178251	.972	74275	.38675
	8.00	254000	.168862	.801	78900	.28100
5.00	1.00	.160500	.164793	.976	36161	.68261
	2.00	.285625	.168862	.693	24938	.82063
	3.00	.259875	.168862	.783	27513	.79488
	4.00	.321500	.168862	.555	21350	.85650
	6.00	.003500	.180521	1.000	56844	.57544
	7.00	.143500	.189332	.994	45636	.74336
	8.00	.067500	.180521	1.000	50444	.63944
6.00	1.00	.157000	.164793	.979	36511	.67911
	2.00	.282125	.168862	.705	25288	.81713
	3.00	.256375	.168862	.794	27863	.79138
	4.00	.318000	.168862	.569	21700	.85300
	5.00	003500	.180521	1.000	57544	.56844
	7.00	.140000	.189332	.995	45986	.73986

		8.00	.064000	.180521	1.000	50794	.63594
	7.00	1.00	.017000	.174400	1.000	53555	.56955
		2.00	.142125	.178251	.992	42262	.70687
		3.00	.116375	.178251	.998	44837	.68112
		4.00	.178000	.178251	.972	38675	.74275
		5.00	143500	.189332	.994	74336	.45636
		6.00	140000	.189332	.995	73986	.45986
		8.00	076000	.189332	1.000	67586	.52386
	8.00	1.00	.093000	.164793	.999	42911	.61511
		2.00	.218125	.168862	.897	31688	.75313
		3.00	.192375	.168862	.945	34263	.72738
		4.00	.254000	.168862	.801	28100	.78900
		5.00	067500	.180521	1.000	63944	.50444
		6.00	064000	.180521	1.000	63594	.50794
		7.00	.076000	.189332	1.000	52386	.67586
AgContent	1.00	2.00	0000244	.00352505	1.000	0111928	.0111439
		3.00	0001154	.00352505	1.000	0112838	.0110528
		4.00	0000019	.00352505	1.000	0111703	.0111663
		5.00	.0269391 [*]	.00382345	.000	.01482529	.0390529
		6.00	0001717	.00382345	1.000	0122855	.0119420
		7.00	0004630	.00404636	1.000	0132831	.0123569
		8.00	0001533	.00382345	1.000	0122671	.0119605
	2.00	1.00	.00002446	.00352505	1.000	0111439	.0111928
		3.00	0000910	.00362725	1.000	0115832	.0114011
		4.00	.00002246	.00362725	1.000	0114697	.0115146
		5.00	.0269635 [*]	.00391787	.000	.01455060	.0393765
		6.00	0001472	.00391787	1.000	0125602	.0122656
		7.00	0004386	.00413570	1.000	0135416	.0126644
		8.00	0001288	.00391787	1.000	0125418	.0122841
	3.00	1.00	.00011549	.00352505	1.000	0110528	.0112838
		2.00	.00009103	.00362725	1.000	0114011	.0115832
		4.00	.00011349	.00362725	1.000	0113786	.0116056
		5.00	.0270546 [*]	.00391787	.000	.01464164	.0394675
		6.00	0000562	.00391787	1.000	0124691	.0123567
		7.00	0003475	.00413570	1.000	0134506	.0127555

	8.00	0000378	.00391787	1.000	0124507	.0123751
4.00	1.00	.00000199	.00352505	1.000	0111663	.0111703
	2.00	0000224	.00362725	1.000	0115146	.0114697
	3.00	0001134	.00362725	1.000	0116056	.0113786
	5.00	.0269411*	.00391787	.000	.01452814	.0393540
	6.00	0001697	.00391787	1.000	0125826	.0122432
	7.00	0004610	.00413570	1.000	0135641	.0126420
	8.00	0001513	.00391787	1.000	0125642	.0122616
5.00	1.00	0269391*	.00382345	.000	0390529	0148252
	2.00	0269635*	.00391787	.000	0393765	0145506
	3.00	0270546 [*]	.00391787	.000	0394675	0146416
	4.00	0269411 [*]	.00391787	.000	0393540	0145281
	6.00	0271108 [*]	.00418838	.000	0403808	0138408
	7.00	0274021 [*]	.00439281	.000	0413198	0134844
	8.00	0270924*	.00418838	.000	0403624	0138223
6.00	1.00	.00017172	.00382345	1.000	0119420	.0122855
	2.00	.00014726	.00391787	1.000	0122656	.0125602
	3.00	.00005622	.00391787	1.000	0123567	.0124691
	4.00	.00016972	.00391787	1.000	0122432	.0125826
	5.00	.0271108 [*]	.00418838	.000	.01384082	.0403808
	7.00	0002913	.00439281	1.000	0142090	.0136263
	8.00	.00001842	.00418838	1.000	0132515	.0132884
7.00	1.00	.00046306	.00404636	1.000	0123569	.0132831
	2.00	.00043860	.00413570	1.000	0126644	.0135416
	3.00	.00034756	.00413570	1.000	0127555	.0134506
	4.00	.00046106	.00413570	1.000	0126420	.0135641
	5.00	.0274021 [*]	.00439281	.000	.01348446	.0413198
	6.00	.00029134	.00439281	1.000	0136263	.0142090
	8.00	.00030976	.00439281	1.000	0136079	.0142274
8.00	1.00	.00015330	.00382345	1.000	0119605	.0122671
	2.00	.00012884	.00391787	1.000	0122841	.0125418
	3.00	.00003780	.00391787	1.000	0123751	.0124507
	4.00	.00015130	.00391787	1.000	0122616	.0125642
	5.00	.0270924 [*]	.00418838	.000	.01382239	.0403624
	6.00	0000184	.00418838	1.000	0132884	.0132515

7.00 -.0003097 .00439281 1.000 -.0142274 .0136079

*. The mean difference is significant at the 0.05 level.

1.00: *Z. mays* Control; 2.00: *Z. mays* 1.0 mg/L c-AgNP; 3.00: *Z. mays* 5.0 mg/L c-AgNP; 4.00: *Z. mays* 0.1 mg/L Ag-lon; 5.00: *Z. mays* + *B. subtilis* Control; 6.00: *Z. mays* + *B. subtilis* 1.0 mg/L c-AgNP; 7.00: *Z. mays* + *B. subtilis* 5.0 mg/L c-AgNP; 8.00: *Z. mays* + *B. subtilis* 0.1 mg/L Ag-lon.