FC2 9257

TIDAL MARSH WILDLIFE AND THEIR HABITATS

Franklin C. Daiber

College of Marine Studies and The School of Life and Health Sciences University of Delaware

Presented at

Delaware Academy of Science Fall Symposium November 10, 1977

CONTENTS

Introduction	n	1
Mollusca.	Pelecypoda:	4
(Gastropoda:	11
Arthropoda.	Crustacea:	15
	Insecta:	28
Aves.		55
Mammalia.		73
References (Cited	85

INTRODUCTION

Man has had an interest in tidal marshes for many centuries. They have provided him with a source of food for himself and his domestic animals as well as material for shelters. They have served as a buffer against storm damage inflicted against upland structures. Also they have been a source of obnoxious insects as well as disease dissemination. Of late years tidal marshes have been looked to as sites for industrial development, recreation, homes, waste disposal and sources of raw materials. As a result, the tempo of interest in tidal wetlands has stepped up with a corresponding increase in interaction and conflict.

Tidal marshes are a dynamic zone between the sea on one hand and the uplands on the other. They are under the profound influence of the flood and ebb of the tide, subjected to tidal erosion and sedimentation and the ingress of all manner of marine organisms. On the other hand there usually is an inflow of fresh water from upland sources and terrestrial animals have easy access to this semi-aquatic habitat.

Because these tidal wetlands are a meeting place between the sea and dry land they acquire characteristics derived from both, yet they display an individuality that clearly sets them aside as a unique habitat. Vegetational types are evident which are not found elsewhere. Therefore, distinctive plant-animal interactions become evident which have been characterized by Shanholtzer (1974) and Kraeuter and Wolf (1974) for the vertebrates and invertebrates respectively.

Teal (1962) delineated the vegetation of a Georgia tidal marsh and placed the animals associated with such vegetational zonation into three general categories: (la) terrestrial animals living in the marsh, (b) terrestrial or freshwater species living on the landward edge; (2) aquatic species with their center of abundance in the estuary, (a) estuarine species limited to the marsh low water level, (b) species in the streamside marsh, (c) estuarine species found well into the marsh; (3) marsh species with aquatic ancestors, (a) species with planktonic larvae and (b) species living entirely within the marsh. Teal recorded 57 percent of the aquatic species (2a and 2b) occupying the lowest portions of the marsh where exposure at low tide is the shortest. Those living on the marsh are living at one edge of their species distributions; their numbers maintained by migrations from the water and by adaptations and tolerance levels. Those living above the mud are subjected to great environmental stress and those penetrating farthest into the marsh have adapted burrowing habits. The remainder are tolerant enough to inhabit the entire marsh. Teal estimated the terrestrial species comprise almost half the marsh fauna, yet have made few adaptations to marsh living and are much less important in community energetics than the aquatic In the same vein, when Kraeuter and Wolf (1974) set species. aside those animals found in the creeks and pools and similar areas, only a few species remain to dominate the marsh surface.

Much work has been done on the game species (waterfowl) and furbearers (muskrats) that are characteristic numbers of

the tidal wetlands fauna (see Daiber, 1974, 1977, for reviews). Among the non-game species various taxonomic groups have attracted a good deal of attention, the salt marsh mosquitos (see numerous articles in various entomological journals), the Foraminifera (Lee and Muller, 1973; Phleger, 1970, 1977), and the fiddler crabs (Teal, 1958; Crane, 1975). Other groups such as the mites (Luxton, 1967), insects in general (Davis and Gray, 1966), beetles (Evans, et al., 1971), amphipods (Phillips, 1978), passerine birds (Tomkins, 1941; Marshall, 1948; Johnston, 1956) have received only spotty examination. Others like the protozoa (other than Foraminifera), oligochaetes, and nematodes have received essentially no attention. Herein exist many interesting studies of species composition and population dynamics.

While attention is focused on the local scene it seems desirable to establish a broader perspective. By so doing, attention can be called to the many interesting studies that have been carried out examining tidal wetland species of which nothing is known from this region. The present paper will review various parameters and their interactions that influence marsh animal distributions which in turn influence facets of life histories such as breeding, feeding and behavior. No attempt will be made to be all inclusive, whole groups of animals will not be covered such as the Foraminifera (see above citations for review).

MOLLUSCA. PELECYPODA:

The Ribbed mussel, Geukensia demissus (Dillwyn 1817) (= Modiolus demissus), is an obvious member of the marsh fauna if one searches along the creek banks and drainage ditches looking for this corregated shelled mollusc partially protruding from the marsh muds. It is restricted to such locations and is a passive feeder, ingesting material only as the flooding tide brings particulate matter past the incurrent siphon. During 1961, as part of a class exercise in Canary Creek marsh, Delaware, the average mussel density was $45.5/m^2$ within a one meter zone bordering drainage ditches. Beyond this the numbers declined to $6.0/m^2$ in the short Spartina alterniflora within a 10 m distance from the ditches. The mussel has been examined in the Sapelo Island marshes of Georgia where random samples gave an estimated population density of $7.8/m^2$ for the inhabited portions of the marsh. They were most dense near the heads of small creeks, an average of 52 animals/ m^2 being common in certain marsh types (Table 1). The tall Spartina (1-3 m) --Creek Head marsh, which occupies only six percent of the total marsh area, contained 46 percent of population by weight. The medium Spartina (0.5-1.0 m) levee had a reported density one-seventh of the tall Spartina -- Creek Head marsh but the levee, occupied 3.5 times as much area and contributed the second largest percentage of total weight (Table 1) (Kuenzler, 1961).

Clumped distributions within uniform marsh areas were a striking feature of the Sapelo Island mussel populations in that

Table 1. Population distribution and density of <u>Geukensia demissus</u> from the Sapelo Island marshes. (From Kuenzler, 1961, by permission of the editor of Limnology and Oceanography.)

	Marsh Type	% of Total	_ #/0.	2m ²	Dens	i + v	m^2	% To Popul	
No.	Description	Marsh	$\overline{\mathbf{X}}$ V	ariance	<u>#</u>		Wt.	#	Wt.
1	Muddy Creek banks	4	0		0	in de la composition de la com	.0	0	0
2	Tall Spartina edge	9	0		0		0	0	0
3	Medium Spartina levee	21	1.14	12.32	5.7		4.35	17	22
4	Short <u>Spartina</u> low elevation	16	1.24	9.29	6.2		3.0	15	12
5	Short <u>Spartina</u> high elevation	20	0.08	0.11	0.4		0.15	1	1 %
6	Salicornia spp. Distichlis spicata	9	0.23	0.37	1.1		0.40	2	1
7	Juncus roemarianus	4	2.64	26.4	13		6.15	9	7
8	Tall Spartina creek heads	6	10.4	146	52		31.6	47	46
9	Medium <u>Spartina</u> medium elevation	11	1.11	5.72	5.6		4.55	9	12
To	tal	100						100	100
Mea	an		1.33		6.66*		4.11		

*The difference in density of $6.66/m^2$ and $7.82/m^2$ reported in the text was due to the mat not being sieved. Mussels less than 25 mg were not recovered but did constitute 18 percent of the total number.

σ

sample variances were larger and more numerous in densely populated marsh types such as the creek heads where they often exceeded one meter in diameter or paralleled the creeks for many meters.

Kuenzler reported the vertical distribution to be approximately 200-240 cm above mean low water. The center of biomass was situated at 220 cm with the maximum density probably at this same level. It was calculated a mussel at the 220 cm elevation would be covered by 77 percent of the flood tides or about 18 percent of the time. The greater density by weight at low and medium elevations was thought to be made possible by longer feeding times. When comparing growth rates, it was always greater in the low marsh types (Short <u>Spartina</u>, low elevation; Tall <u>Spartina</u>, creek heads) than in higher marsh types (Medium <u>Spartina</u>, levee; <u>Salicornia-Distichlis</u>; Medium Spartina, medium elevation).

Lent (1967a, 1967b, 1968, 1969) elaborated on the significance of intertidal distribution and air-gaping in the Ribbed mussel. He compared those mussels typically found in tidal marsh muds with those located on exposed surfaces such as bridge pilings. There was a significant difference in the shell length-height ratios; bridge mussels being longer than marsh mussels. Lent (1967b) attributed the difference between habitats to the crowding effect of clumping in the marsh.

Kuenzler (1961) and Lent (1967b) found an excellent correlation between shell height and the cube root of the shell-free, dry body weight. However, Lent did not find any

relationship between the shell weight-meat weight ratios and intertidal height. Since the mussel is exposed during part of the tidal cycle and air-gapes during exposure, respiration could be carried on and such exposure would thus not reduce meaty tissue weight as rapidly. Air-gaping by increasing the efficiency of aerial respiration would permit a greater landward penetration of the intertidal zone.

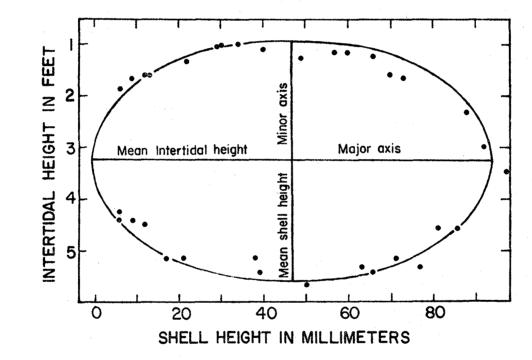
In his 1968 paper, Lent pointed out median survival times in gaseous environments to be proportional to the volume of oxygen present. Lent reported oxygen tension during exposure at 15 mm Hg and 110 mm Hg while inundated (Figure 6, 1968). Lower rates of respiration in air were explained by these lower oxygen tensions of the mantle cavity fluid in air-gaping mussels. Although the rate was lower in air than when submerged, more oxygen was consumed because of extended exposure (82% aerial exposure, Kuenzler, 1961). Mussels survived for long periods exposed as long as they were protected from dessication.

Dessication and temperature changes are environmental parameters with which the exposed mussels are confronted. Lent (1969) considered the mussel to possess a high tolerance to dehydration and a very high enzyme thermostability. Temperature had a significant effect on the rate of dessication at low but not at high humidity (Lent, 1968). Dessication resulted from water loss by air-gaping and was a physical phenomenon in which a surface to volume relationship caused small mussels to reach a median lethal weight loss of 36-38% more rapidly than large mussels.

Mussels can survive over a temperature range of at least 56° C with a recorded minimum of -22° C (Lent, 1969). An upper LD50 for a 10 hour heat stress fell between 36.4 and 37.8°C with the large mussels more labile to thermal stress than small animals (Lent, 1968). Such temperature tolerance would explain the geographic distribution ranging from Prince Edward Island in the north to South America.

Salinity limits are wide but not firmly established. The lower limit is at least $3^{\circ}/\circ\circ$. Mussels can survive water losses of 36-38% by dessication and have at least 71% of their tissue water frozen (Lent, 1969). Therefore the upper limit must be high. Intact animals were able to survive direct transfer into a salinity of $75^{\circ}/\circ\circ$ from an environmental salinity of $30^{\circ}/\circ\circ$. Lent reported <u>Guekensia</u> to be isomotic between 9 and $43^{\circ}/\circ\circ$. Waugh and Garside (1971) reported the mussel to be an osmoconformer at higher salinities but changing to osmoregulation below $10^{\circ}/\circ\circ$.

Mussels living in moist marsh mud occupied a higher intertidal height than those living on exposed bridge pilings. When the intertidal position of bridge mussels was examined as a function of shell size the population fell within an ellipse (Figure 1). The intertidal range was greatest for medium size mussels and was reduced for both larger and smaller groups. Lent (1968) considered this ellipse to be a geometric form within which the natural population could live. The upper surface was generated by the physical factors of dessication and thermal stress whereas predation and competition are biotic factors determining the lower surface.



: 3

Figure 1. Intertidal height as a function of shell height for 33 mussels on the periphery of the distribution of Canary Creek bridge. (From Lent, 1968, by permission of the editor of Biol. Bull.)

43

International Contraction

Lent (1969) proposes air-gaping in the mussel to be a significant behavioral adaptation which permits aerial respiration and penetration of the higher intertidal zone. In addition it is physiologically and biochemically adapted in that it is both eurythermal and euryhaline. These two adaptations provide a tolerance toward dessication, salinity variation, thermal stress, and possibly anaerobic conditions. However, there is no obvious morphological adaptation thus the pelecypod body plan precludes further landward penetration.

An ongoing study of another marsh surface bivalve Cyrenoidea floridana depicts an interesting parallel to the gaping behavior of Geukensia demissa. So far this small bivalve which gets up to 9 mm in length has been found to have a Delaware distribution restricted to the southern portion of the state, not extending beyond the Great Marsh of Lewes. This clam appears to alternate between air-gaping and keeping the valves closed when the animal is exposed to air. Humidity must be quite high in the clam's habitat as individuals tend to be found where there is a thin covering mat of algae and organic debris. Laboratory observations record air bubbles being trapped in the mantle cavity of the juveniles and older individuals. These bubbles and the abundance of periostracal hairs increasing the shell surface area give a positive buoyancy to the clam. Because of the type of reproduction one might expect the young to be clustered near the adults. Preliminary results suggest this is not the case. The clams tend to be distributed among the short Spartina alterniflora and

<u>Distichlis spicata</u> plants of the higher reaches of Canary Creek marsh. Sampling has demonstrated a negative correlation between adult clam and stem densities but a positive association between juveniles and stems. Dispersion of the clam appears to be brought about by the positively buoyant clams floating in the very gently flowing water as it floods over the marsh surface. The plant stems probably provide impedence to the water flow causing the juveniles to be lodged between or swept up against the stems, thus increasing their densities where the stems are the densest (P. Kat, personal communication).

GASTROPODA:

The Coffee bean snail, <u>Melampus bidentatus</u>, is an abundant member of the tidal marsh fauna and has received the attention of numerous workers over the years. Hauseman (1932) found few snails where tides flood the tall marsh grass and where fiddler crabs are common, with snails more numerous above the high tide mark. According to Holle (1957), <u>Melampus</u> can be found only in salt marshes flooded by normal tides and avoids tidal submersion by climbing grass stems or debris. In contrast, Hackney (1944) reported <u>Melampus</u> as common on the mud flats of Beaufort, North Carolina.

Dexter, in a series of papers on the marine molluscs of Cape Ann, Massachusetts region (1942, 1944, 1945), related <u>Melampus</u> distributions to vegetation. In 1942, he often found individuals in small groups under solid objects, occasionally finding the snail in the upper margins of thatch grass or

<u>Spartina glabra</u> (= <u>S</u>. <u>alterniflora</u>) marsh. Small clusters of <u>Melampus</u> were found in the <u>Spartina patens</u> marshes of Little River (1944). Dexter was of the opinion in his 1945 paper that the high marshes of <u>Spartina patens</u>, located for the most part above mean high water, were the only important habitat for Melampus bidentatus.

During a 1961 class exercise in Canary Creek marsh where dwarf <u>Spartina alterniflora</u> dominated the vegetation the mean density of <u>Melampus</u> was $645.2/m^2$. The very large standard deviation of 497.2 suggested a very uneven distribution of snails across the marsh surface, tending to decrease toward the marsh borders. It was noted that those sampling sites located in areas devoid of vegetation had very few or no snails. The inference is the snails crawled up the grass stems to avoid inundation, a pattern of behavior that has been observed many times. Parker (1976) reported similar densities for various marshes bordering both sides of Delaware Bay (Delaware: Canary Creek 693 \pm 340; New Jersey: Dras Creek 445 \pm 184, Fortesque 420 \pm 175.

In his study of the Poropotank River area of Virginia, Kerwin (1972) found the Coffee bean snail associated with the brackish water marshes dominated by <u>Spartina alterniflora-</u> <u>Scirpus robustus</u> stage and <u>Spartina cynosuroides</u>; and with the salt marshes dominated by <u>Spartina alterniflora</u> (short form), <u>S. patens-Juncus-Distichlis</u> zone, in the <u>Distichlis</u> zone and in the upper levels of the area dominated by the dwarf form of <u>Spartina alterniflora</u> of Woods Hole area of Massachusetts. They go on to say Melampus occurred largely in the upper two-

thirds of the zone lying above mean high water of neap tides (MHWN) and below the mean high water of spring tides (MHWS) but was also found in the zone lying between MHWS and the extreme upper limit worked by any tides (Figure 2). They described the snail as occupying the upper 12 percent of the intertidal zone which was bathed for only eight of the 354.4 hours (or 2.3 percent) of each semilunar cycle. Apley et al. (1967), Apley (1970), and Russell-Hunter et al. (1972) described the reproductive cycle of the snail as being synchronized to summertime spring tides.

Melampus bidentatus tends to be isosmotic. Russell-Hunter et al. (1972) submerged early egg masses in water with salinities of 25-110% sea water and found these masses to act like osmometers. Kerwin (1972) recorded Melampus densities of 0.23/m² in the Poropotank River, Virginia, brackish marsh where the salinity range was $4.11-9.38^{\circ}/\circ o$. In that same marsh where the salinity was higher (9.38-14.72⁰/oo) snail densities were 7.24/m². No Melampus was recorded in the fresh and slightly brackish marshes with salinities below 4.11°/00. Parker (1976) recorded much the same pattern for the Delaware Bay marshes; larger numbers in areas of higher salinities. Parker provided some information on the brackish water snail, Detracia floridana, a snail easily confused with Melampus. Detracia abundance tended to increase with declining salinities, apparently totally absent in the highly saline marshes.

Utilizing the aging procedure employed by Apley (1970), Parker (1976) demonstrated a general pattern of increasing length

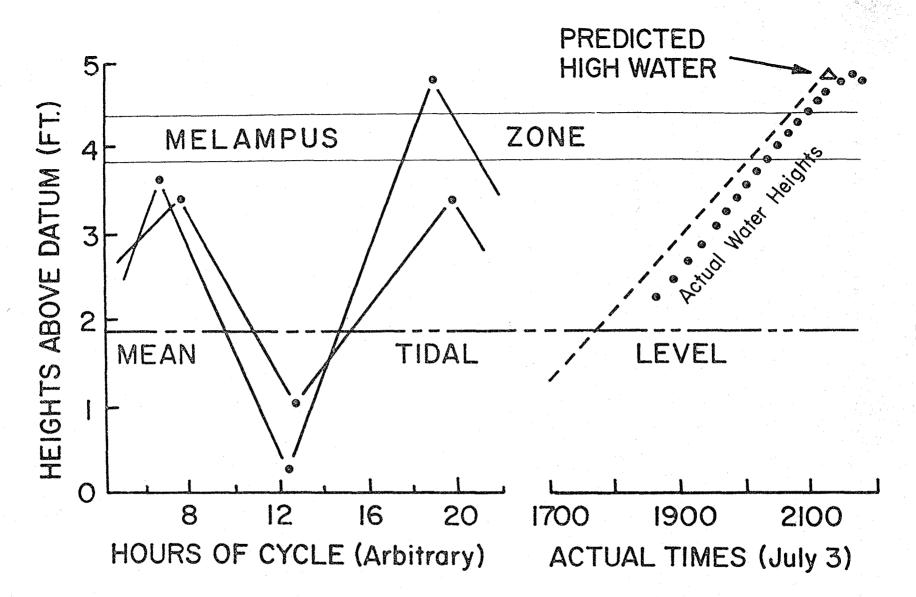


Figure 2.

e 2. The relation of the vertical zone (3.9 to 4.4 ft above datum) occupied by natural populations of <u>Melampus</u> at Little Sippewisset, typical ranges of spring (-0.02 to 4.77 ft) and neap (0.9 to 3.28 ft) tides, and the observed time course of actual tidal heights during one high water of springs. Note that the mean level of all tides (MTL) does not necessarily correspond to mean sea level (MSL). (From Figure 8 in Russell-Hunter et al., 1972, by permission of the editor of Biol. Bull.) in millimeters within each age class along a decreasing salinity gradient. This was particularly evident in those samples taken from the Broadkill River marshes. Laboratory observations by Parker clearly portrayed the euryhaline nature of the snail. Larvae were highly active at salinities ranging from 15 to $40^{\circ}/\circ\circ$, slightly less so at $10^{\circ}/\circ\circ$ and they showed a markedly reduced activity at $5^{\circ}/\circ\circ$. Approximately 90 percent of the larvae survived for 10 hours in salinities varying from 2.5 to $40^{\circ}/\circ\circ$. None survived in fresh water. Parker has reported the adults to be osmoconformers over the salinity range of 8 to $51^{\circ}/\circ\circ$. Above and below these levels osmoregulation might be indicated. However, Parker pointed out the apparent regulation may be merely avoidance reaction as the animals were less active, withdrew into their shells and there was a copious mucus flow above and below $8-51^{\circ}/\circ\circ$.

ARTHROPODA. CRUSTACEA:

Fiddler crabs are an obvious scurrying faunal component of tidal marshes. Pearse (1914) identified three characteristics strongly associated with these crabs, having watched 13 species in North and South America and in the Philipines. (1) Fiddler crabs are diurnal, whereas other crabs are nocturnal as well. (2) A substrate specifity is displayed giving rise to zonation in habitat selection. (3) Fiddler crabs feed on tidal flats during low tide with greatest activity on the ebbing flow. Frantic activity by the crabs was reported by Pearse as the tide began to flood. Crabs retreat to their burrows during high tides

and prolonged flooding, hastily gathering mud pellets to plug the entrance as the tide floods over the marsh surface. They open their burrows but will remain in them during extended periods of low tides which tend to dry out the tidal flats. Dry ing tends to inhibit the feeding process for the crabs.

Marsh crabs can be divided into three behavioral groups associated with tidal cycles (Teal, 1959). (1) Eurytium limosum, Panopeus herbsti, Sesarma reticulatum are active when the tide is high or when the sky is cloudy and the sun is not shining. When the marsh is uncovered at low tide they are in their burrows, near the top, either exposed to air or in the water. Crichton (1960) described Sesarma reticulatum clearing mud from its burrow as the tide receded, depositing it at either side of the entrance where a complete canopy was frequently formed. (2) Uca pugnax and U. minax are active principally when the tide is out and do feed under water when the tide is high or in tide pools during low tide. (3) Uca pugilator and Sesarma cinereum are active only in the air. U. pugilator is in its burrow at high tide while S. cinereum climbs above the water level. The first group is found in the low marsh at the low tide level and the second group occupies an intermediate zone. The third group, with one exception, is located at the high tide level. Teal found colonies of Uca pugilator just above the low tide level on the creek banks of the Georgia marshes. Crabs in the first group build burrows with numerous openings which allow the burrows to fill with water. Uca pugnax and U. minax dig burrows with single

openings but seldom plug them. <u>Uca pugilator</u> almost always plugs its burrow entrance at high tide and thus remains in a pocket of air. However, Teal considers the air pocket incidental to the primary need for a hiding place.

Tidal flooding plays a profound role in the lives of these crabs beyond the daily ebb and flow. <u>Uca minax</u> digs a burrow downward to the point where the chamber is at low tide water level (Gray, 1942). This species excavates new burrows during prolonged low water periods or when the burrows are filled with water from extended flooding. In either case, the burrow is dug to a depth to assure flooding of the chamber. Teal (1958) suggests predation may be influencial in restricting the distribution of <u>Uca pugnax</u> at the lower intertidal levels. This species frequently remains outside its burrows as the tide is flooding while <u>Uca pugilator</u> almost never does. The former is therefore more subject to predation by aquatic organisms.

While Pearse (1914), Gray (1942) and Teal (1958) point out the importance of substrate and tidal flooding in the lives of fiddler crabs, specific associations between crabs and vegetation are also evident. Gray describes <u>Uca minax</u> habitat at Solomons, Maryland, as a high water sand and mud substrate sparsely covered by <u>Spartina cynosuroides</u> and <u>Paspalum</u> sp. <u>Uca minax</u> is found in association with 15 species of marsh plants, occurring with five species more than 20 percent of the time in the Poropotank River marshes of Virginia: <u>Spartina</u> <u>alterniflora</u>, 87.8%; <u>Scirpus</u> robustus, 32.9%; <u>Distichlis spicata</u>, 31.7%; Spartina patens, 24.4%; and S. cynosuroides, 20.7%.

High population densities are reported in the tall and short <u>S. alterniflora</u> and at the edge of the <u>D. spicata-S. patens</u> community with densities ranging from 0-76 burrows/m² (Kerwin, 1971).

Schwartz and Safir (1915) described the burrowing of Uca pugnax where there was dense vegetation cover, while U. pugilator burrowed in sparse vegetation or where it was absent. Teal (1958) found U. pugnax prefers to burrow where there is vegetation cover and is thus found to be most numerous in the Short marshes wherein the Spartina is 10 to 50 cm tall on firm substrate containing varying amounts (40-70%) of sand and the Medium marshes with Spartina one meter tall on soil that is soft with small amounts of sand (0-10%). Gray (1942) and Teal (1958) found Uca minax in certain of the Short Spartina high marshes where the grass is very short (10-20 cm) and the sand content is about 30% and water tends to stand. Uca pugilator was found primarily in the Salicornia-Distichlis marsh and on the tidal creek banks. The former is high, sandy marsh with sparse vegetation and covered only by spring tides. In both situations, U. pugilator seems to be attracted to sparse or absent vegetation.

The large burrows of the Marsh crab <u>Sesarma reticulatum</u> have been reported within five meters of the tidal water of Canary Creek marsh of Delaware and associated with <u>Spartina</u> <u>alterniflora</u> (Daiber and Crichton, 1967). Crichton (1960) considers this marsh crab responsible for patches of bare mud along the creeks and ditches as it crops the cordgrass and

undermines the roots. Frequently these bare areas are swathes separating several feet of lush tall <u>Spartina alterniflora</u> further back on the marsh surface. At other times, they border directly on the bare mud of the stream bank. It is in these swathes that the burrows are the largest and most numerous. The number of burrow openings range from one to ten per square yard with four to five being an average. Crichton considers this burrowing activity to enhance the erosive action of the water and to keep the banks low enough so the tides can flood over the marsh surface.

Salinity plays some part in fiddler crab distributions. The description of the habitat for <u>Uca minax</u> in Solomons Island, Maryland (Gray, 1942) which included <u>Spartina cynosuroides</u>, would suggest this to be a brackish water species. Kerwin (1971) found <u>U. minax</u> widely distributed in brackish and salt marshes of the Poropotank River, Virginia, where the salinities were greater than $2^{\circ}/_{oo}$ and less than $16^{\circ}/_{oo}$ (Table 2).

Miller and Maurer (1973) reported <u>Uca pugnax</u> from high salinities $(21-29^{\circ}/\circ o)$ while <u>U</u>. <u>minax</u> was most abundant in lower salinities and even into fresh waters $(0-12^{\circ}/\circ o)$ of three Delaware tidal streams. An examination of Table 3 shows a shift in abundance for the two species between 15 and $18^{\circ}/\circ o$ for the Broadkill River; i.e., <u>U</u>. <u>pugnax</u> and <u>U</u>. <u>minax</u> being more abundant at 18 and $15^{\circ}/\circ o$, respectively. A similar shift in abundance is shown between 23 and $25^{\circ}/\circ o$ and 18 and $20^{\circ}/\circ o$ for the Mispillion and Murderkill Rivers respectively. Miller and Maurer displayed a significant (at 0.005) positive correlation

Table 2.

PERCENT OF TOTAL SAMPLES, MEAN DENSITY PER SQUARE METER, AND THE MEAN SALINITY RANGE IN EACH MARSH TYPE IN WHICH UCA MINAX BURROWS OCCURRED. (Modified slightly from Table 1, Kerwin, 1971, by permission of the editor, Ches. Sci.)

Marsh Type	Frequency	Density (No./m ²)	Salinity (º/oo)
FRESH	0.0	0.0	0.3 - 0.8
SLIGHTLY BRACKISH	0.0	0.0	0.8 - 4.1
BRACKISH	71.4	7.9	4.1 - 9.4
SALT	83.8	14.3	9.4 - 14.7
TOTAL AREA	66.7	12.4	0.3 - 14.7

Table 3.

Abundance of fiddler crabs in relation to salinity in three Delaware rivers. (Table 1, Miller and Maurer, 1973, by permission of the editor, Ches. Sci.)

1

	Broadkill River			Mispillion River			Murderkill River		
Sta- tion	No. U. pugnax		Salinity %。	No. U. pugnax			No. U. pugnax		Salinity %。
1	0	0	29	30	0	27	30	0	21
2	0	0	29.	30	0	26	30	0	21
3	31	Ο	29	30	0	27	30	0	21
4	25	1	29	30	0	26	30	0	21
5	45	1	29	26	0	26	27	3	21
6	45	2	25	29	4	26	18	12	21
7	37	0	24	20	1	26	23	7	21
8	23	17	23	10	10	25	19	11	21
9	17	7	21	8	20	23	16	14	20
10	16	8	20	8	22	21	6	24	18
11	32	18	18	0	22	15	7	23	11
12	4	13	15	0	0	13	3	27	9
13	3	9 * * *	12	0	30	10	6	24	6
14	6	36	9	0	0	5	0	30	5
15	1	35	8	0	0	4	0	30	3
16	0	30	5	Ο	0	1	0	0	1
17	0	32	1	0	0	0	0	0	0.5
18	0	4	0	Ó	0	0	0	0	0
19	0	3	0	0	0	0	0	0	0
20	0	2	0	0	0	0	0	0	0

20a

between abundance and increasing salinity for <u>Uca pugnax</u> and a significant negative correlation for <u>Uca minax</u> in all three rivers. They went on to say that the mid-point of the crabs distributions, where both species were equally abundant was approximately at 12, 11 and $8^{\circ}/\circ \circ$ in the Broadkill, Mispillion and Murderkill Rivers, respectively. Further examination of Table 3 would suggest the points of equal abundance for both species to be closer to $18-25^{\circ}/\circ \circ$, depending on the river examined. The authors speculated that such mid-points may be the areas where neither species is favored by competitive advantage or where both are equally adapted to salinity regardless of competition.

Something other than salinity may be playing a role in such distributions. Students of the present author reported numerous <u>U</u>. <u>pugnax</u> and few <u>U</u>. <u>minax</u> from one area of the Blackbird Creek marsh in Delaware where the stream salinity was recorded at $4-5^{\circ}/\circ o$. This is supported by Teal (1958) and Kerwin (1971) who suggest substrate as well as salinity and competition influence fiddler crab distributions. Both Teal (1958) and Miller and Maurer (1973) suggested <u>U</u>. <u>pugilator</u> has a preferred salinity range between that of <u>U</u>. <u>minax</u> and <u>U</u>. <u>pugnax</u>.

Teal (1958) tested the salinity tolerance for the three species of <u>Uca</u> by placing them in finger bowls with differing salinities ranging from $0^{\circ}/00$ to $58^{\circ}/00$. It was only in the 0 to $7^{\circ}/00$ salinity that there was a difference in survival. All species survived for three weeks at the higher salinities

and in an experiment where the salinity was varied between 0 and $30^{\circ}/00$ on an alternate day basis. At $0^{\circ}/00$, 50 percent of <u>U. minax</u> had not died at the end of three weeks when the experiment was terminated, 50 percent of <u>U. pugilator</u> died at the end of three and one-half days, while 50 percent of <u>U. pugnax</u> died within one and one-half days. At $7^{\circ}/00$, 50 percent of the other two species survived over the test time of 10 days.

Teal (1958) also provided these three crab species with a choice of freshwater or a salinity of approximately $30^{\circ}/\circ o$. Both sexes of <u>U</u>. <u>minax</u> showed a statistical preference for fresh water over salt. The males of both <u>U</u>. <u>pugilator</u> and <u>U. pugnax</u> showed a significant preference for salt over fresh water. The females of both species were found more often in salt water but the differences were not statistically significant. In all cases the females showed a less strong preference (Table 4).

When examining <u>U</u>. <u>pugnax</u> and <u>U</u>. <u>pugilator</u> at higher salinities, Green et al. (1959) found both species to be hypoosmotic regulators. The sera for those crabs in 100 percent sea water were 12 percent lower in osmotic concentration and for those animals in 175 percent sea water, 22 percent lower than their respective media. Lockwood (1962), in a discussion of osmoregulation among the crustacea, considered <u>U</u>. <u>minax</u>, <u>U</u>. <u>pugilator</u> and <u>U</u>. <u>pugnax</u> to belong to that group of animals wherein the blood is hypo-osmotic to the medium in highly saline water and hyperosmotic in low salinities. It is apparent that

Table 4.

PERCENT OF MALES AND FEMALES OF THREE SPECIES OF FIDDLER CRABS WHICH PREFER FRESH OR SALT WATER ($-30^{\circ}/\circ\circ$) IN A 24-HOUR EXPERIMENT. (P IS PROBABILITY THAT RESULTS IN ROW ARE NOT DIFFERENT.) (Table III, Teal, 1958, by permission of the editor, Ecology).

Percent Preferring

	Fresh Water	Sea Water	P
U. minax male female	68 60	32 40	0.001 0.024
<u>U. pugilator</u> male female	40 48	60 52	0.02 0.6
U. pugnax male female	20 35	80 65	0.001 0.058

these three species have different salinity preferences, wide ranges of tolerance and differing abilities of adjusting to a salinity range within the framework of being hypo- and hyperosmotic regulators.

There is general agreement about substrate preferences by fiddler crabs. This specificity gives rise to fiddler crab zones in habitat selection (Pearse, 1914; Gray, 1942; Teal, 1958; Kerwin, 1971). Uca pugnax burrows in mud (Pearse, 1914) and Teal (1958) found it on the levee marsh and short Spartina low marsh where the sand content was 0 to 10 percent as well as where the sand content was greater. He concluded U. pugnax distribution is determined by a preference for a vegetated muddy substrate and salt water. In contrast, both Pearse (1914) and Teal (1958) recorded U. pugilator with a preference for sand. Teal found this species to be dominant in the Salicornia-Distichlis marsh where the soil was 80 to 95 percent sand. However, in the short Spartina high marsh where the sand content is 40 to 70 percent, Teal reported U. pugilator did not burrow to any appreciable extent. In laboratory tests it burrowed in sand above the water level, presumably because such burrows below the water would be more apt to collapse. Teal concluded U. pugilator preferred sandy habitats but could not colonize all such habitats because of competition with the other species of Uca. Both Gray (1942) and Teal (1958) found U. minax displaying a preference for sand and mud with the sand content exceeding 30 percent (Teal). The soil had to be wet and the salinity could vary over a wide range. Gray (1942) reported the

presence of <u>Spartina cynosuroides</u> coinciding with the presence of <u>U. minax</u> and Teal concluded the species prefers a mud or variable substrate and fresher water.

Results indicate <u>Uca minax</u> prefers substrates of high energy value, i.e. high organic content. This appears to be the case even though such substrates characteristically have the lowest oxygen levels. Substrate oxygen is not considered to be critical to habitat selection and this species can revert to an anaerobic mode of respiration during those times when it is submerged in the water of the burrow (Teal and Carey, 1967). Therefore, organic materials for feeding would be the primary component in habitat selection (Whiting and Moshiri, 1974). Interestingly, Teal (1959) considers <u>U. minax</u>, which does not possess any temperature acclimation, has a lower respiratory rate and is more dependent on oxygen tension, to be greatly limited in distribution in a typical salt marsh. However, as Teal points out, competition probably plays an important role in Uca distributions.

In addition to amounts of organic material and sediment particle size, the differentiation of feeding mechanisms among fiddler crabs had a bearing on crab distributions (Miller, 1961). There are two general processes of feeding involved: flotation and the coordinated action of the mouth parts. There is a copious use of water from the gill cavities to flood the buccal cavity during feeding. There is always the need for water replacement. To continue feeding the crab must have access to an external water supply to replenish its respiratory water. The inclusion

of the flotation process is one major factor keeping <u>Uca</u> from living in a terrestrial habitat. Pearse (1914) comments about these crabs scooping up mud with the chilipeds, feeding on the tidal flats as the tide ebbs. They will remain in their burrows when the flats dry out, feeding apparently being inhibited.

<u>Uca pugnax</u> has been observed to leave its burrow during feeding, dip its thorax to take up water, finding insufficient water, return to the burrow and a source of respiratory water to take up feeding again. This may explain the high degree of burrow territoriality observed for <u>Uca pugnax</u>. Such territoriality has not been observed for <u>U. minax</u> which feeds in muddier areas away from the burrow. The need for water for the flotation process for <u>U. pugilator</u> which inhabits sandy areas has more pronounced effect on this crab's movements. The crab may have to feed at the water's edge away from the burrow should water not remain in the burrow during low tide.

Of these three fiddler crabs $\underline{Uca \ minax}$ has the least modified mouth part hairs. It prefers to feed in low areas where the mud is very fluid. The material carried to the mouth is very fine and little sorting is needed before ingestion. The presence of flat-tipped hairs on the second milliped meropodites of \underline{U} . <u>minax</u> is believed to enhance the sorting process and to enable it to live in a coarser substrate where food is less available. The modification of hair tips into a spoon shape on the second maxilliped meropodites of \underline{U} . <u>pugnax</u> enables this crab to feed in coarser substrates. The luxuriant

fringe of plumose hairs on the outer edges of the second maxillipeds and on the maxillae enables \underline{U} . <u>pugnax</u> to feed on a muddy substrate.

The greatest modification of the hairs of the mouth parts occurs in <u>Uca pugilator</u>, greatly increasing the efficiency of the mouth parts to sort out coarse material. In addition to having to select against coarse mineral material, fiddler crabs are also faced with the greater paucity of food materials in a sandy habitat. In <u>Uca pugilator</u> the efficiency of sorting is further enhanced by a cleaning action which involves drawing sand particles over the bristled surface of the first maxilliped basal endites by the second maxillipeds. This cleaning process retains the bulk of the food particles which are washed by respiratory water toward the base of the buccal cavity.

The import of this is \underline{U} . <u>minax</u> cannot efficiently sort out coarse material and must thus have access to abundant food material on a silty substrate. The more advanced mouth parts of \underline{U} . <u>pugnax</u> enable it to have a wider distribution, living in the sandy areas of a new marsh. The reduced ability of \underline{U} . <u>pugnax</u> to sweep sand particles clean of food probably restricts its ability to live on a sandy beach. The highly modified mouth parts of <u>Uca pugilator</u> enable it to live in sandy areas. When \underline{U} . <u>pugilator</u> feeds in a marsh it cannot separate out the fine silty material and is not as efficient in getting food as when it sweeps the sand grains of food particles in a coarse substrate (Miller, 1961). INSECTA:

Insects are a common component of the tidal marsh fauna, occupying a variety of habitats. Relatively few families can be classed as pests and those noxious to man are of considerable economic importance and thus have received a great deal of attention. Much work has been devoted to the control of salt marsh mosquitos, beginning with Smith (1902) who recognized the relationship between mosquito distributions and zonation of salt marsh vegetation. He advocated filling those depressions on the marsh surface adjoining the uplands where mosquitos breed, or ditching so tidal action could circulate water through such areas.

Tidal inundation has been identified as a factor limiting the distribution of <u>Aedes</u> species of mosquitos on Delaware salt marshes (Connell, 1940). Young larvae fail to appear in portions of the marshes flooded by tides as frequently as 25 days per lunar month. Abundant breeding can be expected only in portions of <u>Spartina alterniflora</u> marsh where the frequency of tidal inundation is less than eight days per lunar month. Table 5 depicts the relationships between vegetation and insect numbers for <u>Aedes sollicitans</u> and <u>Aedes cantator</u>, wherein most of the mosquitos were taken in the <u>Spartina patens</u> zone. LaSalle and Knight (1974) found the <u>Juncus</u> marshes in Carteret County, North Carolina, to be subjected to a higher rate of tidal flooding during the mosquito breeding season than the marshes examined in Pamlico County. Consequently, there was a much lower level of Aedes breeding in the former area than in

Table 5.

AVERAGE NUMBER OF <u>AEDES</u> LARVAE AND PUPAE PER DIP FROM THE SALT MARSHES OF EGG ISLAND, NEW JERSEY (From Ferrigno, 1958).

Aver. number of larvae and pupae per dip

VEGETATION	A. cantator	<u>A.</u> sollicitans	TOTAL
Spartina patens	0.31	6.95	7.26
S. alterniflora	0.01	0.37	0.38
Mixed	0.29	2.42	2.71

Pamlico County. The higher flooding frequency in the Carteret County marshes probably accounted for the presence of <u>Anopheles</u> and <u>Culex</u> mosquitos which were as abundant as or more so than, <u>Aedes</u>. Ferrigno (1958) demonstrates the relationship between flooding and water depth on the success of mosquito breeding caused by Snow geese feeding activity (Table 6). In addition, the restriction of regular tidal flushing in vegetational zones for agricultural purposes has an impact on the numbers of mosquitos produced (Table 7). Diking enhances the growth of salt hay, <u>S. patens</u>, by preventing flooding with salt water. Later reflooding produces great broods of mosquitos, mostly <u>Culex salinarius</u> and <u>Aedes sollicitans</u>.

There appears to be some uncertainty from the earlier literature about the relationship between tidal flooding and vegetation zonation on one hand and the distributions of larval greenhead flies (Tabanidae). Hansens (1952) found them in the wetter portions of the salt marsh while Gerry (1950) described development as taking place primarily in the upper marsh reached only by the higher tides. He went on to say evidence indicates that larvae originate in the creeks from which they migrate to thatch piles at the head of the marsh. Wall and Doane (1960) found various species of tabanids well distributed over the surface of various Cape Cod marshes. Bailey (1948) rarely found <u>Tabanus nigrovittatus</u> where there was standing water on the marshes in the vicinity of Newbury, Massachusetts. Jamnback and Wall (1959) found the larvae associated with several species of salt marsh vegetation but most abundant in Spartina

Table 6.

Mosquito breeding at stations affected by feeding Snow geese in New Jersey marshes. (From Ferrigno, 1958)

	Descrip	tion of eat-outs		i sin			
W	ercent Denuded Aithin 70 foot Radius of tation Marker	<u>Vegetation Type</u>	Water Depth in _Inches_	Numbe <u>Larvae</u> per dip		Number of dips	
A	. 10-30	S. patens	0-2	12.71	6607	520	
В	10-40	<u>S. alterniflora</u>	0-6	1.56	2119	1355	
С	50-90	<u>S. patens-S. alterniflora</u>	1-8	0.05	25	460	
D	50-90	<u>S. alterniflora</u>	1-6	0.03		490	
E	100 (Ponds)	S. patens-S. alterniflora	5-24	0.002	3	1401	

Table 7.

The numbers of mosquito larvae and pupae sampled from natural and diked salt marshes of the Caldwalder Tract, New Jersey. (A portion of Table 2, Ferrigno, 1959.)

	Average number larvae-pupae per dip					
Vegetation Type	<u>Culex</u> salinarius	Aedes sollicitans	Dips	Larvae- Pupae		
UNDIKED			·			
S. alterniflora	0	0.0001	8280	1		
S. patens	0.11	2.74	1080	3293		
P. virgatum	0.003	0	600	2		
Woodland swamp	0.02	0.01	840	620		
DIKED						
S. alterniflora	0.26	4.22	360	1701		
S. patens	0.72	3.54	2760	13376		
S. cynosuroides	2.94	4.66	240	1988		
P. virgatum	0.21	0.75	1320	2219		
D. spicata	0	3.52	600	2761		
J. gerardi	0	2.86	240	780		
Typha	2.01	0.21	120	707		

.

.

<u>alterniflora</u> and <u>S</u>. <u>patens</u>. Contrary to Gerry (1950), they believe the larvae can survive in water for a long time. Olkowski (1966) found most larvae among the <u>S</u>. <u>alterniflora</u> of Delaware marshes with fewer individuals present as ground elevation increased toward the <u>Spartina patens</u> zone. The mature larvae were presumed to be dispersed to higher ground by tidal action.

More recent work gives additional insights into the distributions of tabanid flies. Dukes et al. (1974a) working in Carteret County, North Carolina, and Meany et al. (1976) working on Great Sippewissett Marsh, Massachusetts, found Tabanus nigrovittatus and Chrysops fuliginosus larvae distributed throughout the lower salt marshes occurring in highest densities in areas below mean high tides and dominated by Spartina alterniflora. The preference for low marsh was most marked for C. fuliginosus (Tables 8 and 9). The less abundant Chrysops atlanticus were also characteristic of low marsh areas. The abundance of larvae decreased progressively toward slightly higher elevations with concurrent decrease in Spartina alterniflora and increase in the irregularity of flooding. The authors reported areas of S. cynosuroides, D. spicata and J. roemarianus yielded decreasing numbers of larvae, in that order. However, examination of Table 8 would suggest the order of D. spicata and J. roemarianus be reversed. Dukes et al. did point out that larvae can be found in the high Juncus marsh, particularly in interspersed patches of other vegetation, which is indicative of more frequent flooding. In mixed vegetation, as the proportion Table 8. Recovery of Tabanid larvae from soil in various plant associations in four salt marshes in Carteret County, North Carolina, 1972. (Part of Table 3 from Dukes, et al., 1974, by permission of the editor of Environmental Entomology.)

$ \begin{array}{c c c c c c c c c c c c c c c c c c c $		Samples		Larvae		
$\begin{array}{c c c c c c c c c c c c c c c c c c c $					-	
$\begin{array}{c c c c c c c c c c c c c c c c c c c $			Newport	:/River		
$\begin{array}{c ccccccccccccccccccccccccccccccccccc$	Spartina alterniflora	179			2.95	
$\begin{array}{c c c c c c c c c c c c c c c c c c c $		27	25.9	12		
S. cynosuroides 13 53.8 .17 1.31 S. alterniflora & D. spicata 15 46.7 14 .93 J. roemarianus & D. spicata 5 40.0 7 1.40 Totals and overall average 286 44.9 600 1.28 Born times to emarianus 11 36.4 6 .54 Spartina alterniflora & D. spicata 5 40.0 3 .60 S. alterniflora & D. spicata 5 40.0 3 .60 S. alterniflora & J. roemarianus 5 40.0 3 .60 S. alterniflora & J. roemarianus 5 40.0 3 .60 S. alterniflora & J. roemarianus 5 40.0 2 .40 J. roemarianus 16 50.0 14 .87 None 3 0 2 0 0 Juncus roemarianus 16 50.0 14 .87 S. alterniflora & J. roemarianus 4 50.0 6 1.50 Totals and overall average 20 Dovis Peninsula 37 16.0 <t< td=""><td></td><td>47</td><td>26.1</td><td>29</td><td></td></t<>		47	26.1	29		
S. alterniflora & D. spicata 15 46.7 14 .93 J. roemarianus & D. spicata 5 40.0 7 1.40 Totals and overall average 286 44.9 600 1.28 Hoop Hole Creek Spartina alterniflora 11 36.4 6 .54 Juncus roemarianus 14 7.1 1 .07 S. alterniflora & D. spicata 5 40.0 3 .60 S. alterniflora, D. spicata 5 40.0 3 .60 S. alterniflora, D. spicata & 5 40.0 2 .40 J. roemarianus 5 40.0 2 .40 J. roemarianus 5 40.0 2 .40 J. roemarianus 5 40.0 2 .40 Juncus roemarianus 6 50.0 14 .87 S. alterniflora & J. roemarianus 4 50.0 6 1.50 Juncus roemarianus 10 40.0 5 .50 Totals and overall average 20 50.0 20 1.19		13	53.8	17		
J. roemarianus & D. spicata5 286 40.0 44.9 7 600 1.40 1.28 Totals and overall average286 44.9 600 1.28 600 Spartina alternifiora11 36.4 36.4 6 54 44.9 Juncus roemarianus14 7.1 7.1 1 0.7 S. alternifiora & D. spicata5 40.0 3 3 60 S. alternifiora, D. spicata5 40.0 3 40.0 3 2 None 3 1 roemarianus 5 40.0 2 40.0 None 3 27.2 0 15 0.37 0.37 None 3 20.0 6 1.50 Juncus roemarianus Totals and overall average 16 20.0 0 50.0 Juncus roemarianus Totals and overall average 16 20.0 1.4 61.50 Spartina alternifiora & J. roemarianus Totals and overall average 16 20.0 1.50 1.19 Spartina alternifiora S. alternifiora & D. spicata 37 16.0 16.0 7 1.19 Spartina alternifiora S. alternifiora & D. spicata 4 0 0 0 0 0 S. alternifiora & D. spicata 4 $2.50.0$ 1 50 50 1 Juncus roemarianus 5 250.0 1 1.50 1.50 Juncus roemarianus 5 16.0 7 1.50 1.50 Juncus roemarianus 5 250.0 1 1.50 1.50 Juncus roemarianus 5 14 16.0 7.50 1.50 <t< td=""><td></td><td>15</td><td>46.7</td><td>14</td><td>.93</td></t<>		15	46.7	14	.93	
Totals and overall average 286 44.9 600 1.28 Hoop Hole CreekSpartina alterniflora 5 11 36.4 6 $.54$ 5 11 1 $.07$ 5 $alterniflora$ & D. spicata 5 40.0 3 $.60$ 5 $alterniflora$, D. spicata 5 40.0 3 $.60$ 5 $alterniflora$, D. spicata & 5 40.0 3 $.60$ 5 $alterniflora$, D. spicata & 5 40.0 2 $.40$ J . roemarianus 5 40.0 2 $.40$ J . roemarianus 5 40.0 2 $.40$ J . roemarianus 7.2 15 0.37 None 3 0 0 0 $Totals$ and overall average 16 50.0 14 $.87$ 5 . alterniflora & J. roemarianus 4 50.0 6 1.50 $Totals$ and overall average 20 50.0 20 1.19 $Totals$ and overall average 20 50.0 20 1.19 5 . alterniflora 6 16.7 1 $.17$ $Juncus$ roemarianus 37 16.0 7 $.19$ 5 . alterniflora & D. spicata 4 0 0 0 5 . alterniflora & D. spicata 4 25.0 1 $.50$ J . roemarianus J . roemarianus 2 50.0 1 $.50$ J . roemarianus J . spica			40.0	7		
$\begin{array}{c c c c c c c c c c c c c c c c c c c $		286	44.9	600		
Juncus roemarianus147.11.07S. alterniflora & D. spicata540.03.60S. alterniflora & J. roemarianus540.03.60S. alterniflora, D. spicata &540.02.40J. roemarianus002.40J. roemarianus7.2150.37None3000Totals and overall average4327.2150.37Juncus roemarianus1650.014.87S. alterniflora & J. roemarianus450.061.50Totals and overall average2050.0201.19Davis Peninsula5500201.19Spartina alterniflora1040.05.50Distichlis spicata616.71.17Juncus roemarianus3716.07.19S. alterniflora & D. spicata4000S. alterniflora & D. spicata428.67.50J. roemarianus & D. spicata1428.67.50S. alterniflora, D. spicata &425.01.25J. roemarianus5425.01.25J. roemarianus160001S. alterniflora, D. spicata425.01.25J. roemarianus160001None160000			Hoop Hol	le Creek		
S. alterniflora & D. spicata540.03.60S. alterniflora & J. roemarianus540.03.60S. alterniflora, D. spicata &540.02.40J. roemarianus540.02.40J. roemarianus77.2150.37None $\frac{3}{43}$ $\frac{0}{27.2}$ $\frac{0}{15}$ 0.37Totals and overall average43 27.2 150.37Juncus roemarianus16 50.0 14.87S. alterniflora & J. roemarianus4 50.0 61.50Totals and overall average20 50.0 201.19Davis PeninsulaSpartina alterniflora10 40.0 5.50Distichlis spicata616.71.17Juncus roemarianus3716.07.19S. alterniflora & D. spicata4000S. alterniflora & D. spicata1428.67.50J. roemarianus & D. spicata1428.67.50J. roemarianus20.01.25J. roemarianus16001.25None160000	Spartina alterniflora	11	36.4	6	.54	
$\overline{S. alterniflora \& J. roemarianus}540.03.60\overline{S. alterniflora, D. spicata \&}540.02.40J. roemarianus\overline{S. alterniflora, D. spicata \&}540.02.40None30000Totals and overall average4327.2150.37Juncus roemarianus1650.014.87\overline{S. alterniflora \& J. roemarianus}450.061.50Totals and overall average2050.0201.19Davis PeninsulaDavis PeninsulaDavis PeninsulaSpartina alterniflora1040.05.50Distichlis spicata616.71.17Juncus roemarianus3716.07.19S. alterniflora \& D. spicata4000S. alterniflora \& D. spicata1428.67.50J. roemarianus \& D. spicata \&425.01.25S. alterniflora, D. spicata \&425.01.25S. alterniflora, D. spicata \&425.01.25S. normalianusNone16000$	Juncus roemarianus	14	7.1	1	.07	
S. alterniflora & J. roemarianus540.03.60S. alterniflora, D. spicata &540.02.40J. roemarianus3000None3000Totals and overall average4327.2150.37Juncus roemarianus1650.014.87S. alterniflora & J. roemarianus450.061.50Totals and overall average2050.0201.19Davis PeninsulaSpartina alterniflora1040.05.50Distichlis spicata616.71.17Juncus roemarianus3716.07.19S. alterniflora & D. spicata4000S. alterniflora & D. spicata425.01.50J. roemarianus & D. spicata1428.67.50J. roemarianus & D. spicata1425.01.25J. roemarianus & D. spicata1425.01.25J. roemarianus & D. spicata &425.01.25J. roemarianus160000J. roemarianus160000J. roemarianus160000None160000	<u>S. alterniflora & D. spicata</u>		40.0	3	.60	
J. roemarianusA. roemarianusA. roemarianusNone $\frac{3}{43}$ $\frac{0}{27.2}$ $\frac{0}{15}$ $\frac{0}{0.37}$ North River $\frac{1}{50.0}$ $\frac{14}{50.0}$ $\frac{87}{50.0}$ $\frac{14}{50.0}$ $\frac{87}{50.0}$ Juncus roemarianus $\frac{4}{50.0}$ $\frac{50.0}{20}$ $\frac{6}{1.50}$ $\frac{1.50}{1.19}$ Totals and overall average 20 50.0 20 1.19 Davis PeninsulaSpartina alterniflora 10 40.0 5 $.50$ Distichlis spicata 6 16.7 1 $.17$ Juncus roemarianus 37 16.0 7 $.19$ S. alterniflora & D. spicata 4 0 0 0 S. alterniflora & J. roemarianus 2 50.0 1 $.50$ J. roemarianus & D. spicata 14 28.6 7 $.50$ J. roemarianus & D. spicata 4 25.0 1 $.25$ J. roemarianus 5 $.6$ 7 $.50$ None 16 0 0 0			40.0	3	.60	
None $3 \\ 10 \\ 10 \\ 10 \\ 10 \\ 10 \\ 10 \\ 10 \\ 1$	<u>S. alterniflora, D. spicata &</u>	5	40.0	2	.40	
Totals and overall average 43 27.2 15 0.37 Juncus roemarianus16 50.0 14.87S. alterniflora & J. roemarianus4 50.0 6 1.50 Totals and overall average20 50.0 20 1.19 Davis PeninsulaSpartina alterniflora10 40.0 5.50Distichlis spicata6 16.7 1.17Juncus roemarianus37 16.0 7.19S. alterniflora & D. spicata4000S. alterniflora & J. roemarianus2 50.0 1.50J. roemarianus & D. spicata14 28.6 7.50S. alterniflora, D. spicata &425.01.25J. roemarianus16000	J. roemarianus					
Juncus roemarianus16 50.0 14.87S. alterniflora & J. roemarianus4 50.0 6 1.50 Totals and overall average20 50.0 20 1.19 Davis PeninsulaSpartina alternifloraSpartina alterniflora10 40.0 5.50Distichlis spicata616.71.17Juncus roemarianus3716.07.19S. alterniflora & D. spicata4000S. alterniflora & J. roemarianus2 50.0 1.50J. roemarianus & D. spicata14 28.6 7.50S. alterniflora, D. spicata &4 25.0 1.25J. roemarianus16000	None	3	0		0	
Juncus roemarianus16 50.0 14.87S. alterniflora & J. roemarianus4 50.0 6 1.50 Totals and overall average20 50.0 20 1.19 Davis PeninsulaSpartina alternifloraSpartina alterniflora10 40.0 5.50Distichlis spicata616.71.17Juncus roemarianus3716.07.19S. alterniflora & D. spicata4000S. alterniflora & J. roemarianus2 50.0 1.50J. roemarianus & D. spicata14 28.6 7.50S. alterniflora, D. spicata &425.01.25J. roemarianus16000	Totals and overall average	43	27.2	15	0.37	
S. alterniflora & J. roemarianus Totals and overall average4 2050.0 50.06 201.50 1.19Davis PeninsulaSpartina alterniflora Distichlis spicata10 640.0 16.75 1.50 17Juncus roemarianus S. alterniflora & D. spicata37 16.016.0 7.19 19 .50S. alterniflora & J. roemarianus J. roemarianus & D. spicata14 4 28.628.6 7.50 .50 .50J. roemarianus J. roemarianus None16 160 00 00 0				River		
Spartina alterniflora1040.05.50Distichlis spicata616.71.17Juncus roemarianus3716.07.19S. alterniflora & D. spicata4000S. alterniflora & J. roemarianus250.01.50J. roemarianus & D. spicata1428.67.50S. alterniflora, D. spicata401.50J. roemarianus & D. spicata1428.67.50J. roemarianus250.01.25J. roemarianus16000		16		14		
Spartina alterniflora1040.05.50Distichlis spicata616.71.17Juncus roemarianus3716.07.19S. alterniflora & D. spicata4000S. alterniflora & J. roemarianus250.01.50J. roemarianus & D. spicata1428.67.50J. roemarianus & D. spicata &4000S. alterniflora, D. spicata1428.67.50J. roemarianus250.01.25J. roemarianus425.01.25J. roemarianus16000	<u>S. alterniflora & J. roemarianus</u>	4	50.0		1.50	
Spartina alterniflora1040.05.50Distichlis spicata616.71.17Juncus roemarianus3716.07.19S. alterniflora & D. spicata4000S. alterniflora & J. roemarianus250.01.50J. roemarianus & D. spicata1428.67.50J. roemarianus & D. spicata &425.01.25J. roemarianus7.16000	Totals and overall average	20	50.0	20	1.19	
Distichlis spicata616.71.17Juncus roemarianus3716.07.19S. alterniflora & D. spicata4000S. alterniflora & J. roemarianus250.01.50J. roemarianus & D. spicata1428.67.50S. alterniflora, D. spicata425.01.25J. roemarianus16000				eninsula		
Juncus roemarianus3716.07.19S. alterniflora & D. spicata4000S. alterniflora & J. roemarianus250.01.50J. roemarianus & D. spicata1428.67.50S. alterniflora, D. spicata &425.01.25J. roemarianus16000				5	.50	
S. alterniflora & D. spicata4000S. alterniflora & J. roemarianus250.01.50J. roemarianus & D. spicata1428.67.50S. alterniflora, D. spicata &425.01.25J. roemarianus16000				1		
S. alterniflora & J. roemarianus 2 50.0 1 .50 J. roemarianus & D. spicata 14 28.6 7 .50 S. alterniflora, D. spicata & 4 25.0 1 .25 J. roemarianus 16 0 0 0	Juncus roemarianus		16.0	7	.19	
J. roemarianus & D. spicata1428.67.50S. alterniflora, D. spicata &425.01.25J. roemarianus16000				0		
S. alterniflora, D. spicata &425.01.25J. roemarianusNone16000				. — .		
J. roemarianus None 16 0 0				7		
None <u>16 0 0</u>		4	25.0	· 1 .	.25	
	J. roemarianus					
Totals and overall average9322.0220.26		16	0	0	0	
	Totals and overall average	93	22.0	22	0.26	

ω

MEAN NUMBERS OF TABANID LARVAE/M² FOR THE 1972-1973 AND 1974 SAMPLING PERIODS FOR THE GREAT SIPPEWISSETT MARSH, MASSACHUSETTS, ONLY THE NUMBERS FOR THE CONTROL AREAS ARE SHOWN, EXCLUDING THE RESULTS FROM FERTILIZATION TEST PLOTS. C. ATLANTICUS WAS RECORDED IN LOW NUMBERS FROM THE TEST PLOTS BUT NOT RECORDED FROM THE CONTROL PLOTS. (Modified from Table 1, Meany, <u>et al.</u>, 1976, by permission of the editor, Journal of Applied Ecology.)

	19	72-73	1974			
	Low Marsh	High Marsh	Low Marsh	High Marsh		
Tabanus nigrovittatus	28.4 ± 4.3	22.3 + 4.6	37.7 + 8.3	12.5 <u>+</u> 2.5		
T. lineola	5.7 + 2.7	2.4 + 1.2	1.2 + 0.9	1.2 + 1.2		
Chrysops fuliginosus	28.5 + 4.5	3.8 + 1.5	35.6 <u>+</u> 7.7	0		
C. atlanticus	0	0	0	0		

ິ ເກ of <u>S. alterniflora</u> decreased, the numbers of tabanid larvae also decreased. Sample sites in two of the marshes where there was no vegetation yielded no Tabanid larvae. This concurs with Bailey's (1948) observation that <u>Tabanus nigrovittatus</u> larvae are rarely collected where the marsh peat is devoid of vegetation.

In another paper (Dukes et al., 1974b) the larvae of the Sheep fly Chrysops fuliginosus (82%) and the Greenhead Tabanus nigrovittatus (18%) were found widely distributed throughout a Newport River, North Carolina marsh dominated by Spartina alterniflora. A natural drainage ditch had no apparent influence on distributions. The results were not always consistent with earlier work. Sampling methods have varied and Dukes et al. believe this to be an important factor. However, they propose surface topography to be equally important. The marsh studied was regularly inundated by the tide with no opportunity to dry out. They consider this contrary to many northern marshes dominated by Spartina patens and the entire Newport River marsh was portrayed as being equivalent to the ditch margins of the northern S. patens marshes.

Various authors have reported ditching increases available breeding areas for the Tabanidae in a tidal marsh (Bailey, 1948; Hansens, 1949). Walls and Doane (1960) found the greatest numbers of punkies (Ceratopogonidae) along the edges of the bays and drainage ditches where tall <u>S</u>. <u>alterniflora</u> was the dominant cover or from the moist mud where other vegetation dominated but not from dry or hard soil. Rockel (1969) described the changes

in marsh physiography resulting from ditching New Jersey marshes and the impact on intertidal organisms. Following ditching there was almost a one foot depression on the banks bordering the ditches. With such a soil contour, flooding tides normally covered the marsh to a distance of seven yards from the ditches with plants and animals submerged longer and more frequently than at higher levels. Rockel reported soil drainage to be more complete close to the ditches with the percent of soil water increasing at higher elevations. Soil organic matter and surface thatch increased with elevation, the latter tending to reduce evaporation and soil aeration. The salt water table was closer to the soil surface at higher elevations during ebb tides. There was increasing cordgrass density but decreasing plant height with an increase in elevation. Both Greenhead flies, Tabanus nigrovittatus, and Deer flies, Chrysops spp., exhibited population peaks in marsh areas which were about 0.2 foot below mean high water (MHW). Ditching, by producing more surface area below MHW, provided more marsh offering optimum conditions for larvae of the various biting flies. The impact of ditching was emphasized by Rockel (1969) who stated that 42% of a marsh surface is altered with ditches spaced at 100 foot intervals.

Vertical distributions become evident from the literature. Several authors (Bailey, 1948; Hansens, 1949; Wall and Jamnback, 1957) report <u>Tabanus</u> in the upper strata of marsh soil, usually the topmost one and one-half inches. There are apparent species differences in population concentrations. Rockel and Hansens

(1970a) found populations of Tabanus nigrovittatus and Chrysops spp. were highest below mean high water level on gently sloping banks where the cordgrass was about two feet tall. In their paper of 1970b the Deerfly Chrysops fuliginosus emerged primarily well below mean high water whereas T. nigrovittatus emerged from elevations only slightly below MHW. Rockel and Hansen (1970a) found the Striped horsefly Tabanus lineola and the T. nigrovittatus variant known as "T sp 3" (and tentatively identified in Freeman and Hansens, 1972, as T. n. simulans) mostly at higher elevations where short marsh grass grows. In contrast, Meany et al. (1976) found T. lineola appeared equally between low and high marsh samples. Freeman and Hansens (1972) found the larvae of T. nigrovittatus and T. lineola in more open marsh areas of short grass with T. lineola more abundant in the wetter locations. A greater variance was reported with their finding "T sp 3" more prevalent in the tall grass areas bordering ditches.

Rockel and Hansen (1970a) suggest <u>Tabanus</u> females oviposit primarily on vegetation of a certain height which would influence larval distributions. Emergence appeared to be synchronized with the tide because the lowest areas that support vegetation were exposed long enough to allow hardening of the exoskeleton to permit adults to fly or to climb stems before reflooding (1970b). They found no evidence of larval migrations (1970a). This is in sharp contrast to Gerry's (1950) statement where he indicated the flies originated in the open creek and migrated to thatch piles in the higher areas near the marsh edge. Gerry

considered the reported similarity of prevalence of both mosquito and biting fly concentrations in the upper marsh which is less frequently flooded can be an asset in terms of control of both flies and mosquitos. Since then, Dukes et al. (1974a) believe any control measures directed against the larval stages of tabanids would be inadvisable because such measures would have to be applied to extensive areas with possible adverse impacts on the marsh ecosystem.

Insects in general display an association between the extent of tidal flooding and vegetation zonation. Tide-elevation influences, primarily the length of the hydroperiod, determined the distribution of insect dominants at more stations set up within the North Carolina marshes examined by Davis and Gray (1966). The Homoptera were most important and had a more nearly zonal distribution pattern than did species from the other insect orders (Figure 3). Homoptera numbers decreased as other orders increased with the transition from low marsh to high marsh elevations.

Evidence for zonation among the Homoptera has been enhanced with the recent work of Foster and Treherne (1975, 1976). They found the intertidal aphid, <u>Pemphigus trehernei</u> restricted to the roots of the aster, <u>Aster tripolium</u>, growing in low marshes near the edges of creeks and salt pans in the marshes of Scolt Head Island, Norfolk. Gray (1971, <u>in</u> Foster and Treherne, 1975) classified salt marsh communities containing <u>Aster tripolium</u> into high, mid, and low marshes. High marshes are almost removed from tidal influence while mid marshes are reached by

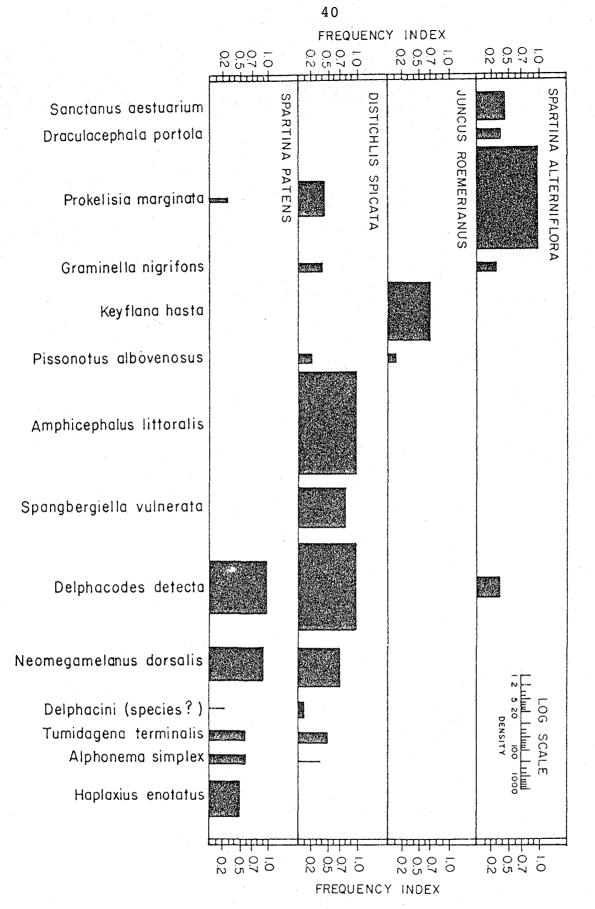


Figure 3. Frequency-density diagram of the principal species of Homoptera from the herbacious strata of four zones of salt marsh vegetation. (From Davis & Gray, 1966, by permission of the editor of Ecol. Monogr.)

all high spring tides and can be distinguished from low marshes by the presence of <u>Plantago</u> <u>maritima</u>. Low marshes are inundated by high neap and spring tides.

Foster and Treherne (1975) reported a relatively high percent of air space near the creek edges where these root aphids were most abundant (Figure 4). Only small aphid populations were found in soils with less than 10 percent air space (Figure 5). Adult aphids were about one millimeter in diameter and were associated with cavities of at least that size. These cavities were categorized and arranged in decreasing order of importance:

- (1) Cavities formed by cracking from drying and erosion
- (2) Cavities formed by decaying roots
- (3) Cavities formed by burrowing of other animals

(4) Cavities formed around mollusc shells and pebbles.Cavities (1) and (3) were usually confined to edge regions while cavities (2) and (4) were found in all marsh areas and were generally water filled.

First instar aphids penetrated the edge soils of the low marshes considerably earlier than the non-edge areas. In eight trials with 12 aphids at each site, the mean number of aphids floating \pm S.E. was 0.88 \pm 0.23 (7.3%) at the edge sites and 6.75 \pm 0.59 (56.3%) at the non-edge sites, a highly significant difference ($x^2 = 47.8$, P < 0.001).

Foster and Treherne found that aphid populations in mid-marsh areas do not show an equivalent edge effect. Such sites did not support high aphid concentrations, the maximum

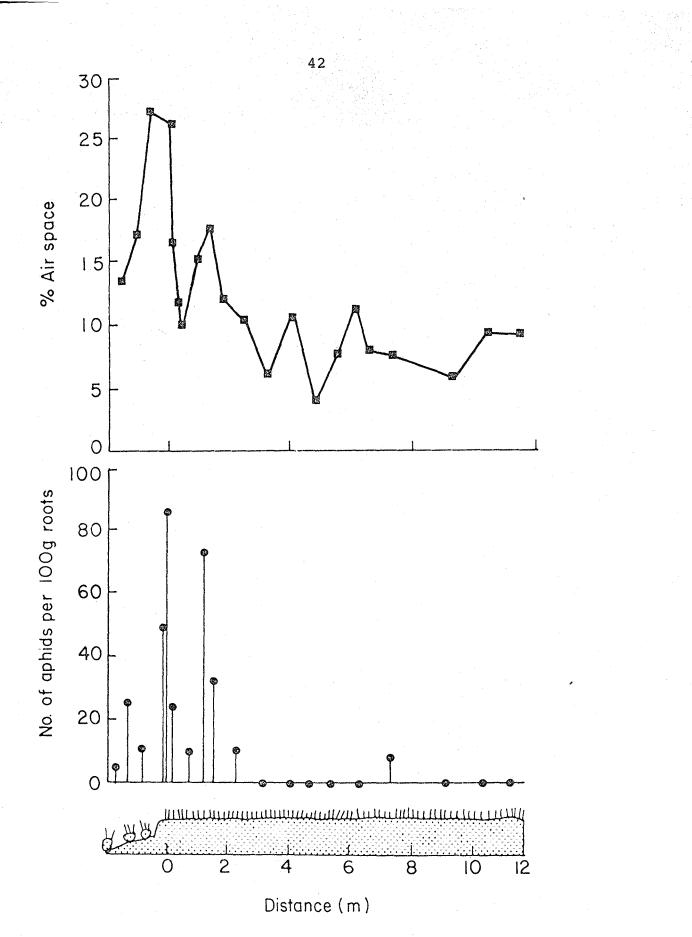


Figure 4. Aphid density and percent soil air-space along a transect on Missel Marsh East, May 26, 1971. All measurements are of single samples. Percent air space was measured by field method. (From Foster & Treherne, 1975, by permission of the editor, Oecologia.)

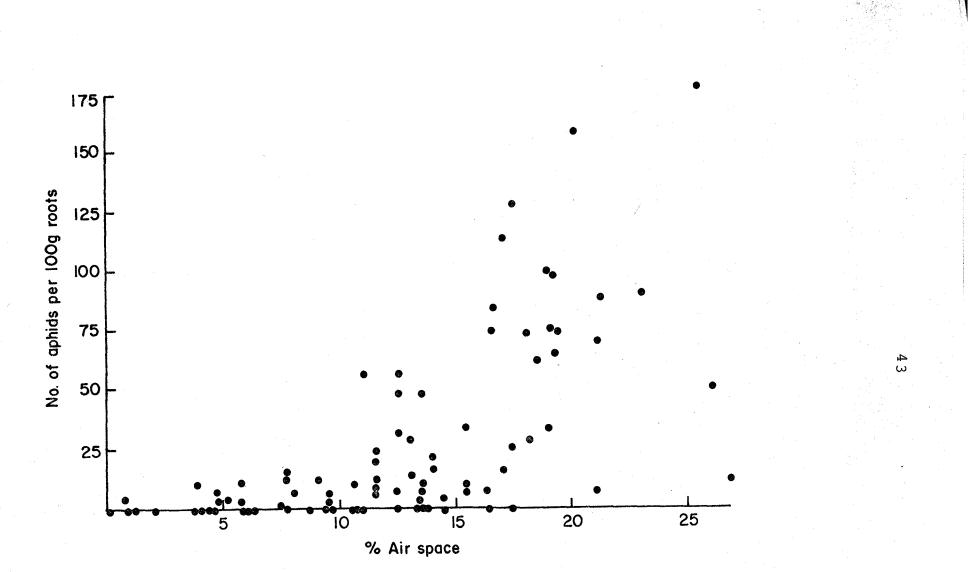


Figure 5. Relationship between aphid density and percent air space in the soil. Measurements made in various sites on Scolt Head in 1970 and 1971. Percent air space measured by field method. Aphid density expressed as numbers per 100g fresh weight of aster roots. (From Foster and Treherne, 1975, by permission of the editor, Oecologia.)

1.17

number per core was 20. The mid-marsh soils are more mature, have a higher organic content, have smaller air spaces, and the large cracks associated with the clay soils do not appear in mid-marsh sites. The small pore size probably limits the aphid populations in the mid-marsh.

Foster and Treherne (1975) suggest that for a salt marsh aphid population to maintain itself, there must be appropriate spaces within the soil physically to accommodate these nonburrowing insects, and within such spaces conditions must be suitable for survival. While there is an enormous potential for parasitic exploitation of <u>Aster tripolium</u> by the aphid only a small portion of the plant's distribution is available. The edge regions appear more accessible to dispersing aphids along the edges. The greater interplant spacing and larger bare areas may enhance aphid landings along the edges in these low marshes. In addition, the greater number of cavities along the edges would be attractive to these animals.

Other works report on intertidal insects that are regularly inundated by the flooding tide. Larsen (1951) reported marsh beetle distributions were strongly influenced by soil type and vegetation, changing as vegetation invaded bare areas. Evans et al. (1971) described their observations of four species of beetles (<u>Bledius spectabilis</u>, <u>Heterocerus fenestratus</u>, <u>Dichirotrichus pubescens</u> and <u>Cillenus lateralis</u>) which were confined effectively to the upper regions of the banks of the drainage channels of the Scolt Head Island, Norfolk salt marshes, in that region between the anaerobic mud of the lower

bank which was usually covered by a film of algae and the dense marsh vegetation, mainly <u>Halimione portulacoides</u> (Figure 6). There were no apparent interspecific differences in beetle distributions within this zone. The beetle area was usually well drained and identified by the castings produced during burrowing. This area differed considerably from the damper muds back from the drainage channels (Figure 7). The area is covered intermittently by high tides to a depth of one meter for seven days in every fortnight. Submergence was reported to last about 3-4 hours at the highest tides.

Evans et al. (1971) portray <u>B</u>. <u>spectabilis</u> burrows, and presumably <u>Cillenus lateralis</u>, being excavated to a depth of 2-5 cm but not penetrating beyond an average depth of 9 cm (Figure 8). <u>H</u>. <u>fenestratus</u> burrowed to a depth of only 1-2 cm, while <u>Dichirotrichus pubescens</u> did not burrow but lived in surface cracks.

Submergence responses were found to induce an apparent anoxic condition (Figure 9). There was a linear relationship between submergence and recovery for <u>H</u>. <u>fenestratus</u> followed by a more rapid recovery (Figure 10). The authors suggest the use of physiological responses to submergence imply the subterranean mode of life largely prevents direct exposure of the beetles to sea water in natural conditions.

A significant proportion of <u>H</u>. <u>fenestratus</u> was found to be vulnerable to flooding and this can be related to the shallow burrows this beetle excavates. This in turn can be related to the mode of respiration in which the delayed onset of deep

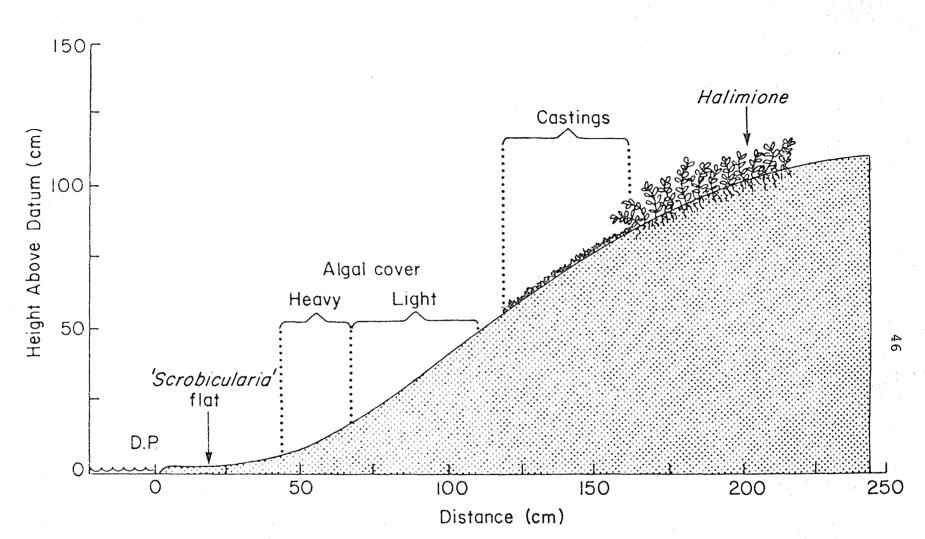


Figure 6. Section of the Hut Drain bank, showing region inhabited by <u>Bledius spectabilis</u>, <u>Heterocerus fenestratus</u>, and <u>Cillenus lateralis</u> (shown by castings). 'D.P.' <u>marks water level of lowest high tide</u>, <u>August/September 1968</u>. (From Evans et al., 1971, by permission of the editor, J. mar. biol. Assoc. U.K.)

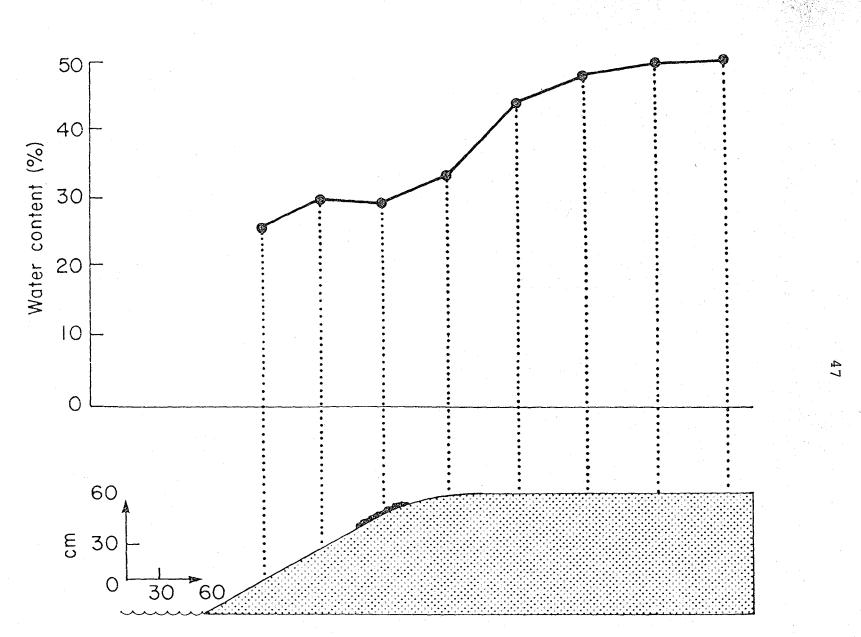


Figure 7. Water content in the top 10 cm of mud in a section of Hut Drain bank. The beetle area is represented by casting. (From Evans et al., 1971, Figure 4, by permission of the editor, J. mar. biol. Assoc. U.K.)

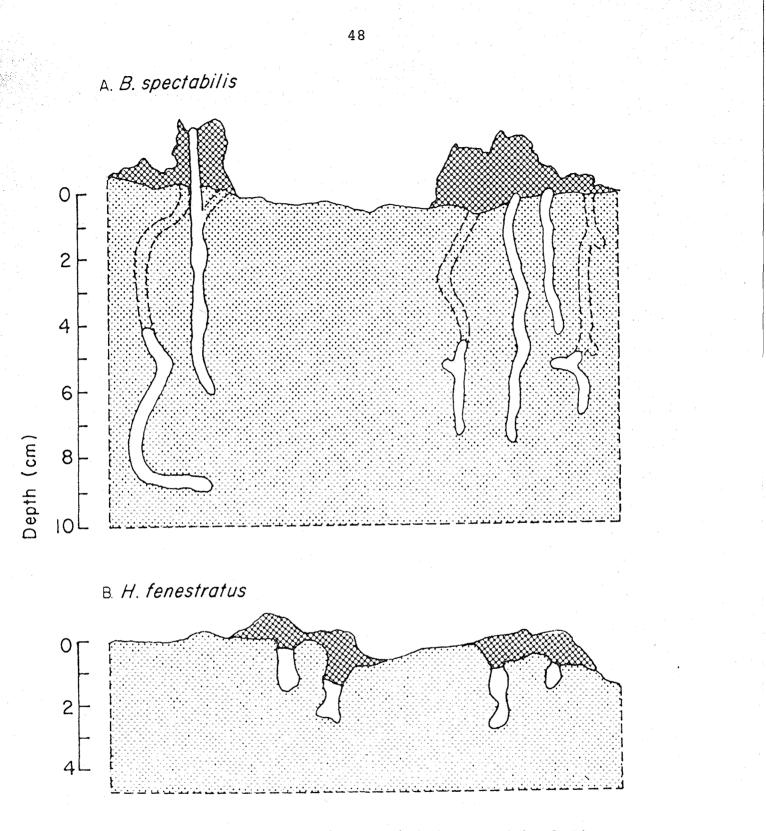
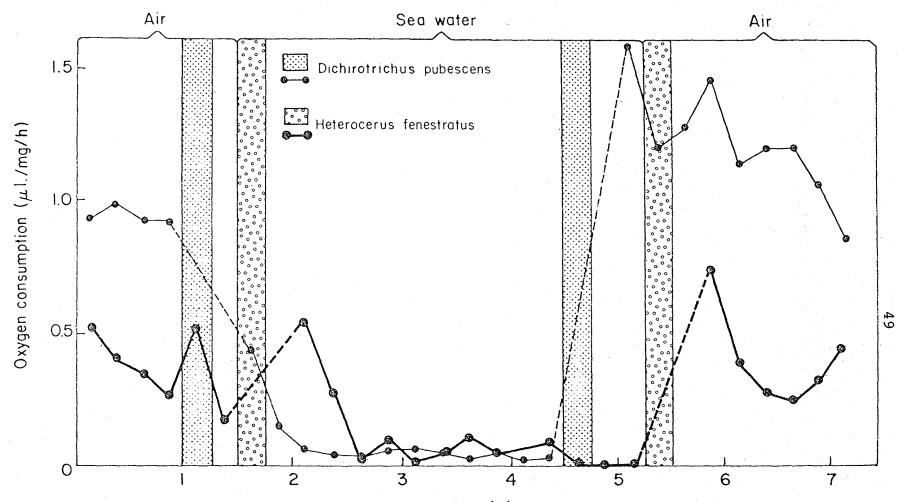


Figure 8. Diagram of burrowing activities of (a) <u>Bledius</u> <u>spectabilis</u> and (b) <u>Heterocerus fenestratus</u>, following introduction of six specimens of each species into 'soil cells.' The solid lines represent open burrows and the broken ones regions of the burrow system previously excavated and subsequently refilled. (From Evans et al., 1971, Figure 6, by permission of the editor, J. mar. biol. Assoc. U.K.)



Time (h)

Figure 9. The effect of submergence in sea water on the oxygen consumption of five adult <u>Dichirotrichus pubescens</u> and thirty <u>Heterocerus fenestratus</u>. The shaded area indicates the equilibrium period allowed on introduction and removal of sea water from the manometer flasks. (Modified from Figs. 12 and 13 in Evans et al., 1971, by permission of the editor, J. mar. biol. Assoc. U.K.)

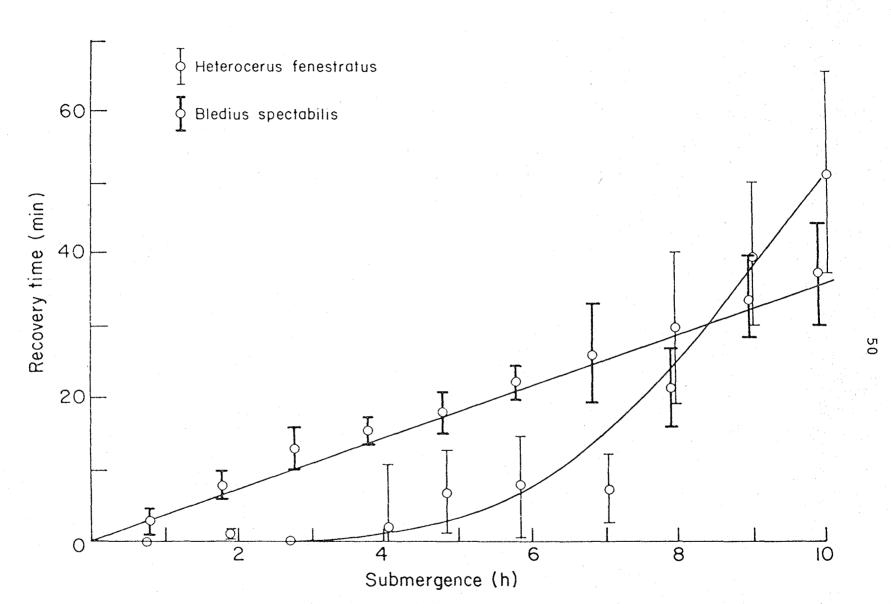


Figure 10. Recovery time for <u>Bledius spectabilis</u> and <u>Heterocerus fenestratus</u> from anoxia induced by varying periods of experimental submergence in sea water. (Modified from Figs. 9 and 11, Evans et al., 1971, by permission of the editor, J. mar. biol. Assoc. U.K.)

anoxia during the first four hours of submergence for H. fenestratus would ensure minimal recovery periods. This is in contrast to the other three species that had a greater and earlier decline in oxygen consumption during submergence but a quicker initial response in recovery (Figures 9 and 10). Apparently, the behavioral and burrowing activities permit this group of insects to survive and flourish in an area of the salt marsh that is regularly inundated by the tides and further work (Treherne and Foster, 1977) provides additional insights into behavioral patterns and survival of these insects. Dicheirotrichus gustavi is widespread on the soft shores of the intertidal zone in the British Isles. On salt marshes, the adults apparently live in cracks in the mud and do not burrow. During tidal emergence D. gustavi exhibited well defined nocturnal activity. Maximum activity occurred after dusk and decays exponentially until dawn when it abruptly terminated by individual beetles retreating into soil cavities (Figure 11). This circadian activity insured that the majority of the population avoided direct submergence during the first few tidal coverages of the marsh surface of a rising tide sequence. The first tidal coverage, in a sequence of spring tides, occurred shortly before dawn when minimal numbers of adult beetles were exposed to sea water. The second and third critical tides each occurred 12.5 hours later, the second just before dusk while the insects were subterranean (Figure 12). Both field and experimental observations indicated that at least two tidal submergences are required to suppress nocturnal activity on the

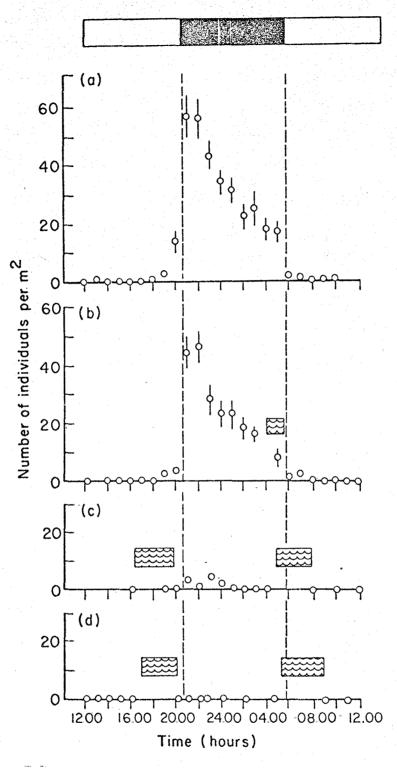


Figure 11. Surface activity of <u>Dicheirotrichus gustavi</u>, in the twenty experimental plots illustrated in Fig. 1, (a) before (31 July 1975) and (b-d) during a period of tidal submergences (horizontal bars) in a sequence of rising tides ((b): 2 Sept. 1975; (c) 3 Sept. 1975; (d) 4 Sept. 1975). The open circles indicate the mean number of individuals on the soil surface and the vertical lines the extent of twice the standard error of the mean. Dawn and dusk are indicated by the broken lines. (From Treherne & Foster, 1977, by permission of the editor of the J. Anim. Ecol.)

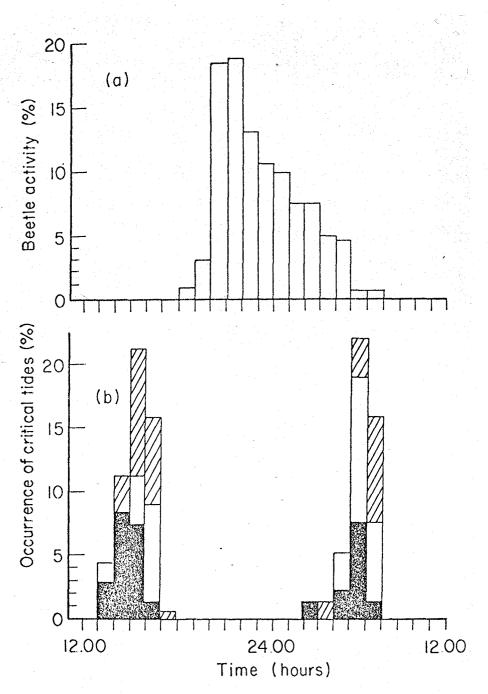


Figure 12. Timing of 'critical' tides (first tides to cover the beetle area after a period of emergence) in relation to beetle activity. It is assumed that tides of 20.7m above the Chart Datum just cover the beetle area (as on 3 Sept. 1975). A first 'critical' tide is any tide exceeding 20.7m which comes after at least three The second and third 'critical' tides follow non-flooding tides. (a) Hourly beetle activity the first at about 12.5 h intervals. Pooled data from (Fig. 2(a), (b), (c)) expressed as in the field. a percentage of the total number of beetles observed during this (b) Hourly occurrence of 'critical' tides in 1975 and period. 1976, expressed as a percentage of the total number of 'critical' First 'critical' tide, solid columns; second 'critical' tides. tide, open columns; third 'critical' tide, hatched columns. (Data from Admiralty Tide Tables.) (From Treherne & Foster, 1977, by permission of the editor of J. Anim. Ecol.)

soil surface. This change in behavior did not seem to be related to changed soil character but from direct experience of changed conditions in the cavities of submerged soil. Reversion to nocturnal surface activity followed an absence of two tidal submergences.

While Evans et al. (1971) found four species effectively confined to the upper intertidal regions of the banks of the Scolt Head marsh drainage channels they found no apparent interspecific differences in distribution within this zone. The mud is aerobic and brown with a quantity of organic material. In contrast, Larsen (1951) described beetle distributions to be very restricted, strongly influenced by soil types and vegetation, replacing each other in successive biotypes. As soon as conditions changed on those Danish marshes at Skallingen, beetle species distributions changed in accordance to their specific demands. As a rule different communities were kept in pure formations as were the plant communities. However, populations intermingled as an area passed through a transition period. Larsen (1951) established three groupings of beetles based on soil and plant associations.

1. Bare algal clay flats. <u>Bledius spectabilis</u> was associated with bare sand with a clay covering and an algal layer covering 20-60%, soil water content 20-25%, abundant in the tidal zone which is covered by water each day. <u>Bledius</u> <u>taurus</u> has much the same substrate requirements but is found with a high soil water content and prefers loose sand outside the tidal zone. Bledius diota and Trogophlous schneideri are

associated with no vegetation, high water content (20%), and high salinity $(30-100^{\circ}/\circ\circ)$.

2. Bare humid sand flats. <u>Bledius arenarius</u> was found on pure sand, no clay, no algal mats or other vegetation with a salinity of $5-20^{\circ}/\circ\circ$ and a water content of 10-20%.

3. Tolerating plant covered soil. <u>Bledius trichornis</u> found mainly in overgrown biotypes and sandy places with varied salinities, 0-60^O/oo. <u>Dichirotrichus pubescens</u> can be found in sandy places but prefers clay and dense vegetation. <u>Heterocerus flexuosus</u> was found in many places except where the sand was pure and bare. It invaded muddy banks with an algal mat and areas of pure sand covered by clay and overgrown with vegetation.

At least three of the beetles described by Evans et al. (1971) are found in Larsen's (1951) categories 1 and 3, and still they are confined to a particular zone in the Scolt Head Marsh with no apparent differences within the zone. The divergence probably can be explained by the response of these beetles and the aphids (Foster and Treherne, 1975, 1976) to submergence.

AVES.

There is a definite relationship between bird distributions in a salt marsh and the interplay between vegetational zonation, tidal flooding and salinity as it affects feeding and reproductive activities. Urner (1935) identified the ava fauna specialized to nest and feed in salt marshes in relation to the marsh wetness, creek and pond depth, salinity and accessibility of the tides. Both he and Ferrigno (1961) called attention to changes in bird populations following any alterations in these various parameters.

The resident Clapper rail, <u>Rallus longirostris waynei</u> of the Georgia salt marshes displayed distinct distributional preferences within the dominant grass, <u>Spartina alterniflora</u> (Oney, 1954). This grass could be divided into three categories. A Tall Grass zone was located on the edge of natural drainage ditches, creeks and rivers with a soft mud substrate and an average plant density of 110/square yard, two to four feet tall. The Short Grass zone was found in the lowest parts of the Georgia marshes, had a high sand content with an average plant density of 362/square yard, five inches to two feet tall.

The square-back marsh crabs, <u>Sesarma</u>, the primary food of the rail, were most abundant in the Tall Grass zone. The fiddler crabs, <u>Uca</u>, and the periwinkle, <u>Littorina irrorata</u>, of secondary and tertiary food value, were found in the Medium and Short Grass zones, respectively. However, during his three year study, Oney found a definite nesting preference for the Medium Grass zone: 118 nests measured, 30 were in the Tall Grass zone, 87 or 74% were in the Medium Grass zone and only one nest was recorded for the Short Grass zone. Furthermore, his prior studies in 1949 and 1950 indicated the bird's preference for the medium type grass bordering the Tall Grass zone along a small ditch or creek. The average distance from a nest to the creek at low tide was 458 feet with a range of

3-1200 feet. The average nest was 20 feet from a change in vegetation zone with a range of 3-80 feet. Earlier, Stewart (1951) had called attention to the importance of the edge between vegetation zones in placement of Clapper rail nests. Working in the Eastern short marshes of Maryland, he found a high correlation between nest densities and the amount of edge between the tall (and dense) and the short (and sparse) growth form of <u>Spartina alterniflora</u>. Lower correlations were derived from pure stands of short and tall <u>Spartina</u>. Stewart recorded nest density in the best edge at $2.5 \pm 0.3/acre$ with nests within 15 feet of the creeks.

Changes in vegetational cover can affect Clapper rail nesting. Ferrigno (1957) reported a decrease in observed nests on the Coney's area but an increase on the Keye's area, both in Cape May County, New Jersey. He attributed the decrease at Coney's to the development of large barren areas which had been formerly covered with <u>Spartina alterniflora</u>. Possible causes given were either adverse weather conditions or over-population of fiddler crabs killing the vegetation by burrowing activities. Rafting of debris by storm tides may have smothered the vegetation.

Several references call attention to the height of nest placement and the impact of inundation on hatching success of the rails. In his early work on the life histories of North American marsh birds, Bent (1963) located the second nestings of the Clapper rail, <u>Rallus longirostris crepitans</u>, on higher drier ground covered with only a few inches of water at high tide.

He went on to say that most nests were built in small clumps of grass along creek banks in soft wet mud: nest heights varied 8-12 inches above the mud, probably high enough to escape ordinary high tides but not spring tides. King rail nests were usually found in the shallow water portion of the marsh with water depths of 4-24 inches. The nest height above water was dependent on water depth; the shallower the water, the lower the nest (Meanley, 1969). Most nests of the resident Clapper rail of Georgia were 8-9 inches off the ground with the average distance from the ground to egg level at 14.8 inches with extremes of nine to 30 inches. These nests could be covered by as much as 12-19 inches of water and still support a hatch (Oney, 1954). Earlier, Stewart (1952) found 1951 hatching success of the first sets of eggs in the Chincoteague, Virginia marshes to be less than 45 percent compared to 94 percent in 1950. Due to high storm tides during the early part of the nesting season, production of young was greatly staggered. Ferrigno (1957) clarified the impact of storm tide flooding on hatching success (Table 10). On May 10, 1955, there had been a high tide with strong winds but little nest damage since the birds had just begun to nest. High tides of June 8-11, resulted in virtually complete nest destruction. There was subsequent renesting as evidenced by a later hatching There were high lunar tides in 1956 but with no wind. peak. Although 14 percent of the nests were destroyed by a flood tide on June 8, 83 percent of the first nests hatched successfully in that year.

Table 10.

CLAPPER RAIL NEST CENSUS DURING 1955 AND 1956 IN CAPE MAY COUNTY, NEW JERSEY. (From Table 1, Ferrigno, 1957.)

	obse	mber of Number oserved destroyed nests (by tides)		oyeđ	Number hatched successfully		
Area	1955	1956	1955	1956	1955	1956	
CONEY'S	29	20	9	4	20	16	
KEYE'S	_7	10	3	1	_4	9	
TOTAL	36	30	12	5	24	25	

The Clapper rail has been described primarily as a resident of the more saline low marshes (Stewart, 1951; Oney, 1954). In contrast, the King rail, <u>Rallus elegans</u>, and Virginia rail, <u>R. limicola</u>, are essentially inhabitants of the fresh and brackish marshes; numbers varying with vegetation (Bent, 1963; Meanley, 1969). They are common in the coastal marshes of Louisiana, abundant in the South Carolina low country fresh and brackish marshes, especially where the the giant cut grass <u>Zizaniopsis miliacea</u> and the fiddler crab, <u>Uca minax</u>, are present. The grass provides good nesting cover with a nest density as high as one per acre. In the Chesapeake Bay area, the King rail is most abundant where the big cord grass, <u>Spartina cynosuroides</u> is dominant, providing good cover.

Both the King and Clapper rails species occur in transition areas, especially in the lower reaches of brackish marshes where interbreeding sometimes produces hybrids. In Delaware they were taken in the Broadway meadows located between Flemings Landing and Woodland Beach. In that section at Taylors Gut where mixed populations occurred, the dominant vegetation was <u>Spartina alterniflora and Scirpus robustus</u> with the salinity range reported at $5.7-7.2^{\circ}/oo$. Inland at Flemings Landing, the dominant vegetation was <u>Spartina patens</u> and <u>Distichlis spicata</u> with a salinity range of $3.7-4.4^{\circ}/oo$ and only the King rail was observed. Only Clapper rails were found at Woodland Beach with <u>S</u>. <u>alterniflora</u> and <u>S</u>. <u>robustus</u> as the dominant vegetation and a salinity range of $7.5-7.6^{\circ}/oo$.

Both rails establish nesting territories but there is no evidence that either species confines its feeding to a territory. Instead, they forage in areas adjacent to or in the preferred nesting areas. The King rail often takes advantage of muskrat runs in its search for food. Both rails search along creek banks where the Clapper rail finds its preferred food the marsh crab <u>Sesarma</u> or the fiddler crab. The King rail frequently repairs to a favorite feeding site, such as the top of a muskrat house, for dismemberment and ingestion of the food item (Oney, 1954; Meanley, 1969).

While both rails can be identified with the low marshes, Bent (1929) described the Eastern willet, Catoptrophorus semipalmatus, as a decidedly coastal bird, seldom seen far from coastal marshes, beaches and islands. He described the nesting area as being on sand islands overgrown with tall grass or on dry uplands close to marshes. Vogt (1938) was a bit more specific in his observations in the Fortesque, New Jersey marshes. He recorded most nests in the dense Spartina patens of the high marsh, near the boundary with Spartina alterniflora or Typha augustifolia. In addition, Vogt found the Willet most abundant where S. patens was regularly mowed or burned, presumably making it easier for the birds to feed. This clamorous bird was also found to be abundant where wintering Brant or Greater Snow geese had grazed S. patens in the absence of eel-grass (Zostera marina). Stewart and Robbins (1958) reported breeding densities of 10.5/100 acres of brackish hay marsh during a 1956 survey of Dorchester County, Maryland. In

contrast to earlier observations these workers described the habitat as a strip along the tidal creek 220 yards wide.

The Willet takes most of its food from within the established territory. Some trips are made to the bayshore or uplands and, when the water level is right, the marsh ponds are visited, the birds picking food from the grass blades. However, the extent of this off-territory feeding is affected by the wind (Vogt, 1938).

The casual visitor to a salt marsh seldom sees the secretive rails as they slip through the vegetation but, during the first half of the summer, a visitor can be announced by the clattering presence of the Willet. However, as the summer progresses these clamorous calls decline to silence although the bird can still be seen flying over the expanses of grass or walking along a creek bank. Black ducks can be put to flight during the early summer nesting season but waterfowl typically are most obvious during the fall migration flights and over the wintering areas.

Thirteen major types of waterfowl habitat have been categorized by Stewart (1962) in the upper Chesapeake Bay area; six in open tidewater areas, five marsh types and two in the coastal plain interior were designated as river bottoms and impoundments. The five marsh habitats were distinguished by salinity distributions which in turn had an influence on vegetational composition. Stewart considered the brackish estuarine bays to be the most important habitat for waterfowl populations as a whole. Fresh estuarine bays, brackish estuarine marshes and estuarine river marshes also attracted large numbers

of waterfowl (Table 11). While Stewart considered the brackish estuarine bays to be the most valuable waterfowl habitat, the marshes bordering such bays had fewer recorded species (10) than the less saline fresh estuarine bay marshes (13) and estuarine river marshes (12). The three marshes with one or two designated principal species had higher numbers of secondary species. It is evident that nine species were restricted to certain marsh habitats while seven were recorded from all five types. The Black duck was the most ubiquitous, being classed as a principal species for all five marsh habitats. Making an arbitrary selection of those seven species found throughout the marshes and using estimated population size data available from Stewart, Table 12 depicts the population distributions for four of the seven species. The widely distributed Black duck seemed to be more associated with the open bays or agricultural lands during migration and wintering. Large breeding populations were found along the shallow brackish estuarine bays and associated marshes. In contrast, while the recorded population was low (Table 12) the great majority (99%) of Green-wing teal were found in the tidal marshes. Stewart indicated the same pattern for the Blue-wing teal. It should be pointed out that waterfowl numbers within a particular type of habitat vary from location to location.

Vegetation dominance and distributions as well as salinity levels had an influence on waterfowl distributions by providing food and nesting sites. <u>Spartina alterniflora</u> was the predominant plant in the coastal embayed salt marshes located back of the

PRINCIPAL (P) AND SECONDARY (S) WATERFOWL SPECIES ASSOCIATED WITH VARIOUS TYPES OF UPPER CHESAPEAKE BAY MARSHES DURING 1958-1959. (Derived from Stewart, 1962.)

	Emb Ma	stal ayed rsh	Salt Brackish Estuarine Estuarin Bay Marsh Bay Mars		arine Marsh	Fresh Estuarine Bay Marsh P S		Estuarine River Marsh P S		
	P	S	P	S	P	S	P	S	<u> </u>	S
CANADA GOOSE	x			х		X	X			X
BLACK DUCK	X		Х		X ,		x		X	
SNOW GOOSE		x								
MALLARD		x		х		х		X	X	
GREEN-WING TEAL		x		Х	Х			х	X	5 A A
BLUE-WING TEAL		x		x	х			х	X	
SHOVELER		X		х		x		x		
AMERICAN WIDGEON		x		X	х			x		x
HOODED MERGANSER		x				Х				X
COMMON MERGANSER								x		X
RING NECK DUCK								X		X
AMERICAN COOT		x						x		X
PINTAIL		х		x		Х		x	x	
GADWALL				х		X		х		
WHISTLING SWAN								x		
WOOD DUCK		1. A.					,		х	
	2	9	1	8	4	6	2	11	6	6
	1	1		9	1	0.	1	.3		2

64

Table 11.

Table 12.

POPULATION DISTRIBUTION OF WATERFOWL ASSOCIATED WITH VARIOUS TYPES OF UPPER CHESAPEAKE BAY MARSHES DURING 1958-1959.

(Derived from Stewart, 1962 --- Tables 44, 45, 76, 84.)

Percent of total population*

Waterfowl	Total Population (X 1000)	Coastal Embayed Marsh 21,000 acres	Salt Estuarine Bay Marsh 113,000 acres	Brackish Estuarine Bay Marsh 47,000 acres	Fresh Estuarine Bay Marsh 30,000 acres	Estuarine River Marsh 67,000 acres	
BLACK DUCK	317	12	10	3	3	5	
	517		± 0	~	,		
CANADA GOOSE							
MALLARD	148	6	1	3	5	7	
GREEN-WING TEAL	10	32	23	13	6	25	
BLUE-WING TEAL					р Долгания Сталана Ст		
AMERICAN WIDGEON	196	8	5		4		
PINTAIL							

*The percentage figures are averaged values derived from five observation periods from October 2, 1958 through March 16, 1959.

The difference between the total percentage figure for each waterfowl species and 100% indicates the birds were observed on the open waters of the adjoining bays and areas other than in the marshes.

ហ

barrier beaches. The Black duck was the only species that was common and widely distributed while the Canada goose was numerous along tidal creeks and guts. Canada and Snow geese were most numerous on extensive cord grass areas or mud flats. Pintails, Shovelers, Blue-wing teal and American widgeon preferred open ponds with poor drainage or stable water levels in artifically created ponds. Scattered pairs of Black ducks along with a few Mallards and Blue-wing teal nested in these marshes. They preferred marsh islands rather than shore zone marshes and all nests found by Stewart (1962) were in the drier more elevated areas.

The salt estuarine bay marshes were characterized by a high salinity and narrow tidal fluctuation. Widgeon grass <u>Ruppia maritima</u>, salt grass <u>Distichlis spicata</u>, salt marsh cord grass <u>Spartina alterniflora</u>, salt meadow cord grass <u>S. patens</u>, the bulrush <u>Scirpus robustus</u> and <u>Juncus roemarianus</u> were the common plant species. The Black duck was the only common and widely distributed waterfowl. All others were scarse.

The brackish estuarine bay marshes comprised a complex mosaic of ponds, creeks and marshes. The principal plant species were <u>Ruppia</u>, <u>Distichlis</u>, <u>Spartina cynosuroides</u>, <u>S. alterniflora</u>, <u>S. patens</u>, <u>Scirpus olneyi</u> and <u>Juncus roemarianus</u> depending on the locale. There was a diverse fauna with raccoons and crows as important waterfowl predators. As in the coastal marshes, the Black ducks and Green-wing teal were generally distributed showing a definite preference for tidal creeks and ponds in drainage systems with marginal mud flats exposed at low tide.

The American widgeon and Gadwells concentranted on stable ponds with beds of widgeon grass or musk grass (<u>Chara sp.</u>). The Hooded mergansers were restricted to the larger tidal creeks while Canada geese preferred the larger ponds. Large numbers bred in these marshes with the Blue-wing teal restricted almost entirely to the marsh meadow (presumably <u>Spartina</u> patens).

The fresh estuarine bay marshes had next to the smallest total recorded acreage but a much greater diversity of vegetation. Along with this vegetation variety, Stewart recorded the greatest numbers of waterfowl species, though eleven of the 13 were classed as secondary (Table 11) and most of the population was located on the bays adjoining these fresh estuarine bay marshes (Table 12) rather than in the marshes themselves. As with some of the other marsh types, the Canada geese were found on the larger ponds while the Black ducks and teals were found in well drained areas of creeks and ponds with exposed mud flats.

The vegetation varied with salinity in the estuarine river marshes of the upper Chesapeake. This was an area with a great variety of emergent vegetation and possessed a greater tidal fluctuation than some other marsh types. The greatest concentrations of waterfowl, including many dabbling ducks, were located here between the fresh and brackish water habitats where excellent cover and food was available.

An examination of the numerous tables in Stewart (1962) depicting waterfowl food habits in the upper portion of Chesapeake Bay discloses a relationship between the horizontal salinity-

vegetational distribution and food habits. A greater variety of food species was consumed as salinities decreased from one marsh type to another. Presumably this was a reflection of increased diversity and availability associated with less saline areas. However, there may be more than availability playing a part in the food habits of waterfowl associated with tidal marshes. By using two criteria, timing of usage and relative quantity of food per unit area remaining after usage, Owen (1971) ascertained the winter feeding habits of White-fronted geese in England to be highly selective in the choice of vegetation zone. By both criteria, the Agrostis zone which is the lowest grassy area and dominated by Agrostis stolonifera was preferred to all others. Lolium perenni, Festuca rubra, Hordeum sicalinum and Juncus gerardi were found at higher elevations, and were consecutively grazed later in the season. Selection of a zone was not absolute however; when large numbers of geese were present all zones were occupied. Owen presented evidence that plants may be selected on basis of protein, fiber and carbohydrate content. If so, Puccinellia maritima was considered to be preferred by White-fronted geese over Agrostris and Festuca.

Among the passerine species the magnitude of flooding influences habitat selection for both the Coastal Savannah sparrow <u>Passerculus sandwichensis alaudinus</u> and the Eastern Sharp-tailed sparrow <u>Ammospiza c. caudacuta</u>. Both are found at the higher marsh elevations. The former is found in the <u>Salicornia</u> associes, less so among the <u>Distichlis spicata</u>. The Sharp-tailed sparrow is found in the drier Spartina patens zone with some

nests on the ground but most of them are at varying heights above ground level, presumably above summer high water levels (Bent et al., 1968).

Adequate feeding grounds and suitable nesting cover within easy flight of the feeding grounds provided two components of a suitable habitat for both the Northern Seaside sparrow, Ammospiza m. maritima (Bent et al., 1968) and Macgillvrays Seaside sparrow, Ammospiza maritima macgillivraii (Tomkins, The latter is the breeding Seaside sparrow of South 1941). Carolina and Georgia and in this locale these two requirements are often separated by a short distance. Food requirements were far stronger than nesting needs in determining habitat limitations with population shifts caused by erosion of feeding sites. Feeding grounds were the wet banks of salt creeks where Spartina alterniflora grows rankest, the ponds that are at the heads of salt creeks and the patches of S. alterniflora located on the outer beaches that were flooded by tides. Tidal flooding of the feeding grounds made them unsuitable for nesting sites. Nests were built in a variety of places; a few inches above the mud in Sporobolus-Paspalum to three feet in Spartina or Juncus and up to five feet in Baccharis. Nesting preferences appeared to follow this same order with the taller shrubbery the least desirable site.

Much attention has been focused on the Song sparrows <u>Melospiza melodia</u> of the salt marshes of the San Francisco Bay region (Marshall, 1948; Johnston, 1956a, 1956b). These investigators discussed habitat, abundance, annual cycles,

population structure and maintenance, and concluded that tidal flooding and vegetational zonation were the determining influences. Song sparrow distributions were more circumscribed than vegetation distributions as birds are less tolerant of drying (Marshall, 1948). <u>Scirpus acutus</u> and <u>Typha latifolia</u> were dominant plants in the brackish marshes comprising a large portion of the region. <u>Scirpus californicus</u> and <u>Typha latifolia</u> grew upstream beyond the tidal flow while <u>Scirpus acutus</u>, being more tolerant of salt, graded into the <u>Spartina</u> (largely <u>S. foliosa</u>) marshes. There was no interruption of linear sequence of breeding Song sparrows through these transitions and birds of any intermediate zone did not sort out into different plant associations in the brackish marshes (Marshall, 1948).

Both Marshall (1948) and Johnston (1956a,b) noted the Song sparrow occupied territories strung out along the tidal sloughs, exercising definite preferences for vegetation types and not dispersing over the marsh surface. Marshall called attention to differences in territorial spacing. Sparrow pairs were about 48 yards apart along the bayshore and about 52-70 yards between pairs along the brackish marsh sloughs, depending on the width of fringe plants, 15 and 5 yards respectively. The birds used the tallest plants whether it was <u>Scirpus acutus</u>, <u>Spartina</u> or <u>Grindelia</u> in the center of patches of these species for song and calling perches. Plants on levees were visited but territory headquarters were always at slough margins. Flooded areas behind these levees, or any stagnant areas where tidal flow

had been cut off, were avoided as were low, dense growing stands of <u>Scirpus campestris</u>. The birds would visit drained <u>Salicornia</u> patches but did not establish territories in such locations.

Marshall (1948) went on to describe the relationship between Song sparrow distributions and the salt marsh vegetation dominated by Spartina foliosa, Salicornia ambigua, and Grindelia cunefolia. Spartina was found at the lowest elevations, Salicornia was covered only by the highest tides while Grindelia grew on the levees and elevated banks of the sloughs in the Salicornia zone. Song sparrow pairs were spaced at 76 yards, single file, along each bank in the Spartina marsh (Marshall, 1948) with an average territory width of 30 feet (Johnston, 1956b). In the Salicornia-Grindelia marsh, pairs were 30-100 yards apart, being closer together where the Grindelia was widest (Marshall, 1948). Johnston (1956b) disagreed somewhat since he found larger territories at the heads of the sloughs where the height and amount of vegetation, especially Grindelia, was reduced. He described the Spartina zone territories as being larger than those in the Salicornia and, during periods of high numbers, actual densities were 8-10 pairs per acre.

The breeding season is influenced by tidal heights (Johnston, 1956a) with the salt marsh song sparrow breeding about two weeks earlier than the upland form. This seems to be an adaptation to marsh life wherein the birds nest mainly during the lower tidal conditions existing in March. Late nests suffered high mortality from the extended high run of tides in April to June. Nests were placed off the ground, most commonly in <u>Salicornia</u>, <u>Grindelia</u>, <u>Distichlis</u> or <u>Spartina</u>. Any nest less than five inches above the ground would be flooded out. Average nest height above ground was 9.5 inches for the whole marsh and 12 inches in the low marsh. The birds usually tried to put the nests in the highest vegetation but not always. Johnston is of the impression that predators may exert some selective pressure on the upper limit of nest height. No nests were used more than once with nests of the season scattered about in the territory. Generally an increase in vegetation height during the growing season enabled the birds to nest at successfully higher levels. Such action paralleled the increased height of tides during the breeding season.

Marshall (1948) described the Song sparrow to be the only ground foraging bird of the San Francisco Bay salt marshes, dropping down through a <u>Grindelia</u> plant from a singing perch. In fact, the species was limited to areas covered by tides where flows are unimpeded. Both Marshall and Johnston stressed the Song sparrow distributions are limited by vegetation, tidal water, accessibility of the ground for foraging, habitat selection and the sedentary nature of the populations. Such characteristics favor the maintenance of integrity and distinctness of Song sparrow populations in the salt marshes of the San Francisco Bay region.

Sibley (1955) described the responses of salt marsh birds to extreme tides in the San Francisco Bay area at times other

than nesting periods. White-crowned, Savannah and Song sparrows and Long-billed marsh wrens were concentrated along the levees. When flushed they did not fly out over the flooded marsh. Willets and the Least and Western sandpipers settled on floating debris. The Clapper, Sora, and Virginia rails, normally not seen at low or average tidal conditions, were quite evident clinging to <u>Spartina</u> stems or huddled in clumps of emergent vegetation. Predators availed themselves of these concentrations as California and Ring-billed gulls, Marsh hawks and Short-eared owls were reported hunting along the levees and over clumps of emergent vegetation.

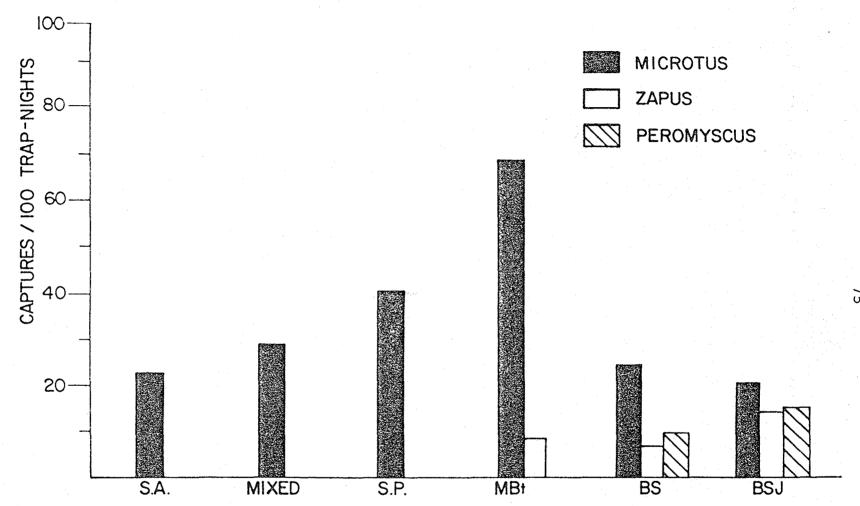
MAMMALIA.

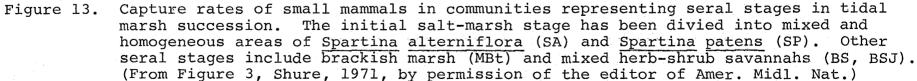
A number of small mammals have been associated with the fringes of salt marshes (Paradiso and Handley, 1965). Shure (1970, 1971) found a definite relationship between small mammal distributions and the topographically controlled pattern of barrier beach vegetation of a New Jersey marsh habitat. The Meadow mouse <u>Microtus pennsylvanicus</u> was the most abundant, constituting 97 percent of the captures associated with the dense herbacious cover of the marshes of the bay shore dominated by <u>Spartina patens</u>. <u>Zapus hudsonius</u>, Meadow jumping mouse and <u>Peromyscus leucopus</u>, the White-footed mouse were fairly abundant, the former being most evident in the brackish marshes and upland borders. <u>Peromyscus</u>, the Masked shrew <u>Sorex</u> <u>cinereus</u> and the House mouse <u>Mus musculus</u> were taken occasionally in the salt marsh but primarily at trap sites in the marsh-upland

border. There is a change in small mammal distributions during successional stages of a salt marsh (Figure 13). Shure (1971) suggests the role of vegetation preference and interspecific relations as contributing factors and seems to prefer the latter.

Microtus distribution is related to tidal flooding wherein Shure (1971) found a significantly greater number of Meadow vole in the drier more dense S. patens vegetation in contrast to the lower wetter S. alterniflora. This preference was further borne out by the significantly greater number of trap captures from the Spartina patens covered spoil piles bordering drainage ditches than from other marsh locations. Several authors (Harris, 1953; Johnston, 1957) comment on the willingness of Microtus to swim. Such behavior must be related to unusual conditions such as storms because Shure (1971) showed a significant restriction of Microtus movements caused by the presence of mosquito control ditches wherein the vole apparently had a reluctance to swim across such a waterway that ranged up to one meter in width. This reluctance was further substantiated by Shure recording significantly smaller recapture distances in a brackish marsh than a salt marsh area. The brackish marsh was highly compartmentalized by ditches and most Microtus were repeatedly recaptured with the same compartment.

Several authors (Harris, 1953; Fisler, 1965; Rudd et al., 1971) comment on the presence of a home range for <u>Microtus</u> in a tidal marsh with Harris identifying the size at one quarter to





A State and the

one half acre. Feeding is presumed to occur within that home range and recent observations in Canary Creek marsh on the frequent recapture of the same individuals and the presence of feeding remains on grass in the same general location would tend to confirm the presence of <u>Microtus</u> to a restricted area.

All of these mammals are inconspicuous marsh fringe inhabitants. The Muskrat, <u>Ondatra zibethica</u>, is a much more obvious animal, conspicuous by its houses dotting the marsh surface and its aquatic runs among the vegetation. Much has been written about its activities both in fresh and brackish water marshes. Ditching for mosquito control which presumably lowered water levels is reported to have an adverse effect on vegetation needed for muskrat food and house construction (Stearns et al., 1939, 1940). Important muskrat foods such as <u>Scirpus olneyi</u> and <u>Spartina cynosuroides</u> have been replaced by species such as <u>Hibiscus oculiroseus</u>, marsh mallow, <u>Kosteletzkya virginia</u>, salt marsh mallow, <u>Solidago remprevirens</u>, seaside goldenrod, <u>Bidens trichoxperma</u>, tick seed sunflower and <u>Aster novi-belgii</u>, presenting a brilliant picture but of no value to the muskrat (Figure 14).

Aerial views of Muskrat, <u>Ondatra zibethicus</u>, activity tend to display discrete sites for feeding with the house or feeding platform being the focal point. During the fall it is hard from such aerial views to distinguish between feeding activity and harvesting or grass for house construction. Errington (1963) has very ably documented the presence of home range and territorialism among muskrats and in addition has described the

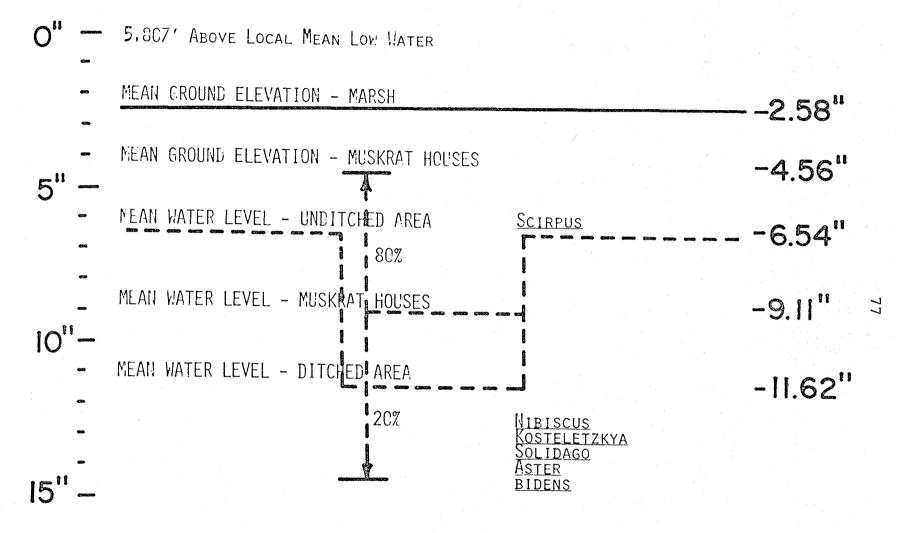


Figure 14. Water levels in ditched and unditched tidal marshes as related to muskrat houses and vegetation. (From Stearns, et al., 1939.)

increased intraspecific strife induced by continued compression of home range and territory resulting from the expansion of the population. He described instances of animals apparently suffering from malnutrition repeatedly being captured within a limited area degraded by overgrazing.

Muskrat food and populations decrease as salinity increases (Harris, 1937; Dozier, 1947; Dozier et al., 1948). The Muskrats prefer the less saline types of vegetation such as the threesquare sedges, Scirpus olneyi, S. robustus, S. americanus, and the cat-tails Typha. Less favored foods are Spartina cynosuroides, S. alterniflora, S. patens, Distichlis spicata and Juncus roemarianus. Dozier et al. (1948) reported the heaviest muskrats, with average weights of 2.25 and 2.26 lbs., consuming Group I plants (Scirpus olneyi, S. americanus, Typha (salinity 0.0-5.0°/00)) and Group II vegetation (S. olneyi, Typha $(5-10^{\circ}/\circ\circ)$). Those feeding in higher salinity areas on Group IV vegetation (S. patens, S. alterniflora, S. olneyi and Typha (15.0-25.0⁰/oo)) and Group V plants (S. patens, S. alterniflora and Juncus roemarianus (25.0-43.0°/00)) had average weights of 2.16 and 2.20 lbs, respectively. On the surface, this weight differential does not appear significant. However, even casual observation reveals the divergence in plant distributions and numbers of muskrat houses along a salinity gradient. Dozier (1947) made a special point about the impact of salinity on tide water muskrat production. The animals were found in the upper reaches of tidal streams and were abundant where tidal influence was reduced. Diked areas or marshes that were flooded by high

salinity water during storms or through evaporation tended to have reduced muskrat populations through loss of food plants and drinking water.

Palmisano (1972) portrayed the interaction between vegetation types, salinity, extent of flooding and drought on the distribution and abundance of the muskrat in Louisiana coastal marshes. Palmisano recognized four major plant communities: (1) saline marshes adjacent to the Gulf of Mexico dominated by relatively few salt tolerance species, (2) brackish marshes forming a broad zone of moderate salinity where plant growth was vigorous, (3) intermediate, slightly brackish marshes, and (4) fresh water marshes. Muskrats occurred in all the coastal marshes examined; however, population densities varied greatly. Brackish marshes composed of a mixed community of Scirpus olneyi and Spartina patens were a preferred habitat (Table 13). Although Palmisano recorded approximately equal percentage values for southeastern and southwestern Louisiana, population densities were much higher in the southeast. The saline marshes in the southeast contained populations equal to overall average density. The saline marshes of the southwest were poor muskrat habitat, possibly because of their well drained nature. The intermediate marshes had below average population densities but high or average levels in restricted areas adjacent to the brackish marshes. The fresh water marshes exhibited the lowest densities of any type; 31.4 percent of the total area examined supported only 4.1 percent of the total houses counted. Populations were recorded at their highest

Table 13.

Acreage of marsh vegetative types in coastal Louisiana, percent of acreage surveyed and percent of houses counted. (Modified from Tables 2, 3, 4, Palmisano, 1972.)

	Southwestern			Southeastern			Total		
		Percent		-		cent		Percent	
		Surveyed			Surveyed			Surveyed	
Marsh Type	Acreage	Acres	Houses	Acreage	Acres	Houses	Acreage	Acres	Houses
Saline	45,507	2.6	1.0	753,130	22.2	20.8	798,637	12.4	14.2
Brackish	463,938	37.5	70.3	722,202	36.1	70.0	1,186,140	36.8	72.6
Intermediate	368,703	31.5	25.0	283,076	7.2	3.7	651,779	19.4	9.1
Fresh	390,757	28.4	3.7	830,769	34.5	5.5	1,221,526	31.4	4.1
Total	1,268,905	100.0	100.0	2,589,177	100.0	100.0	3,858,082	100.0	100.0

during periods of high precipitation and low salinity. The greatest house counts were recorded in February when water levels were generally high, temperatures low and the spring breeding season about to begin.

Palmisano reported wide population fluctuations in localized areas with rapid increases for 3 to 4 years followed by sharp declines to almost zero in a few months. Such a decline was observed in the brackish marshes in December 1971 following two dry summrers. The other marsh types were not as severely affected by the drought with populations actually increasing in a few locations. These marshes served as important reservoirs for muskrats during periods of stress in the brackish marshes. In December 1971, when populations were below normal, non-brackish marshes accounted for over 50 percent of the houses recorded. During normal years less than 20 percent were recorded outside the brackish marsh zone.

Mammals respond to storm tides on a salt marsh much as do the birds; Norway rats and meadow mice are often found in the same clumps of grass with birds (Sibley, 1955). Extremely high tides that drive small mammals from cover occur during the winter months in the San Francisco Bay area (Johnston, 1957; Fisler, 1965; Rudd et al., 1971). At such times only 4-5 inches of vegetation remain exposed, mostly <u>Grindelia cuneifolia</u>, a woody perennial growing along the elevated banks of tidal sloughs. During flooding conditions, these mammals use the emergent vegetation, floating debris, or move to higher ground. Since high ground is some distance from centers of mammal

populations, most animals use emergent vegetation or floating debris. Some are lost and a number of young are drowned in the spring (Johnston, 1957). Swimming ability plays a part. The shrew Sorex vagrans swims and dives well and most nests are on high ground greater than six feet above mean sea level. Rudd et al. (1971) found the greatest numbers of Sorex sinuosus at the interface between the Salicornia ambigua marsh and the Baccharis pilularis and grasses of the levees and found no difference in numbers between a dry summer marsh and a flooded The Meadow mouse Microtus californicus swims winter marsh. and dives well (Johnston, 1957) but prefers to hide in what emergent vegetation exists (Fisler, 1965). Since the mouse does not normally move out of its home range during and after a storm tide, only a rare combination of factors would reduce or displace a population. Such a reduction was reported by Stark (1963) wherein only those voles with home ranges on higher ground survived several hours of winter storm tide and rain. Later observations by Rudd et al. (1971) depicted a significant decrease in the M. californicus population from a summer dry marsh to an inundated winter marsh. They suggested the decrease could be explained by drowning, increased predation or exclusion caused by the high numbers of Peromyscus maniculatus or Mus musculus. Harris (1953) was not clear in his own mind how Microtus pennsylvanicus survived high water in the Blackwater Refuge marshes of Maryland. It was not evident whether high tides cause the voles to move to muskrat houses from the marsh surface or whether the tides restrict the Microtus to the vicinity of the

houses, thereby making concentrating signs of their activity as well as making it easier to take in traps. Harris considered movement to high ground (landward edge of the marsh, islands in the marsh or muskrat houses) as means for survival but if the mice had to travel any distance exposure to predation would be increased. While he acknowledged the mice's ability to swim and dive with ease and that some did drown. Harris found it difficult to accept that Microtus could scramble from one Spartina culm to the next for any length of time and still survive. Fisler found few mice in the Spartina foliosa, more in Salicornia ambigua and the largest concentrations in the Grindelia growing on the levees. This preference for high ground by M. californicus agrees with the association between Microtus pennsylvanicus and high marsh Spartina patens in New Jersey (Shure, 1970, 1971). However, Johnston (1957) and Rudd et al. (1971) found Microtus nests at all levels of the marsh surface, the breeding season occurring during the fewest excessive tides.

The Norway rat, <u>Rattus norwegicus</u> was described by Johnston (1957) as an excellent swimmer and diver but suffering considerable losses of young because its nests are located on the ground. Apparently during the non-breeding season the Norway rat is better able to survive (Stark, 1963). In contrast, the Harvest mouse is a poor swimmer; its fur wets easily and it does not take to the water readily. It normally nests above ground in old sparrow nests and Johnston did not observe any nests being flooded out. This may explain why Rudd et al. (1971) found only a slight increase in Reithrodontomys raviventris numbers moving

REFERENCES CITED

- Apley, M. L. 1970. Field studies on life history, gonadal cycle and reproductive periodicity in <u>Melampus bidentatus</u> (Pulmonate: Ellobiidae). Malacol. 10(2): 381-397.
- Apley, M. L., W. D. Russell-Hunter and R. J. Avolizi. 1967. Annual reproductive turnover in salt marsh pulmonate snail, <u>Melampus</u> <u>bidentatus</u>. (Abstract) Biol. Bull. 133(2): 455-456.
- Bailey, N. S. 1948. A mass collection and population technique for larvae of Tabanidae (Diptera). Bull. Brooklyn Ent. Soc. 43: 22-29.
- Bent, A. C. 1929. Life histories of North American shore birds. U. S. Nat. Mus. Bull. 146: I-IX, 1-412.
- Bent, A. C. 1963. Life histories of North American marsh birds. U. S. Nat. Mus. Bull. 135. 490 pp.
- Bent, A. C. and collaborators. Compiled and edited by O. L. Austin, Jr. 1968. Life histories of North American cardinals, grosbeaks, buntings, towhees, finches, sparrows and allies. U. S. Nat. Mus. Bull. 237. 1889 pp.
- Connell, W. A. 1940. Tidal inundation as a factor limiting distribution of <u>Aedes</u> spp. on a Delaware salt marsh. Proc. N. J. Mosq. Exterm. Assoc. 27: 166-177.
- Crane, J. 1975. Fiddler crabs of the world: Oxypodidae: Genus Uca. Princeton Univ. Press. 736 pp.
- Crichton, O. W. 1960. Marsh crab -- intertidal tunnel-maker and grass-eater. Est. Bull. 5: 3-10.
- Daiber, F. C. 1974. Salt marsh plants and future coastal salt marshes in relation to animals. In: R. J. Reimold and W. H. Queen, eds. Ecology of Halophytes. Academic Press, N. Y. pp. 475-510.
- Daiber, F. C. 1977. Salt-marsh animals: distributions related to tidal flooding, salinity and vegetation. Chapter 5 in: Wet Coastal Ecosystems. V. J. Chapman, ed. Elsevier Scientific Publ. Co., Amsterdam.
- Daiber, F. C. and O. Crichton. 1967. Caloric studies of <u>Spartina</u> and the marsh crab <u>Sesarma</u> <u>reticulatum</u> (Say). <u>Ann. Pittman-Robertson Rept. Del. Bd. Game and Fish Comm.</u> <u>Proj. W-22-R-2</u>, Job No. 4. 20 pp.
- Davis, L. V. and I. E. Gray. 1966. Zonal and seasonal distribution of insects in North Carolina salt marshes. Ecol. Monogr. 36(3): 275-295.

- Dexter, R. W. 1942. Notes on the marine mollusks of Cape Ann, Massachusetts. Nautilus 56(2): 57-61.
- Dexter, R. W. 1944. Annual fluctuations of abundance of some marine mollusks. Nautilus 58(1): 20.
- Dexter, R. W. 1945. Zonation of the intertidal marine mollusks at Cape Ann, Massachusetts. Nautilus 58(2): 56-64.
- Dozier, H. L. 1947. Salinity as a factor in Atlantic Coast tide water muskrat population. Trans. N. Amer. Wildl. Conf. 12: 398-420.
- Dozier, H. L., M. H. Markley and L. M. Llewellyn. 1948. Muskrat investigations on the Blackwater National Wildlife Refuge, Maryland, 1941-1945. J. Wildl. Mgmt. 12(2): 177-190.
- Dukes, J. C., T. D. Edwards and R. C. Axtell. 1974a. Associations of Tabanidae (Diptera) larvae with plant species in salt marshes, Carteret County, North Carolina. Environ. Entomol. 3(2): 280-286.
- Dukes, J. C., T. D. Edwards and R. C. Axtell. 1974b. Distribution of larval tabanidae (Diptera) in a Spartina alterniflora salt marsh. J. Med. Entomol. 11(1): 79-83.
- Errington, P. L. 1963. Muskrat populations. Iowa State Univ. Press. 665 pp.
- Evans, P. D., C. N. Ruscoe and J. E. Treherne. 1971. Observations on the biology and submergence behavior of some littoral beetles. J. mar. biol. Assoc. U.K. 51(2): 375-386.
- Ferrigno, F. 1957. Clapper rail study. In: J. W. Aldrich, et al. Investigations of woodcock, snipe and rails in 1956. U. S. Fish and Wildl. Serv. Spec. Sci. Rept., Wildl. 34: 81-85.
- Ferrigno, F. 1958. A two-year study of mosquito breeding in the natural and untouched salt marshes of Egg Island. Proc. N. J. Mosq. Exterm. Assoc. 45: 132-139.
- Ferrigno, F. 1961. Variations in mosquit-wildlife associations on coastal marshes. Proc. N. J. Mosq. Exterm. Assoc. 48: 193-203.
- Fisler, G. F. 1965. Behavior of salt marsh <u>Microtus</u> during winter high tides. J. Mamm. 42: 37-43.
- Foster, W. A. and J. E. Treherne. 1975. The distribution of an intertidal aphid, <u>Pemphigus trehernei</u> Foster, on marine salt marshes. Oecologia 21(2): 141-155.

- Foster, W. A. and J. E. Treherne. 1976. The effects of tidal submergence on an intertidal aphid, <u>Perphigus</u> trehernei Foster. J. Anim. Ecol. 45(1): 291-301.
- Freeman, J. V. and E. J. Hansens. 1972. Collecting larvae of the salt marsh greenhead <u>Tabanus</u> <u>nigrovittatus</u> and related species in New Jersey: comparison of methods. Environ. Entomol. 1: 653-658.
- Gerry, B. I. 1950. Salt marsh fly control as an adjunct to mosquito control in Massachusetts. Proc. N. J. Mosq. Exterm. Assoc. 37: 189-193.
- Gray, E. H. 1942. Ecological and life history aspects of red jointed fiddler crab, <u>Uca minax</u> (Le Conte), region of Solomons Island, Maryland. <u>Md. Bd.</u> Nat. Res., Dept. Res. and Ed. CBL Publ. 51. pp. 3-20.
- Green, J. W. M. Harsch, L. Barr and C. L. Prosser. 1959. The regulation of water and salt by the fiddler crabs, Uca pugnax and Uca pugilator. Biol. Bull. 116(1): 76-87.
- Hackney, A. G. 1944. List of mollusca from around Beaufort, North Carolina, with notes of Tethys. Nautilus 58(2): 56-64.
- Hansens, E. J. 1949. The biting fly problem in New Jersey resorts and its relation to mosquito control. Proc. N. J. Mosq. Exterm. Assoc. 36: 126-130.
- Hansens, E. J. 1952. Some observations on the abundance of salt marsh greenheads. Proc. N. J. Mosq. Exterm. Assoc. 39: 93-98.
- Harris, E. S. 1937. Muskrat culture and its economic significance on New Jersey. Proc. N. J. Mosq. Exterm. Assoc. 24: 20-25.
- Harris, V. T. 1953. Ecological relationships of meadow voles and rice rats in the tidal marshes. J. Mammal. 34: 479-487.
- Hauseman, S. A. 1932. A contribution to the ecology of the salt-marsh snail, <u>Melampus</u> <u>bidentatus</u> Say. Amer. Natur. 66: 541-545.
- Holle, P. A. 1957. Life history of the salt marsh snail, Melampus bidentatus Say. Nautilus 70: 90-95.
- Jamnback, H. and W. J. Wall. 1959. The common salt marsh Tabanidae of Long Island, New York. Bull. N. Y. St. Mus. No. 375. 77 pp.

- Johnston, R. F. 1956a. Population structure in salt marsh song sparrows. I. Environment and annual cycle. Condor 58: 24-44.
- Johnston, R. F. 1956b. Population structure in salt marsh song sparrows. II. Density, age structure, and maintenance. Condor 58: 254-272.
- Johnston, R. F. 1957. Adaptation of salt marsh mammals to high tides. J. Mamm. 38(4): 529-531.
- Kerwin, J. A. 1971. Distribution of the fiddler crab (Uca minax) in relation to marsh plants within a Virginia estuary. Ches. Sci. 12(3): 180-183.
- Kerwin, J. A. 1972. Distribution of the salt marsh snail (Melampus bidentatus Say) in relation to marsh plants in the Poropotank River area, Virginia. Ches. Sci. 13(2): 150-153.
- Kraeuter, J. N. and P. L. Wolf. 1974. The relationship of marine macroinvertebrates to salt marsh plants. In: R. J. Reimold and W. H. Queen, eds. Ecology of Halophytes. Academic Press, N. Y. pp. 449-462.
- Kuenzler, E. J. 1961. Structure and energy flow of a mussel population in a Georgia salt marsh. Limnol. Oceanogr. 6(2): 191-204.
- Larsen, E. B. 1951. Studies on the soil fauna of Skallingen. Oikos 3(2): 166-192.
- LaSalle, R. N. and K. L. Knight. 1974. Effects of salt marsh impoundments on mosquito populations. Water Resources Res. Inst., Univ. N. Car. Rept. No. 92. 85 pp.
- Lee, J. J. and W. A. Muller. 1973. Trophic dynamics and niches of salt marsh Foraminifera. Am. Zool. 13: 215-223.
- Lent, C. M. 1967a. Effects and adaptive significance of airgaping by the ribbed mussel <u>Modiolus</u> (Arcuatula) <u>Demissus</u> (Dillwyn). Ph.D. dissertation. Univ. Delaware. 77 pp.
- Lent, C. M. 1967b. Effect of habitat on growth indices in the ribbed mussel <u>Modiolus</u> (Arcuatula) <u>demissus</u>. Ches. Sci. 8(4): 221-227.
- Lent, C. M. 1968. Air-gaping by the ribbed mussel, Modiolus demissus (Dillwyn): Effects and adaptive significance. Biol. Bull. 134: 60-73.
- Lent, C. M. 1969. Adaptations of the ribbed mussel, <u>Modiolus</u> <u>demissus</u> (Dillwyn) to the intertidal habitat. Am. Zool. 9: 283-292.

Lockwood, A. P. M. 1962. The osmoregulation of Crustacea. Biol. Rev. 37(2): 257-305.

- Luxton, M. 1967. The ecology of salt-marsh Acarina. J. Anim. Ecol. 36(2): 257-277.
- Marshall, J. T., Jr. 1948. Ecologic races of song sparrows in the San Francisco Bay region. I. Habitat and abundance. Condor 50: 193-215.
- Meanley, B. 1969. Natural history of the king rail North American fauna. Bur. Sports Fish. and Wildl., Fish and Wildl. Serv., Publ. No. 67. 108 pp.
- Meany, R. A., I. Valiela and J. M. Teal. 1976. Growth, abundance and distribution of larval tabanids in experimentally fertilized plots on a Massachusetts salt marsh. J. Appl. Ecol. 13(2): 323-332.
- Miller, D. C. 1961. The feeding mechanisms of fiddler crabs with ecological considerations of feeding adaptations. Zool. 46(8): 89-101.
- Miller, K. G. and D. Maurer. 1973. Distribution of the fiddler crabs, Uca pugnax and Uca minax, in relation to salinity in Delaware rivers. Ches. Sci. 14(3): 219-221.
- Olkowski, W. 1966. Biological studies of salt marsh Tabanids in Delaware. Master's thesis. Univ. Delaware. 116 pp.
- Oney, J. 1954. Final report: clapper rail survey and investigation study. Ga. Game and Fish Comm. 50 pp.
- Owen, M. 1971. The selection of feeding site by White-fronted geese in winter. J. Appl. Ecol. 8(3): 905-917.
- Palmisano, A. W. 1972. The distribution and abundance of muskrat (Ondatra zibethica) in relation to vegetative types in Louisiana coastal marshes. Proc. Ann. Conf. Southeast. Assoc. Game and Fish Comm. 26: 1-31.
- Paradiso, J. L. and C. O. Handley, Jr. 1965. Checklist of mammals of Assateaque Island. Ches. Sci. 6(3): 167-171.
- Parker, N. H. 1976. The distribution, growth and life history of <u>Melampus bidentatus</u> (Gastropoda: Pulmonata) in the Delaware Bay region. Master's thesis. Univ. Delaware. 65 pp.

Pearse, A. S. 1914. Habits of fiddler crabs. Ann. Report Smith. Inst. 1913: 415-428.

- Phillips, N. W. 1978. Spatial distribution and population dynamics of Orchestia spp. (Amphipoda: Talitridae) in the Canary Creek salt marsh, Delaware. Master's thesis. Univ. Delaware. 187 pp.
- Phleger, F. B. 1970. Foraminifera populations and marine marsh processes. Limnol. Oceanogr. 15: 522-534.
- Phleger, F. B. 1977. Soils of marine marshes. In: V. J. Chapman, ed. Wet Coastal Ecosystems. Elsevier Sci. Publ. Co. pp. 69-77.
- Rockel, E. G. 1969. Marsh physiography: influence on distribution of intertidal organisms. Proc. N. J. Mosq. Exterm. Assoc. 56: 102-115.
- Rockel, E. G. and E. J. Hansens. 1970a. Distribution of larval horse flies and deer flies (Diptera: Tabanidae) of a New Jersey salt marsh. Ann. Entomol. Soc. Am. 63: 681-684.
- Rockel, E. G. and E. J. Hansens. 1970b. Emergence and flight activity of salt marsh horseflies and deerflies. Ann. Entomol. Soc. Am. 63: 27-31.
- Rudd, R. L., H. C. Hadaway and J. R. Newman. 1971. Differential responses of five species of salt marsh mammals to inundation. J. Mamm. 52: 818-820.
- Russell-Hunter, W. D., M. L. Apley and R. D. Hunter. 1972. Early life history of <u>Melampus</u> and the significance of semilunar synchrony. <u>Biol. Bull.</u> 143(3): 623-656.
- Schwartz, B. and S. R. Safir. 1915. The natural history and behavior of the fiddler crab. Cold Spring Harbor Monogr. 8: 1-24.
- Shanholtzer, G. F. 1974. Relationship of vertebrates to salt marsh plants. In: R. J. Reimold and W. H. Queen, eds. Ecology of Halophytes. Academic Press, N. Y. pp. 463-474.
- Shure, D. J. 1970. Ecological relationships of small animals in a New Jersey barrier marsh habitat. J. Mamm. 51: 267-278.
- Shure, D. J. 1971. Tidal flooding dynamics: its influence on small mammals in barrier beach marshes. Amer. Midl. Nat. 85(1): 36-44.
- Sibley, C. G. 1955. The responses of salt-marsh birds to extremely high tides. Condor 57: 241-242.
- Smith, J. B. 1902. The salt marsh mosquito, <u>Culex</u> sollicitans Wlk. Spec. Bull. N. J. Agr. Exp. Sta. 10 pp.

- Stark, H. E. 1963. Nesting habits of the California vole, <u>Microtus californicus</u>, and microclimatic factors affecting its nests. Ecol. 44(4): 663-669.
- Stearns, L. A., D. MacCreary and F. C. Daigh. 1939. Water and plant requirements of the muskrat on a Delaware tide water marsh. Proc. N. J. Mosq. Exterm. Assoc. 26: 212-221.
- Stearns, L. A., D. MacCreary and F. C. Daigh. 1940. Effects of ditching on the muskrat population of a Delaware tide water marsh. Univ. Delaware Agr. Exp. Sta. Bull. 225. 55 pp.
- Stewart, R. E. 1951. Clapper rail populations of the Middle Atlantic States. Trans. N. Amer. Wildl. Conf. 16: 421-430.
- Stewart, R. E. 1952. Clapper rail studies. In: J. W. Aldrich, et al. Investigations of woodcock, snipe and rails in 1951. U. S. Fish and Wildl. Serv. Spec. Sci. Rept., Wildl. 14: 56-58.
- Stewart, R. E. 1962. Waterfowl populations in the upper Chesapeake region. U. S. Fish and Wildl. Serv. Spec. Sci. Rept., Wildl. 65. 208 pp.
- Stewart, R. E. and C. S. Robbins. 1958. Birds of Maryland and the District of Columbia. N. Amer. Fauna No. 62. Bur. Sports Fish. and Wildl., Fish and Wildl. Serv. 401 pp.
- Teal, J. M. 1958. Distribution of fiddler crabs in Georgia salt marshes. Ecol. 39(2): 185-193.
- Teal, J. M. 1959. Respiration of crabs in Georgia salt marshes and its relation to their ecology. Physiol. Zool. 32(1): 1-14.
- Teal, J. M. 1962. Energy flow in the salt marsh ecosystem of Georgia. Ecol. 43(4): 614-624.
- Teal, J. M. and F. G. Carey. 1967. The metabolism of marsh crabs under conditions of reduced oxygen pressure. Physiol. Zool. 40(1): 83-91.
- Tomkins, I. R. 1941. Notes on Macgillivray's seaside sparrow. Auk 58: 38-51.
- Treherne, J. E. and W. A. Foster. 1977. Diel activity of an intertidal beetle, <u>Dicheirotrichus gustavi</u> Crotch. J. Anim. Ecol. 46(1): 127-138.
- Urner, C. A. 1935. Relation of mosquito control in New Jersey to bird life in the salt marshes. Proc. N. J. Mosq. Exterm. Assoc. 22: 130-136.

- Vogt, W. 1938. Preliminary notes on the behavior and ecology of the eastern Willet. Proc. Linn. Soc. N. Y. 49: 8-42.
- Wall, W. J., Jr. and O. W. Doane, Jr. 1960. A preliminary study of the blood-sucking Diptera on Cape Cod, Massachusetts. Mosq. News 20: 39-44.
- Wall, W. J., Jr. and H. Jamnback. 1957. Sampling methods used in estimating larval reproduction of salt marsh Tabanids. J. Ecol. Entomol. 50: 389-391.
- Waugh, D. L. and E. T. Garside. 1971. Upper lethal temperatures in relation to osmotic stress in the ribbed mussel Modiolus demissus. J. Fish. Res. Bd. Canada 28(4): 527-532.
- Whiting, N. H. and G. A. Moshiri. 1974. Certain organismsubstrate relationships affecting the distribution of Uca minax. Hydrobiol. 44(4): 481-493.